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1 Research paper

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Combined bioreduction and volatilization of Se^{VI} by *Stenotrophomonas bentonitica*: formation of trigonal selenium nanorods and methylated species

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- 23

- 24 Abstract
- 25

26 Nowadays, metal pollution due to the huge release of toxic elements to the environment has 27 become one of the world's biggest problems. Bioremediation is a promising tool for reducing 28 the mobility and toxicity of these contaminants (e.g. selenium), being an efficient, 29 environmentally friendly, and inexpensive strategy. The present study describes the capacity of Stenotrophomonas bentonitica to biotransform Se^{VI} through enzymatic reduction and 30 31 volatilization processes. HAADF-STEM analysis showed the bacterium to effectively reduce Se^{VI} (200 mM) into intra- and extracellular crystalline Se⁰ nanorods, made mainly of two 32 different Se allotropes: monoclinic (*m*-Se) and trigonal (*t*-Se). XAS analysis appears to indicate 33 34 a Se crystallization process based on the biotransformation of amorphous Se⁰ into stable *t*-Se 35 nanorods. In addition, results from headspace analysis by gas chromatography-mass 36 spectometry (GC-MS) revealed the formation of methylated volatile Se species such as DMSe 37 (dimethyl selenide), DMDSe (dimethyl diselenide), and DMSeS (dimethyl selenenyl sulphide). 38 The biotransformation pathways and tolerance are remarkably different from those reported with this bacterium in the presence of Se^{IV}. The formation of crystalline Se⁰ nanorods could 39 have positive environmental implications (e.g. bioremediation) through the production of Se of 40 41 lower toxicity and higher settleability with potential industrial applications.

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43 Keywords: bacteria, selenate, reduction, bioremediation, nanorod, volatilization

- 44 **1. Introduction**
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46 Exhaustive investigations in the past decade have shed light on the occurrence of a wide 47 diversity and distribution of microorganisms recognized as capable of selenium (Se) oxyanion 48 bio-transformations (Avendaño et al. 2016; Presentato et al. 2018; Ojeda et al. 2020). Oxidation, 49 reduction, and volatilization have been reported as the main pathways involved in the biotransformation of the different Se oxidation states: selenate (Se^{VI}), selenite (Se^{IV}), elemental 50 51 Se (Se⁰), and selenide (Se^{II-}). Selenate and selenite are known to be environmentally hazardous due to their high solubility, mobility, and bioavailability, while Se⁰ is insoluble and less toxic. A 52 53 few studies have reported the oxidation of reduced Se species by microorganisms (Dowdle and 54 Oremland 1998; Losi and Frankenberger 1998; Luo et al. 2022; Nancharaiah and Lens 2015). 55 However, microbial oxidation of Se is not usually considered to be of major relevance for the 56 environment because of the low rates at which these reactions occur (Eswayah et al. 2016). Se 57 volatilization is a biotransformation process now considered as a promising mechanism for 58 bioremediation purposes. Some selenium-resistant microorganisms are able to volatilize Se 59 through biomethylation processes (Eswayah et al. 2017). Selenium methylated compounds such 60 as DMSe or DMDSe reportedly present limited bioavailability, toxicity, and solubility in comparison to Se oxyanions (Doran 1982; Hasanuzzaman et al 2020; Ranjard et al. 2003). 61 Finally, the reduction of oxidized and toxic forms of Se (Se^{VI} and Se^{IV}) to Se⁰ nanoparticles 62 63 (NPs) has been extensively studied (Martínez et al. 2020; Tugarova et al. 2020). Even though 64 changes in the oxidation state —hence bioreduction— have been confirmed in many cases, the 65 specific mechanisms, regulators, and biochemical pathways involved in this process have not yet been fully elucidated. To date, the number of Se^{IV}-reducing microbial isolates is 66 considerably higher than Se^{VI}-reducing ones, which generally require a two-step process in 67 which Se^{IV} act as an intermediate product. Therefore, both volatilization and reduction are 68 69 potential mechanisms to be used in bioremediation of contaminated environments since they 70 involve the removal of toxic Se species. However, there is still much to be investigated about 71 bioremediation and its possible in-situ application, so any step forward in its research would be 72 of crucial importance in the search for an effective strategy.

