

**The prevalence of onchocerciasis in Africa and Yemen,
2000–2018: a geospatial analysis**

SCHMIDT, C.A., CROMWELL, E.A., HILL, E., DONKERS, K.M., SCHIPP, M.F., JOHNSON, K.B., PIGOTT, D.M., LBD 2019 NEGLECTED TROPICAL DISEASES COLLABORATORS, and HAY, Simon I.

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Additional File 1: Supplementary Appendix for *The prevalence of onchocerciasis in Africa and Yemen, 2000–2018: A geospatial analysis*

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1.0 GATHER compliance

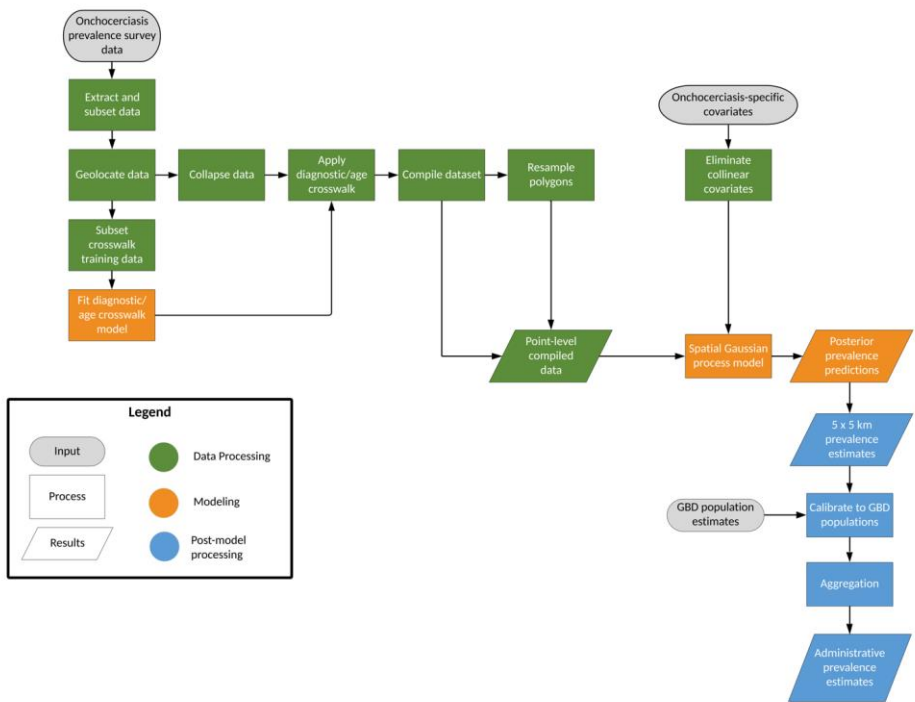
Supplementary Table 1. Guidelines for Accurate and Transparent Health Estimates Reporting (GATHER) checklist.