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Of industrial and environmental interest is the type, form, and location of the Se⁰NPs produced after Se bioreduction. Indeed, a wide array of nanoparticle shapes (spheres, nanotubes, nanorods, etc.), structures (amorphous, trigonal, monoclinical, etc.), sizes, and cellular locations has been described. This suggests that there is no unique Se biotransformation mechanism. Elemental Se exists in nature in several allotropic forms including both amorphous (*a*-Se) and crystalline varieties (monoclinic (*m*-Se) and trigonal Se (*t*-Se)). Amorphous Se tends to change to the more stable crystalline Se by heating, use of chemical reagents, and other physico81 chemical methods (Zhang et al. 2012). However, some bacteria can transform a-Se to m-Se and 82 t-Se at room temperature and without the use of additional reactive (Komova et al. 2018; Ruiz-83 Fresneda et al. 2020; Pinel-Cabello et al. 2021). Indeed, some authors propose that a-Se 84 nanospheres are released from the cell and transformed to Se crystal nanostructures of different shapes (Wang et al. 2010; Ruiz-Fresneda et al. 2018; Ruiz-Fresneda et al. 2020). For example, 85 the bacterium Stenotrophomonas bentonitica have been recently identified to reduce Se^{IV} to Se⁰ 86 87 nanospheres (a-Se), different crystalline nanostructures (m-Se and t-Se), and with the formation 88 of volatile methylated Se (Ruiz-Fresneda et al. 2020). They proposed a transformation 89 mechanism from a-Se to Se crystals including the intracellular synthesis of the a-Se 90 nanospheres and their subsequent release, aggregation, and transformation in the extracellular 91 space. Unfortunately, it remains unclear how these relatively huge nanospheres —in comparison 92 with the cell size— are formed, assembled, released, and transformed. It is likely that many 93 unknown proteins, enzymes, and transport complexes may be involved. Interestingly, results from recent proteomic studies indicate the possible role of specific proteins in Se^{IV} reduction. 94 95 including RND (resistance-nodulation-division) transport systems, and glutathione reductase 96 (Pinel-Cabello et al. 2021), which has been found to be important in Se reduction in other 97 bacterial species (Ni et al. 2015; Martínez et al. 2020).

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99 The importance of Se crystal formation lies in the potential number of applications derived from 100 it. For instance, crystalline Se has been described to be more settleable than a-Se (Lenz et al. 101 2009), which may be beneficial for the decontamination of Se polluted environments. Recent 102 experiments in bioreactors highlight bioremediation as potential tool for contaminated water 103 treatment (Dessì et al. 2016; Ojeda et al. 2020). Likewise, certain Se-reducing bacteria may play 104 an important role in the immobilization of Se, positively affecting the safety of deep geological 105 repositories (DGR) (Ruiz-Fresneda et al. 2018; Ruiz-Fresneda et al. 2019), the most accepted 106 option for final disposal of radioactive residues. In addition, the utility of t-Se in many industrial 107 and medical applications is well-known. t-Se is a photoconductor of broad spectral sensitivity, 108 making it very useful in solar cells, photocells, rectifiers, photographic exposure meters, and 109 xerography (Ibragimov et al. 2000; An et al. 2003; Zhu et al. 2019). From a medical standpoint, 110 Se nanoparticles hold potential as antitumoral and antibacterial agents (Kuršvietienė et al. 2020; 111 Filipović et al. 2021). It is worth mentioning that for many of the previous applications, 112 nanoparticles must be crystalline, smaller than 100 nm, and as flawless as possible (An et al. 113 2003). Although many studies are reporting the bioproduction of selenium nanoparticles 114 (SeNPs) through bacteria, archaea, plants, fungi, etc., there is still no optimized eco-friendly 115 methodology applied to the industrial production of NPs. For this reason, any contribution to the 116 field could be of great help in the search for an effective synthesis method.

The study presented here describes the mechanisms involved in the reduction of Se^{VI} by S. 118 bentonitica as compared to those reported for Se^{IV} (Ruiz-Fresneda et al. 2018). This bacterium 119 reduced Se^{VI} to extracellular and intracellular Se⁰ nanorods (*m*-Se and *t*-Se) and methylated Se 120 compounds. The results evidenced a different and novel Se^{VI} reduction mechanism entailing the 121 formation of intracellular crystalline nanorods, never described before, with no observation of 122 123 a-Se nanospheres. This study thus provides new information to be considered in the 124 development of novel bioremediation strategies and tools. In addition, S. bentonitica is clearly 125 identified as a candidate for environmentally friendly methodologies in t-Se nanorod 126 fabrication.

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2. Materials and methods

131 2.1. Bacterial species and growing conditions under Se^{VI} stress

133 The bacterium Stenotrophomonas bentonitica used in the present study was isolated from 134 Spanish bentonite clays (Almeria, Spain) and selected based on previous genomic and metal 135 interaction studies (Sánchez-Castro et al. 2017). The bacterial cells were grown aerobically in 136 Luria-Bertani (LB) broth medium (tryptone 10 g/l, NaCl 10 g/l, and yeast extract 5 g/l and, pH 7.0 ± 0.2) at 28 °C and 180 rpm. Specifically, the cells were inoculated to an initial optical 137 density (OD) of 0.1 at a wavelength of 600 nm for all the experiments. Se^{VI} tolerance by S. 138 139 bentonitica was studied by growing the cells aerobically in liquid LB (25 ml) added with 50, 140 100, and 200 mM Se^{VI} at 28 °C by shaking at 180 rpm. Growth was quantified at different 141 incubation times (0, 8, 24, 48, and 72 h) by calculating the cell protein content following the 142 method employed by Ruiz-Fresneda et al. (2018, 2019, 2020) based on Bradford's reagent.

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2.2. X-ray diffraction (XRD)

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146 XRD analysis was used to identify the crystalline phase of the Se reduction products. For this
147 purpose, high volumes of *S. bentonitica* cultures (500 ml) treated with different initial Se^{VI}
148 concentrations (50, 100, 150, and 200 mM) were harvested after 24 h (10,000 x g for 10 min).
149 Dried powder samples were obtained as indicated by Ruiz-Fresneda et al. (2018) and measured
150 with a Bruker D8 Advanced diffractometer linked to a LINXEYE detector.

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2.3. X-ray absorption spectroscopy (XAS) measurements

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154 XAS is a synchrotron-based analytical technique widely employed for local chemical structure 155 determinations of different materials, and hence can be successfully used to determine the 156 structure and oxidation state of Se allotropes produced by the cells of S. bentonitica (Lopez-157 Fernandez et al. 2020). For XAS experiments, S. bentonitica cells were grown in LB supplemented with 200 mM Se^{VI}. After 17, 24 and 48 h of incubation the samples were 158 159 collected, washed, dried, and powdered as described above in section 2.2. Subsequently, the 160 powder samples were pressed on Kapton tape as indicated by Ruiz-Fresneda et al. (2020). Se standards (sodium selenate-Na₂SeO₄ (Se^{VI}), sodium selenite-Na₂SeO₃ (Se^{IV}), Se⁰ foil (*t*-Se) and 161 selenium sulphide-SeS₂ (Se^{-II})) were prepared with cellulose forming small disks following the 162 163 procedures of Ruiz-Fresneda et al. 2020. The XAS data of the experimental samples and Se 164 standards were collected in fluorescence and transmission mode, respectively.

Selenium K–edge X-ray absorption spectra were measured at the MARS beamline (SOLEIL synchrotron facility in Paris, France), which is a bending magnet beamlinefor Multi Analyses on Radioactive Samples. The experimental setup and technical parameters behind the measurements were same as those followed in Ruiz-Fresneda et al. (2020); they are detailed in the **Supplementary Material 1.1**.

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172 2.4. Electron Microscopy

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174 The crystallographic/structural analysis and cellular location of the Se reduction products were 175 analyzed by means of high-angle annular dark field scanning transmission electron microscopy 176 (HAADF-STEM) fitted with energy dispersive X-ray (EDX), selected-area electron diffraction 177 (SAED), and Fast Fourier Transform (FFT). The samples consisting of Se^{VI}-treated cells (150 and 200 mM Se^{VI}) were prepared as described in Merroun et al. (2005) after 24 and 48 h 178 179 incubating, and subsequently examined under a HAADF-STEM microscope FEI TITAN G2 80-180 300 (University of Granada, Granada, Spain). The samples were further examined under FEG-181 ESEM (field emission gun environmental scanning electron microscopy) on a FEG-SEM 182 Microscope FEI QEMSCAN 650F (University of Granada, Spain). The samples were prepared 183 using the critical point drying method as described in Ruiz-Fresneda et al. (2018).

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185 2.5. Gas Chromatography Mass Spectrometry (GC-MS).

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GC-MS combined with thermal desorption (250°C) system was employed for the qualitative analysis of volatile Se-containing compounds released by *S. bentonitica* cells. To this end, the cells were incubated with 2 and 100 mM Se^{VI} in special conical flasks (QuickfitTM) capped with Suba-Seals (SigmaTM) rubber septa as described in Ruiz-Fresneda et al. (2020). Headspace gases were sampled and analyzed after 144 h of incubation as described previously (Eswayah et al. 2017). Se^{VI}-free cultures (biotic) and Se^{VI}-added media (abiotic) were employed as controls. All measurements were performed in duplicate.

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3. Results and discussion

199 3.1. Selenate reduction and tolerance assays for *S. bentonitica*

The reduction of Se^{VI} was confirmed by the formation of red precipitates characteristic of Se⁰ in 201 Se^{VI}-treated cultures after 24 h of incubation (Figure 1A). The non-formation of such 202 precipitates in Se^{VI}-untreated cultures (Figure 1B) and Se^{VI}-treated media (biotic and abiotic 203 controls) (Figure 1C-D) indicated that Se^{VI} reduction is a biological process mediated by the 204 cells of S. bentonitica. Interestingly, the formation of the reddish colour and thus of Se^{0} 205 nanoparticles of interest in the Se^{VI}-treated cultures was only observed at high initial 206 concentrations (100-200 mM), but not at lower initial concentrations (10 and 50 mM) (Figure 207 S1-Supplementary Material). These results suggest that Se^{VI} would not exert a high toxic 208 209 effect upon the cells of S. bentonitica at low metalloid concentrations. For this reason, 200 mM Se^{VI} was selected as initial concentration for most of the experiments. It is noteworthy that 210 transcriptomic studies showed that the cells respond metabolically differently with Se^{VI} 211 212 concentration (Pinel, 2021). Thus, genes encoding for efflux transporters and several 213 oxidoreductases were clearly overexpressed at concentrations below 200 mM (e.g. 50 mM). In 214 this case, where no red precipitationis present, the cells would be capable of reducing Se^{VI} to Se^{IV}, which is removed most probably by these transporters to the extracellular space. For all 215 216 the above reasons, it is quite probable that an additional resistance mechanism apart from 217 reduction —such as volatilization, which does not trigger a colour change in the cultures— is 218 involved.



Figure 1. Cultures of *Stenotrophomonas bentonitica* in LB broth medium with (A) and without 200 mM
 Se^{VI} (B) after 24h incubation. Dead cell (C) and abiotic (D) controls were both supplemented with 200
 mM Se^{VI}.