Item #	Checklist item	Reported on page #
Objectives and Funding		
1	Define the indicator(s), populations (including age, sex, and geographic entities), and time period(s) for which estimates were made.	Main Text: Introduction, Methods (Data inputs); Methods (Covariates)
2	List the funding sources for the work.	Main text: Acknowledgments
Data Inputs		
<i>For all data inputs from multiple sources that are synthesized as part of the study:</i>		
3	Describe how the data were identified and how the data were accessed.	Main Text: Methods (Data inputs); Supplementary Information: 3.2 Systematic review
4	Specify the inclusion and exclusion criteria. Identify all ad-hoc exclusions.	Main Text: Methods (Data inputs); 3.0 Supplementary data; Supplementary Information: 3.2 Systematic review
5	Provide information on all included data sources and their main characteristics. For each data source used, report reference information or contact name/institution, population represented, data collection method, year(s) of data collection, sex and age range, diagnostic criteria or measurement method, and sample size, as relevant.	Supplementary Information: 3.0 Supplementary data
6	Identify and describe any categories of input data that have potentially important biases (e.g., based on characteristics listed in item 5).	Main text: Discussion (Limitations)
<i>For data inputs that contribute to the analysis but were not synthesized as part of the study:</i>		
7	Describe and give sources for any other data inputs.	Main text: Methods (Geospatial covariates), Supplementary Information: 4.0 Supplementary covariates
<i>For all data inputs:</i>		
8	Provide all data inputs in a file format from which data can be efficiently extracted (e.g., a spreadsheet rather than a PDF), including all relevant meta-data listed in item 5. For any data inputs that cannot be shared because of ethical or legal reasons, such as third-party ownership, provide a contact name or the name of the institution that retains the right to the data.	Available at (GHDx link will be added upon review) Supplementary Information: 3.0 Supplementary data
9	Provide a conceptual overview of the data analysis method. A diagram may be helpful.	Main text: Methods (Age and diagnostic adjustment), Supplementary Information: 5.1 Age and diagnostic crosswalks; Supplementary Information: Table 6: data used in estimation of age and diagnostic crosswalk
10	Provide a detailed description of all steps of the analysis, including mathematical formulae. This description should cover, as relevant, data cleaning, data pre-processing, data adjustments and weighting of data sources, and mathematical or statistical model(s).	Main text: Methods (Geospatial analysis), Supplementary Information: 5.3 Geostatistical model
11	Describe how candidate models were evaluated and how the final model(s) were selected.	Main text: Methods (Geospatial analysis), Supplementary Information: 5.4 Model validation
12	Provide the results of an evaluation of model performance, if done, as well as the results of any relevant sensitivity analysis.	Main text: Methods (Geospatial analysis), Supplementary Information: 5.4 Model validation
13	Describe methods for calculating uncertainty of the estimates. State which sources of uncertainty were, and were not, accounted for in the uncertainty analysis.	Main text: Methods (Geospatial analysis), Supplementary Information: 5.3.4. Model description
14	State how analytic or statistical source code used to generate estimates can be accessed.	Available at (GHDx link will be added upon review)
Results and Discussion		
15	Provide published estimates in a file format from which data can be efficiently extracted.	Raster files for spatial data and CSVs of estimates available at (GHDx link will be added upon review)
16	Report a quantitative measure of the uncertainty of the estimates (e.g., uncertainty intervals).	Supplementary Information: Supplementary Figure 11 and Table 9

17	Interpret results in light of existing evidence. If updating a previous set of estimates, describe the reasons for changes in estimates.	Main text: Discussion (Strengths)
18	Discuss limitations of the estimates. Include a discussion of any modelling assumptions or data limitations that affect interpretation of the estimates.	Main text: Discussion (Limitations)

2.0 Supplementary discussion

This document outlines the major data processing, modelling, and validation steps for the onchocerciasis prevalence analysis described in the main text (Supplementary Figure 1). We present a detailed description of model inputs, including data coverage, covariate sources, and geo-referencing. The geospatial model is described along with model validation metrics.



Supplementary Figure 1. Flowchart of major steps in data processing and modelling of onchocerciasis prevalence.

3.0 Supplementary data

In the following section, we present a detailed summary of the data inputs used to estimate the prevalence of onchocerciasis in Africa and Yemen. Broadly, we aimed to include all published sources of onchocerciasis infection prevalence, as well as routine programme monitoring data collected to monitor progress toward onchocerciasis elimination. Data inputs were retained for analysis if they could be accurately geo-referenced.

3.1 Geographical restrictions

Supplementary Table 2 lists all countries included in our MBG (model-based geostatistical) modelling region. Inclusion was partially based on ESPEN (Expanded Special Project for Elimination of Neglected Tropical Diseases)

onchocerciasis endemicity classifications,(1) with extensions to some neighbouring countries outside the WHO (World Health Organization) AFRO (Regional Office for Africa) region which have evidence of onchocerciasis endemicity or uncertain endemicity status (Sudan, Somalia, and Yemen). Portions of Mauritius, Namibia, and Zambia are considered by ESPEN as candidates for future elimination mapping but were not included in our geospatial model. Conversely, The Gambia is considered non-endemic but we included it in our model for geographical continuity. Countries in the Americas with historical or residual onchocerciasis burdens were not modelled due to their highly localised endemicity.

Supplementary Table 2. Geographical restrictions.

The geographical definition of the modelling region is indicated, with country-level onchocerciasis endemicity status per ESPEN, and which countries were included in our geospatial analysis from 2000 to 2018. Endemicity classifications were derived from ESPEN data at the level of intervention units (IU) (retrieved 14 February 2020). Countries outside the WHO AFRO region are not covered by the scope of ESPEN; countries in the WHO EMRO (Eastern Mediterranean Regional Office) region are listed here if they are in Africa or have evidence of onchocerciasis endemicity or uncertain status. We considered countries endemic if they had at least one intervention unit that was flagged as endemic by ESPEN. Number of observations: The number of data rows in the final dataset for a given country.