The toxicity of Se^{VI} was evaluated by determining the bacterial growth rate under different Se^{VI} 223 concentration levels (0, 50, 100, and 200 mM). As can be seen in Figure 2, cell growth curve of 224 the biotic control in the absence of selenate (0 mM Se^{VI}) showed a normal pattern of bacterial 225 growth. Considerable differences in cell growth in the presence of 50 and 100 mM during the 226 first 48 h confirmed the toxicity of Se^{VI}, with an observable small delay in the lag phase of 227 growth of about 8 h. However, after 48 h growing, no differences were appreciated between the 228 control and these Se^{VI} concentrations. When the cells were grown under 200 mM Se^{VI}, the 229 230 growth rate is clearly affected. The lag and the exponential phases are prolonged until 24 and 48 231 h, respectively. However, despite the evident toxic effect, the cells were able to reach growth 232 levels compare with the control after 72 h, revealing the capacity of S. bentonitica to tolerate high concentration levels of Se^{VI}. All these observations suggest that the cells cope with the 233 toxicity of Se^{VI} through several mechanisms depending on concentrations, including enzymatic 234 reduction, export of Se^{VI} by oxidoreductases and transporters, and/or mitigation of oxidative 235 236 stress by enzymes such as glutathione S-transferase (Pinel, 2021). The bacterium of study was previously characterized as efficiently reducing Se^{IV} at lower initial concentrations (0.1 to 2 237 mM) than those employed here for Se^{VI} (Ruiz-Fresneda et al. 2020) —in their study, S. 238 bentonitica growth rate was strongly affected by Se^{IV} at 2 mM. This finding reveals that S. 239 240 *bentonitica* cells are more resistant to Se^{VI} than to Se^{IV}.



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and 200 mM Se^{VI}.

Although Se is usually present in nature at relatively low concentrations, there is a growing increase in the release of this contaminant into the environment in the last years. For instance, Chang et al. (2019) reported concentrations up to 7007 mg/kg in seleniferous soil areas in

central China, similar to those tested in the present study. In addition, several plant species such
as Astragalus L., Oonopsis (Nutt.) Greene, Oryzopsis Michx., Xylorhiza Nutt., or Mentzelia L.
(Presser, 1999), can accumulate from 1000 to 10,000 mg/kg selenium (dry weight), being a
potential source of Se in the environment. There is no report in the literature with similar results
with bacterial strains tolerating and reducing such high concentrations. In this regard, S. *bentonitica* is presented herein as a potential microorganism to be used for decontamination of
highly polluted environment.

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3.2. XRD analysis

- The structure of the Se^{VI}-reduction products was characterized using the XRD technique. The 258 259 XRD patterns of the samples obtained after 24 h of incubation show the main peaks 260 characteristic of crystalline t-Se (COD-9008579) at 2θ values of 23.4° , 29.7° , 41.1° , 43.6° and 261 45.3°, and 51.7° (Figure 3), corresponding to the crystal planes (1 0 0), (0 1 1), (1 1 0), (1 0 2), (1 1 1), and (0 2 1), respectively, for initial Se^{VI} concentrations of 100, 150, and 200 mM. For 262 50 mM Se^{VI}, no peaks were observed, most probably due to the lower toxicity exerted by this 263 264 element at this concentration and non-formation of Se crystals. Interestingly, for 100 and 150 265 mM concentrations, peaks belonging to the t-Se phase can be observed in the absence of a 266 colour change. This suggests that the reduction rate increased, but not enough to produce the 267 characteristic red precipitates. No differences in the diffraction patterns were observed when 268 varying the incubation time from 17 to 48h at the same concentration (200 mM Se^{VI}) (Figure 269 **S2-Supplementary Material**)
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Figure 3. X-ray diffraction patterns of cultures of *S. bentonitica* supplemented with 50, 100, 150, and 200 mM Se^{VI} after 24h of incubation. Peaks for Se with a *t*-Se were detected for cultures supplemented with 100, 150, and 200 mM as indicated the corresponding crystal planes (black arrows). X-ray reflections of *t*-Se (COD-9008579) and m-Se (COD-9008581) obtained from Crystallography Open Database
 (http://www.crystallography.net/cod/) are shown at the bottom.

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3.3. FEG-ESEM and HAADF-STEM analysis

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280 A combination of FEG-ESEM and STEM/HAADF was used to determine the structure and physical properties of the Se^{VI} reduction products needed for the elucidation of biological Se^{VI} 281 282 biotransformation by the cells of S. bentonitica. Three-dimensional (3D) images obtained by a 283 FEG-ESEM system showed the presence of electron-dense nanorods mostly in the intracellular 284 space of the cells, but also extracellularly, after 24 and 48 h of incubation with 150 (Figure 4A-**B**) and 200 mM Se^{VI} (Figure 4C-E). Energy Dispersive X-Ray Analysis (EDX) clearly showed 285 286 these nanostructures to be composed of Se in addition to small peaks of sulfur (S) (Figure 4F). 287 The presence of both Se and S suggest the participation of S-containing compounds as 288 intermediates and biocatalysts in Se^{VI} reduction. A set of biochemical pathways implicating S-289 doped molecules such as glutathione (GSH), glutathione reductase (GR), glutathione peroxidase 290 (GP), or thioredoxin reductase (TrxR) in formation of Se⁰ through a series of redox reactions 291 have been widely reported before (Xu and Barton 2013; Eswayah et al. 2019). Interestingly, organic Se molecules such as selenocysteine amino acids are known to be essential in the catalytic activity of TrxR or GP forming a redox active disulfide bond in the active site (Hawkes and Alkan 2010). Both enzymes GR and TrxR could be involve in reduction to Se^0 in *S. bentonitica* since the genes encoding them are present in its genome (accession number: MKCZ00000000).