Location	ESPEN endemicity	Included in 2000–2018 model?	Number of Observations
<i>AFRO</i> (1)			
Algeria	Non-endemic	no	-
Angola	Endemic	yes	876
Benin	Endemic	yes	300
Botswana	Non-endemic	no	-
Burkina Faso	Endemic	yes	72
Burundi	Endemic	yes	263
Cabo Verde	Non-endemic	no	-
Cameroon	Endemic	yes	1216
Central African Republic	Endemic	yes	1079
Chad	Endemic	yes	753
Comoros	Non-endemic	no	-
Republic of the Congo	Endemic	yes	490
Côte d'Ivoire	Endemic	yes	95
Democratic Republic of the Congo	Endemic	yes	4574
Equatorial Guinea	Endemic	yes	341
Eritrea	Non-endemic	no	-
Eswatini	Non-endemic	no	-
			6

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Location	ESPEN endemicity	Included in 2000–2018 model?	Number of Observations
Ethiopia	Endemic	yes	1062
Gabon	Endemic	yes	157
The Gambia	Non-endemic	yes	0
Ghana	Endemic	yes	78
Guinea	Endemic	yes	193
Guinea-Bissau	Endemic	yes	0
Kenya	Consider oncho elimination mapping	yes	94
Lesotho	Non-endemic	no	-
Liberia	Endemic	yes	114
Madagascar	Non-endemic	no	-
Malawi	Endemic	yes	333
Mali	Endemic	yes	160
Mauritania	Non-endemic	no	-
Mauritius	Mapping gap (one district)	no	-
Mozambique	Endemic	yes	291
Namibia	Consider oncho elimination mapping (some districts)	no	-
Niger	Endemic	yes	0
Nigeria	Endemic	yes	3445
Rwanda	Consider oncho elimination mapping	yes	90
São Tomé and Príncipe	Non-endemic	yes	0
Senegal	Endemic	yes	67
Seychelles	Non-endemic	no	-
Sierra Leone	Endemic	yes	223
South Africa	Non-endemic	no	-
South Sudan	Endemic	yes	466
Togo	Endemic	yes	160
Uganda	Endemic	yes	620
Tanzania	Endemic	yes	478

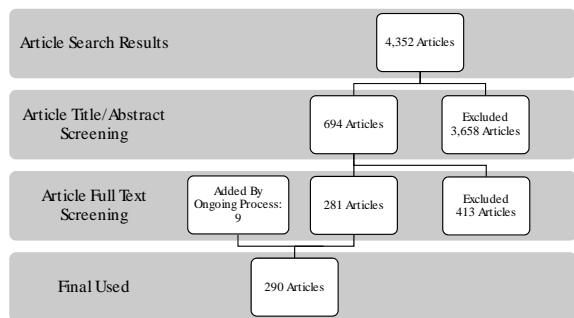
Location	ESPEN endemicity	Included in 2000–2018 model?	Number of Observations
Zambia	Consider oncho elimination mapping (some districts)	no	-
Zimbabwe	Non-endemic	no	-
<i>EMRO</i> (2)			
Djibouti	-	no	-
Egypt	-	no	-
Libya	-	no	-
Morocco	-	no	-
Somalia	-	yes	0
Sudan	(Endemic)	yes	26
Tunisia	-	no	-
Yemen	(Endemic)	yes	0

3.2 Systematic review

Articles related to onchocerciasis were found by searching PubMed, Scopus, and Web of Science using the following keywords: “Oncho”, “river blindness”, “O. Volvulus”, “robles disease”, “blinding filariasis”, “coast erysipelas”, and “sowda”. A systematic review of these reports, all published before July 7, 2017, was then conducted. A second round of formal reviews was undertaken on June 6, 2019, to cover new articles published after July 7, 2017, and followed the same process as the first round.