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298 The formation of Se nanoparticles having diverse morphologies is widely described (Presentato 299 et al. 2018; Fischer et al. 2020). Yet to the best of our knowledge, it is the first time that tubularshaped Se nanoparticles produced from a Se^{VI} reduction process could be observed 300 intracellularly. Interestingly, these relatively big and long nanorods can be observed 301 302 intracellularly without signs of cell lysis. The strain of study S. bentonitica has been highlighted 303 in recent years for its ability to reduce Se^{IV} to Se⁰ forming amorphous Se nanospheres and 304 trigonal nanorods (Ruiz-Fresneda et al. 2018; Ruiz-Fresneda et al. 2020). Based on such 305 findings, the authors invoked a biotransformation mechanism whereby intracellular amorphous 306 Se nanospheres are released to the extracellular space during the first 24-72 h, followed by 307 transformation to different Se allotropes (monoclinic and trigonal) after 144 h. Other authors 308 suggest a similar transformation pathway in different bacterial species: Bacillus subtillis and 309 Zoogloea ramigera (Wang et al. 2010; Srivastava and Mukhopadhyay 2013). In the present 310 study, however, Se nanorod formation was observed at 24 h and mainly located intracellularly, 311 evidencing that different transformation and interaction processes may occur in S. bentonitica when the initial Se source is Se^{VI} instead of Se^{IV} (Ruiz-Fresneda et al. 2018). Our findings 312 313 suggested the reduction and Se nanorod synthesis occurred intracellularly and are then released 314 somehow. Cytoplasmic enzymes from several microorganisms (Burkholderia fungorum 95, 315 Burkholderia fungorum DBT1, or Bacillus mycoides SeITE01) have been reported to be 316 involved in Se^{IV} reduction (Khoei et al. 2017; Lampis et al. 2014). Lampis et al. (2014) proposed thioredoxin reductase (TrxRed) systems and thiol-containing molecules with redox 317 activityin Se⁰ nanoparticle formation of *B. mycoides* SeITE01. Different export mechanisms 318 319 have been hypothesized by several authors causing SeNPs leak outside the cells. Some of them 320 suggested the SeNPs are released after cell death and lysis (Lampis et al. 2014). A few 321 extracellular nanorods were produced in our case, but no signs of bacterial lysis were observed 322 after 24h. Other export system proposed were based on secretion of the NPs through outer 323 membrane vesiculation and encapsulation (Kulp et al. 2010). For example, the gram-negative 324 bacterium Thauera selenatis use the protein Se factor A (SefA) for SeNPs binding, stabilization, 325 and secretion to the extracellular space (Debieux et al 2011). Further upcoming investigations 326 will help to elucidate the synthesis pathway for S. bentonitica.

328 From an industrial point of view, the faster production of crystalline SeNPs presented herein 329 points out a most promising crystalline NPs synthesis process compared to others. Wang et al. 330 (2010) reported a more rapid formation of crystalline Se nanorods (12 h) by B. subtilis, but with 331 the help of reactive agents such as ethanol. Similar Se tubular structures were produced by 332 anaerobic granular sludge after 18 h at 55°C (Jain et al. 2017). Interestingly, S. bentonitica can 333 form t-Se nanorods without the use of chemicals and room temperature, providing a more 334 ecologic and less expensive synthesis method. In terms of quantitative data, S. bentonitica is 335 highlighted herein as promising SeNPs producer as observed in the micrographs of thin sections 336 with up to 6 nanorods in a single cell (Figure 5).

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Figure 4. FEG-ESEM of cultures of *S. bentonitica* supplemented with 150 mM of Se^{VI} after 24 (A) and
48h (B), and with 200 mM of Se^{VI} after 24 (C and D) and 48h (E). EDX analysis of a single nanorod
(highlighted area in panel in E) confirmed the presence of Se (F).

343 Ultrathin sections observed with an HAADF-STEM system allowed for further structural 344 characterization of the Se nanostructures produced by *S. bentonitica*. By means of this technique 345 we again observed that the predominantly Se nanostructure was found intracellularly in the form 346 of nanorods (**Figure 5A**). Element-distribution maps indicated the presence of both Se and S in 347 the nanorods (**Figure 5B-D**), strongly supporting the possible relevance of thiol-containing 348 proteins in the reduction pathway.



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Figure 5. HAADF-STEM micrographs of a thin section showing Se nanostructures produced by S.
 bentonitica after 24 h of incubation with Se^{VI} (A). EDX element-distribution maps showing their Se and S
 elemental composition (B-D).

354 A combination of tools of the HAADF-STEM system --including SAED (Selected Area 355 Electron Diffraction), FFT (Fast-Fourier Transform), and STEM high resolution (HRTEM)-356 attested to the crystalline nature of the Se nanostructures. SAED and FFT patterns from a 357 selected Se nanorod revealed signs of crystallization (Figure 6A-C); specifically, three different 358 d-spacings of 0.29, 0.37, and 0.48 nm could correspond to crystal planes of monoclinic-Se (m-359 Se) (Pinel-Cabello et al. 2021) (Figure 6B). The space lattices 0.29 and 0.37 nm are 360 characteristic as well in trigonal Se (t-Se) according to the American Mineralogist Crystal 361 Structure Database (http://rruff.geo.arizona.edu) [accessed December, 2021]. Yet the detection 362 of the 0.48 nm d-spacing confirmed the monoclinic structure of the nanorods. Similar results were obtained recently for Se nanorods produced by S. bentonitica contacted with Se^{IV} (Pinel-363 364 Cabello et al. 2021). HRTEM analyses, along with the FFT from a nanorod selected area, agree 365 with the SAED pattern, revealing the presence of 0.29-0.3 and 0.38 nm lattice spacings (Figure 366 6C-F). As mentioned above, these spacings could correspond to different crystal planes of both 367 *m*-Se and *t*-Se. More specifically, the d-spacing of 0.3 nm could correspond to planes (1 0 1 or 0 368 1 1) of t-Se and many different planes of m-Se. The d-spacing of 0.37 nm may correspond to the 369 plane (1 0 0) of t-Se, and (0 2 2) or (4 0 0) of m-Se. Therefore, the existence of t-Se should not 370 be discarded since common d-spacings (0.3 and 0.38 nm) between t-Se and m-Se were 371 observed. Even the co-existence of mixed crystal phases in one same nanocrystal is possible in 372 our Se nanostructures, as reported before for Sn nanoparticles (Haq et al. 2019). In fact, t-Se 373 was previously detected in our samples with XRD (see section 3.2.). Naturally, XRD is a more 374 representative and precise technique, as the bulk sample is measured, whereas under HAADF-375 STEM some selected crystals were analyzed. Ultimately, these results could indicate the 376 biotransformation of *m*-Se (as a transitional allotropic form) to *t*-Se, as was reported for Se^{IV} 377 (Ruiz-Fresneda et al. 2018). Still, the results clearly underline the importance of a 378 multidisciplinary approach, combining microscopic and spectroscopic techniques, to 379 exhaustively characterize nanoparticle structures.



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Figure 6. HAADF-STEM micrograph of Se nanostructures produced by *S. bentonitica* incubated with
 200 mM Se^{VI} for 24 h (A). SAED pattern (B), FFT pattern (C), and HRTEM image (D) derived from a
 single nanostructure (highlighted area in panel A). Magnified HRTEM images (E and F) corresponding to
 highlighted area in D.

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386 3.4. XAS analysis

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388 X-ray absorption spectroscopy (XAS) spectra of the Se^{VI}-reduction products obtained after 17, 389 24, and 48 h of incubation with *S. bentonitica* (200 mM Se^{VI} as initial concentration) provided 390 more detailed structural information concerning the local coordination environment, as well as 391 the oxidation state of the Se in the studied samples. The X-ray absorption near-edge structure 392 (XANES) region clearly showed that the local coordination of Se is dominated by Se⁰ at all 393 incubation times, as indicated by the maximum peak positions obtained at 12659.3 eV, 12659.3 394 eV, and 12659.2 eV after 17, 24, and 48h incubating (**Figure 7A**) (**Figure S3 and S4**- Supplementary Material). Generally, most of the bioreduced SeNPs in solid form consisted of
Se in the zero-valent oxidation state (Zhang et al. 2012; Vogel et al. 2018; Ruiz-Fresneda et al.
2020).

The extended X-ray absorption fine structure (EXAFS) spectra of the Se^{VI}-reduction products 398 and Se foil (as Se⁰ reference compound), along with their corresponding Fourier transforms 399 (FT) and fit parameters of the obtained spectra, are respectively presented in Figure 7B-C, 400 401 Figure S5-Supplementary Material, and Table 1. FT peak distances are reported in units of Å 402 and are expressed as $R + \Delta R$. The fit of the EXAFS spectra of the 3 experimental samples 403 indicated the presence of one Se-Se coordination shell at a bond distance of about $2.35-2.37 \pm$ 404 0.02\AA (Table 1). In previous XAS analyses obtained for the Se nanostructures produced by S. 405 bentonitica with Se^{IV} this parameter ranged between 2.33 and 2.34 Å and were attributed to amorphous Se in view of the literature (Eswayah et al. 2017; Vogel et al. 2018; Ruiz-Fresneda 406 407 et al. 2020). Slight increases of Se-Se bond distance values could indicate a crystallization 408 process, while slight decreases in this frame point to an amorphization. Such is the case of the 409 experiments by Zhao et al. (2004) and Breynaert et al. (2008), who found an amorphization 410 process of crystalline Se upon a decrease from 2.37 to 2.35 Å. The increase of the Se-Se bond 411 distances with incubation time found in our samples accordingly suggests a heightened 412 structural order and therefore crystallization of Se⁰ over time. In view of the literature discussed 413 above, the bond distances of 2.35 and 2.36 Å found after 17 and 24 h could correspond to a 414 mixture of amorphous and trigonal Se, while the value of 2.37 Å found for 48 h samples might 415 mark a Se crystalline phase mainly as trigonal Se (Zhao et al. 2004; Breynaert et al. 2008). 416 However, our results are not conclusive about a dependent time crystallization process.