3.2.1 Systematic review data processing

The systematic review process is illustrated in Supplementary Figure 2. Throughout the systematic review we excluded publications that met the following criteria: no measurement of onchocerciasis prevalence, data collected before 1985, case-control studies, qualitative research publications, duplicative data from cohort studies, hospital-based studies, and publications that did not report the location of data collection. The search identified 4048 publications, which were reduced to 579 after screening titles and abstracts. A full-text review was completed for the remaining publications. The full-text review yielded 259 publications which met the inclusion criteria and were extracted. The literature review was updated on June 6, 2019, by searching PubMed with the same search string used in 2017 for articles published after July 7, 2017. The search returned 304 results, which were narrowed to 115 articles after screening titles and abstracts. A full-text review resulted in 22 articles that were eligible for extraction. Additional publications were identified outside of the literature review and screened for inclusion by the same criteria, up through April 14, 2020. The final dataset drew from 290 articles. Overall, 16 096 datapoints were extracted and geo-located from 290 publications. Data extracted from each source included survey year, age-range of individuals tested, diagnostic information, sample size, number of individuals tested positive, and sampling strategy details.



Supplementary Figure 2. Onchocerciasis article review and data extraction flowchart.

Each step of the extraction process is outlined, from article identification and screening to extraction, including the number of articles or records that were processed or removed in each step before reaching the final dataset. Additional articles outside of the literature review were identified and screened for inclusion (“ongoing process”) on an ongoing basis up until April 14, 2020.

3.3 Geo-positioning

Geographical information associated with onchocerciasis prevalence data were verified and geo-located to ensure accuracy. Data associated with locations smaller than 5×5 km were treated as points and geo-located as latitude/longitude coordinates. Coordinates provided by a data source were mapped to ensure that the coordinates were located in the correct administrative units. If coordinates were not reported, points were geolocated and vetted by comparing results from Google Maps, Fuzzy Gazetteer, Geonames, and Open Street Map. Locations larger than 5×5 km were treated as polygon data and geo-located to the most granular administrative boundary possible (most commonly district level). Custom shapefiles were created to geolocate areas that did not align with administrative boundaries. In the event that a literature source only included a map of locations sampled without any other information, ArcGIS software was utilised to overlay the map onto existing administrative boundaries, and location coordinates or custom polygons were manually created and recorded. Prevalence data that were reported for administrative areas were matched to their appropriate polygon by searching our administrative shapefile database. If place names were unidentifiable across multiple shapefile libraries or geo-referencing sources, they were excluded from the analysis.

3.4 Data processing

Data were excluded from the analysis for the following reasons: did not report survey year ($N = 8236$); did not report post-1988 data ($N = 2648$); did not report skin snip microscopy or nodule prevalence ($N = 12\ 121$); did not report sample size or prevalence ($N = 30$); or did not report data for countries in the modelling region or could not be accurately georeferenced ($N = 580$). Duplicate records were identified and excluded ($N = 1592$); these generally reflected different data sources (eg, literature extraction and ESPEN) which contained the same survey, as judged by identical year, location, sample size, cases, and reference when available, although there was some deduplication within individual data sources. Prior to polygon resampling (see section 5.2), the dataset consisted of 18 116 georeferenced prevalence observations, consisting of 17 896 point-referenced observations and 220 areal observations. A total of 14 314 observations represented prevalence based on nodule examinations, and 3802 observations represented microfiladermia prevalence as measured by skin snip assays. After resampling, the full modelling dataset consisted of 20 124 georeferenced datapoints.

Supplementary Table 3 provides citations for data sources used in our onchocerciasis MBG model. The geographical coverage of the final dataset is summarised in Supplementary Figure 3.

Supplementary Table 3. Citations for data inputs.

The NID is a unique identifier cataloguing all data inputs in the Global Health Data Exchange (<http://ghdx.healthdata.org>). Note: Records are listed here in alphabetical order by geography, but some sources provided data for multiple countries; such sources are listed here only once.

NID	Geographies	Citation
332798	Angola	Carne B, Ntsoumou-Madzou V, Samba Y, Yebakima A. Prevalence of depigmentation of the shins: a simple and cheap way to screen for severe endemic onchocerciasis in Africa. <i>Bull World Health Organ.</i> 1993; 71(6): 755–8.
334477	Benin	Gallin M, Adams A, Kruppa TF, Gbaguidi EA, Massougbedji A, Sadeler BC, Brattig N, Erttmann KD. Epidemiological studies of onchocerciasis in southern Benin. <i>Trop Med Parasitol.</i> 1993; 44(2): 69–74.
125405	Burkina Faso	Kabore JK, Cabore JW, Melaku Z, Druet-Cabanac M, Preux PM. Epilepsy in a focus of onchocerciasis in Burkina Faso. <i>Lancet.</i> 1996; 347(9004): 836.
334475	Burkina Faso	De Sole G, Remme J. Importance of migrants infected with <i>Onchocerca volvulus</i> in west African river valleys protected by 14 to 15 years of Simulium control. <i>Trop Med Parasitol.</i> 1991; 42(2): 75–8.
332903	Burkina Faso, Côte d'Ivoire	Toè L, Adjami AG, Boatin BA, Back C, Alley ES, Dembélé N, Brika PG, Pearlman E, Unnasch TR. Topical application of diethylcarbamazine to detect onchocerciasis recrudescence in west Africa. <i>Trans R Soc Trop Med Hyg.</i> 2000; 94(5): 519–25.
334481	Burundi	Newell E d. Comparison of the use of skin scarification and skin biopsies to determine the prevalence and intensity of <i>Onchocerca volvulus</i> infection. <i>Ann Trop Med Parasitol.</i> 1997; 91(6): 633.
338571	Burundi	Newell ED, Hicuburundi B, Ndimuruvugo N. [Endemicity and clinical manifestations of onchocerciasis in the province of Bururi, Burundi]. <i>Trop Med Int Health.</i> 1997; 2(3): 218–26.
338573	Burundi	Newell ED, Ndimuruvugo N, Nimpa D. [Endemicity and clinical manifestations of onchocerciasis in the provinces of Cibitoke and Bubanza (Burundi)]. <i>Bull Soc Pathol Exot.</i> 1997; 90(5): 353–7.
136492	Cameroon	Cho-Ngwa F, Amambua AN, Ambele MA, Titanji VPK. Evidence for the exacerbation of lymphedema of geochemical origin, podoconiosis, by onchocerciasis. <i>J Infect Public Health.</i> 2009; 2(4): 198–203.
159316	Cameroon	Matthews GA, Dobson HM, Nkot PB, Wiles TL, Birchmore M. Preliminary examination of integrated vector management in a tropical rainforest area of Cameroon. <i>Trans R Soc Trop Med Hyg.</i> 2009; 103(11): 1098–104.
324729	Cameroon	Kamga G-R, Dissak-Delon FN, Nana-Djeunga HC, Biholong BD, Mbigha-Ghogomu S, Souopgui J, Zoure HGM, Boussinesq M, Kamgno J, Robert A. Still mesoendemic onchocerciasis in two Cameroonian community-directed treatment with ivermectin projects despite more than 15 years of mass treatment. <i>Parasites Vectors.</i> 2016; 9(1): 581.
327960	Cameroon	Wanji S, Kengne-Ouafo JA, Esum ME, Chounna PWN, Tendongfor N, Adzemye BF, Eyong JEE, Jato I, Datchoua-Poutcheu FR, Kah E, Enyong P, Taylor DW. Situation analysis of parasitological and entomological indices of onchocerciasis transmission in three drainage basins of the rain forest of South West Cameroon after a decade of ivermectin treatment. <i>Parasites Vectors.</i> 2015; 8: 202.
327968	Cameroon	Katabarwa MN, Eyamba A, Chouaibou M, Enyong P, Kuété T, Yaya S, Yougouda A, Baldiagui J, Madi K, Andze GO, Richards F. Does onchocerciasis transmission take place in hypoendemic areas? a study from the North Region of Cameroon. <i>Trop Med Int Health.</i> 2010; 15(5): 645–52.