417

Figure 7. XANES (A), EXAFS (B), and their corresponding FT spectra (C) of Se reference compounds [Se^{VI}
(Na₂SeO₄), Se^{IV} (Na₂SeO₃), Se⁰ (Se foil), and Se^{-II} (SeS₂)] and *S. bentonitica* samples incubated with 200 mM Se^{VI} at different incubation times (17, 24, and 48 h).

422

Table 1. EXAFS structural parameters of the Se foil and the Se^{VI}-reduction products.

Sample	Shell	N ^a	R[Å] ^b	$\sigma^2 [\mathring{A}^2]^c$	$\Delta E[eV]$
Se foil	Se-Se ₁	3.3 ± 0.2	2.37	0.0043	-5.6
Se ^{VI} - Cells-17h	Se-Se ₁	0.7 ± 0.1	2.35	0.0031	-3.1
Se ^{VI} - Cells-24h	Se-Se ₁	0.7 ± 0.1	2.36	0.003	-3.8
Se ^{VI} - Cells-48h	Se-Se ₁	0.6 ± 0.1	2.37	0.0028	-1.2

423 ^a Errors in coordination numbers are ±25% and standard deviations as estimated by EXAFSPAK

424 b Errors in distance are ± 0.02 Å

425 ° Debye-Waller factor

426

427 3.5. Biological production of volatile Se compounds

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429 Headspace analysis using GC-MS of extracted volatiles from 100 mM Se^{VI}-treated cultures 430 indicated the formation of dimethyl selenide (DMSe), dimethyl diselenide (DMDSe) and 431 dimethyl selenenyl sulphide (DMSeS) by the cells after 144 h incubating (Figure 8). In 432 contrast, no volatile Se-containing species were detected in biotic and abiotic controls (Figure 433 S6-Supplementary Material). The Se^{VI} initial concentration of 100 mM was selected for this 434 analysis to increase the amount of volatiles that are produced. At this concentration, trigonal Se^0 435 red precipitates derived from reduction processes were not as observable as for 200 mM, 436 suggesting that other interaction mechanisms such as volatilization may have happened. The 437 formation of Se volatile compounds presented herein revealed that volatilization is involved as a biotransformation mechanism in S. bentonitica, in addition to reduction. It is worth noting that 438 no Se volatile species were produced by the cells when the Se^{VI} initial concentration was 2 mM 439 (Figure 8). This suggests, as discussed in previous sections, that a very high Se^{VI} initial 440 441 concentration is needed to let biotransformation occur, whether for volatilization or reduction 442 process.

443

Interestingly, this bacterium does not produce DMSe when amended with Se^{IV} instead of Se^{VI} 444 445 (Ruiz-Fresneda et al. 2020), thereby suggesting a different biotransformation pathway 446 depending on the oxidation state of Se, even in the same microorganism. Thus, the versatile role 447 of S. bentonitica in Se volatilization depends on the physico-chemical conditions. Not only is 448 the Se oxidation state important, but also the type of SeNPs produced can influence Se 449 volatilization. Otsuka and Yamashita (2020) showed that the structural nature of SeNPs affects 450 the Se volatilization rates in P. stutzeri NT-I, a bacterial strain producing higher amounts of 451 DMDSe when amorphous SeNPs were used as substrate as opposed to crystalline SeNPs. The 452 differences between the nanoparticles formed by S. bentonitica (nanospheres and nanorods from Se^{IV}; only nanorods with Se^{VI}) might therefore condition the volatile compounds produced. 453



456 Figure 8. GC-MS chromatograms of the headspace gas of *S. bentonitica* cultures supplemented with Se^{VI}
457 (2 and 100 mM) and after 144 h of incubation. All GC-MS chromatograms were obtained by selecting the
458 80 m/z ion specific for selenium.

460 No evident biochemical pathway is expounded in the literature to clearly explain how Se 461 biovolatilization occurs. Most studies indicate that Se volatilization relies on reduction and 462 methylation reactions when the initial form is a Se oxyanion. Accordingly, Se^{VI} or Se^{IV} are firstly reduced to Se⁰ or Se^{-II}, and subsequently methylated to DMSe (Eswayah et al. 2016). 463 464 Methylation appears to be the most crucial or differential step, since a great number of reactions 465 and intermediates ---MeSeH, dimethyl disulfide (DMDS), dimethyl selenone, and Se 466 aminoacids (SeMet and SeCys) have been suggested to be involved (Chasteen and Bentley 467 2003; Winkel et al. 2015). DMDS may be involved in the formation of DMSeS through its 468 reaction with DMDSe (Chasteen 1993). The presence of these three compounds (DMDS, 469 DMDSe, and DMSeS) in our samples comes to support this hypothesis as a possible mechanism 470 active in S. bentonitica and the role of S-containing enzymes in the reduction process. Current 471 evidence would suggest that the biochemical pathways responsible for Se volatilization remains 472 to be elucidated, and it is likely that a variety of mechanisms could be involved depending on 473 the given organism and the form of Se. From an environmental perspective, Se biomethylation 474 is held to constitute a detoxification mechanism, as it enables the removal and transformation of 475 toxic inorganic Se precursors toward less toxic methylated forms (Wilber 1980; Ranjard et al. 476 2003). This fact in itself points to a positive role of S. bentonitica in the bioremediation of Se 477 contaminated sites. In the context of deep disposal of radioactive waste, however, Se production 478 by S. bentonitica could compromise the integrity of such repositories by causing gas 479 overpressure.

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455

481 According to the results obtained both enzymatic reduction and volatilization are involved in Se^{VI} interactions with *S. bentonitica*. When Se^{VI} initial concentration is below 100 mM Se^{VI} is 482 probably reduced to Se^{IV} by S-containing oxidoreductases (e.g. GR or TrxR) and then removed 483 484 through efflux transporter to the extracellular space since no red precipitates were observable. 485 This hypothesis agrees with preliminary transcriptomics studies (to be published). However, at concentrations over 100 mM the cells are capable of reducing Se^{VI} to Se⁰ and Se^{-II} volatile 486 487 compounds due to these more stressful conditions. The experimental data suggested the 488 reduction to Se⁰ nanorods occurs intracellularly, before being released probably through vesicle 489 secretion. The observation of organic matter surrounding some nanorods could be a sign of this 490 (Figure 4D).

491

492

3.6. Environmental and industrial significance of the different Se allotropes produced

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494 To sum up, the combination of spectroscopic and microscopic techniques showed the great potential of *S. bentonitica* to reduce Se^{VI} to Se⁰ as crystalline nanorods with different allotropes 495 496 (monoclinic and trigonal Se). The presence of t-Se crystalline phase in the biogenic nanorods 497 was demonstrated by XRD. Lattice d-spacing values corresponding to m-Se could be observed 498 by high resolution TEM analysis. These findings suggest a crystallization process from m-Se to 499 the most thermodynamically stable phase t-Se. The XANES region of the XAS spectra 500 confirmed the presence of the zero valent oxidation state of the Se reduction products, while the 501 EXAFS region of the spectra showed the possible presence of different Se allotropes 502 (amorphous, monoclinic, and trigonal). In addition, these Se nanorods could be stabilized by their interactions with organic matter as it was demonstrated for Se nanostructures derived from 503 504 reduction of Se^{IV} by this bacterial strain (Ruiz-Fresneda et al. 2020). However, Se-C 505 coordination shell does not seem to be visible in the EXAFS spectra of the studied samples. 506 Therefore, further studies in the characterization of organic matter coating the Se nanorods 507 should be performed using spectroscopic techniques like Infrared spectroscopy.

508

509 Ultimately, the results presented here point to *S. bentonitica* as an interesting bacterial model for 510 bioremediation systems and for its potential as a new, green, and faster (only 24 h) way to 511 produce crystalline SeNPs of substantial interest for industrial and medical applications.

- 513 **4.** Conclusions
- 514

512

515 The present work describes, for the first time, the ability of the bacterium *S. bentonitica* to 516 reduce Se^{VI} to Se⁰ nanorods and Se volatile compounds. The results clearly demonstrate a higher 517 tolerance capacity of this bacterial species against Se^{VI} when compared to Se^{IV} experiments 518 performed in previous research. Differences in the structure, location, and formation speed of 519 the Se reduction products when Se^{VI} is the initial source suggest a different biotransformation pathway than the one proposed for Se^{IV}. Still, more research is needed to fully understand the 520 521 specific biotransformation processes. An exhaustive combination of spectroscopic and 522 microscopic techniques applied here revealed important structural and chemical data, including 523 the oxidation state and the crystalline phases of bioreduced Se. The zero-valent oxidation state, 524 tubular shape, and crystallinity of the Se nanostructures produced would uphold S. bentonitica 525 as a potential bioremediation candidate in contaminated environments. The formation of volatile 526 methylated species furthermore points to a positive role in the removal of toxic Se from polluted 527 soils and waters, in terms of atmospheric bioremediation. Finally, S. bentonitica is put forth as a 528 candidate for novel, environmentally friendly, and relatively quick Se nanorod fabrication.

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531

530 **CRediT authorship contribution statement**

532 Miguel A. Ruiz-Fresneda: Conceptualization, Methodology, Validation, Formal Analysis, 533 Investigation, Writing-Original Draft, Writing-Review & Editing, Visualization. María V. 534 Fernández-Cantos: Conceptualization, Methodology, Validation, Formal Analysis, 535 Investigation, Writing-Review & Editing. Jaime Gómez-Bolívar: Methodology, Validation. 536 Abdurrahman S. Eswayah: Methodology, Validation. Philip H. E. Gardiner: 537 Conceptualization, Resources, Writing-Review & Editing, Supervision. Maria Pinel-Cabello: 538 Formal analysis, Writing-Review & Editing. Pier L. Solari: Methodology, Formal Analysis. 539 Mohamed L. Merroun: Conceptualization, Methodology, Formal Analysis, Resources, 540 Writing-Original Draft, Writing-Review & Editing, Supervision, Project Administration, 541 Funding Acquisition.

- 542
- 543 **Declaration of CompetingInterest**
- 544
- 545 Non-declared
- 546

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548

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