NID	Geographies	Citation
327992	Cameroon	Katabarwa MN, Eyamba A, Nwane P, Enyong P, Kamgno J, KuetÃ© T, Yaya S, Aboutou R, Mukenge L, Kafando C, Siaka C, Mkpouwoueiko S, Ngangue D, Biholong BD, Andze GO. Fifteen years of annual mass treatment of onchocerciasis with ivermectin have not interrupted transmission in the west region of cameroon. <i>J Parasitol Res.</i> 2013; 2013: 420928.
328065	Cameroon	Katabarwa MN, Eyamba A, Nwane P, Enyong P, Yaya S, BaldiagÃ© J, Madi TK, Yougouda A, Andze GO, Richards FO. Seventeen years of annual distribution of ivermectin has not interrupted onchocerciasis transmission in North Region, Cameroon. <i>Am J Trop Med Hyg.</i> 2011; 85(6): 1041-9.
328095	Cameroon	Kamga HLF, Shey DN, Assob JCN, Njunda AL, Nde Fon P, Njem PK. Prevalence of onchocerciasis in the Fundong Health District, Cameroon after 6 years of continuous community-directed treatment with ivermectin. <i>Pan Afr Med J.</i> 2011; 10: 34.
332745	Cameroon	Pion SDS, Clarke P, Filipe J a. N, Kamgno J, Gardon J, BasÃ©ez M-G, Boussinesq M. Co-infection with <i>Onchocerca volvulus</i> and <i>Loa loa</i> microfilariae in central Cameroon: are these two species interacting?. <i>Parasitology.</i> 2006; 132(Pt 6): 843–54.
332783	Cameroon	Boussinesq M, Chippaux JP, Ernoult JC, Quillevere D, Prod'hon J. Effect of repeated treatments with ivermectin on the incidence of onchocerciasis in northern Cameroon. <i>Am J Trop Med Hyg.</i> 1995; 53(1): 63–7.
332786	Cameroon	Kollo B, Mather FJ, Cline BL. Evaluation of alternate methods of rapid assessment of endemicity of <i>Onchocerca volvulus</i> in communities in southern Cameroon. <i>Am J Trop Med Hyg.</i> 1995; 53(3): 243–7.
332788	Cameroon	Ngoumou P, Walsh JF, Mace JM. A rapid mapping technique for the prevalence and distribution of onchocerciasis: a Cameroon case study. <i>Ann Trop Med Parasitol.</i> 1994; 88(5): 463–74.
332790	Cameroon	Mendoza Aldana J, Piechulek H, Maguire J. Forest onchocerciasis in Cameroon: its distribution and implications for selection of communities for control programmes. <i>Ann Trop Med Parasitol.</i> 1997; 91(1): 79–86.
332840	Cameroon	Ayong LS, Tume CB, Wembe FE, Simo G, Asonganyi T, Lando G, Ngu JL. Development and evaluation of an antigen detection dipstick assay for the diagnosis of human onchocerciasis. <i>Trop Med Int Health.</i> 2005; 10(3): 228-33.
332848	Cameroon	Wanji S, Tendongfor N, Esum M, Ndindeng S, Enyong P. Epidemiology of concomitant infections due to <i>Loa loa</i> , <i>Mansonella perstans</i> , and <i>Onchocerca volvulus</i> in rain forest villages of Cameroon. <i>Med Microbiol Immunol.</i> 2003; 192(1): 15-21.
332893	Cameroon	Kamgno J, Boussinesq M. [Hyperendemic loiasis in the Tikar plain, shrub savanna region of Cameroon]. <i>Bull Soc Pathol Exot.</i> 2001; 94(4): 342–6.
332897	Cameroon	Esum M, Wanji S, Tendongfor N, Enyong P. Co-endemicity of loiasis and onchocerciasis in the South West Province of Cameroon: implications for mass treatment with ivermectin. <i>Trans R Soc Trop Med Hyg.</i> 2001; 95(6): 673–6.
338584	Cameroon	Kamgno J, BouchitÃ© B, Baldet T, Folefack G, Godin C, Boussinesq M. [Study of the distribution of human filariasis in West Province of Cameroon]. <i>Bull Soc Pathol Exot.</i> 1997; 90(5): 327–30.
332820	Cameroon, Central African Republic, Gabon	Ozoh G, Boussinesq M, Bissek A-CZ-K, Kobangue L, Kombila M, Mbina J-RM, Enyong P, Noma M, SÃ©kÃ©tÃ©li A, Fobi G. Evaluation of the diethylcarbamazine patch to evaluate onchocerciasis endemicity in Central Africa. <i>Trop Med Int Health.</i> 2007; 12(1): 123-9.

NID	Geographies	Citation
327872	Cameroon, Nigeria, Sudan, Uganda	Ozoh GA, Murdoch ME, Bissek A-C, Hagan M, Ogbuagu K, Shamad M, Braide EI, Boussinesq M, Noma MM, Murdoch IE, Sèkétéli A, Amazigo UV. The African Programme for Onchocerciasis Control: impact on onchocercal skin disease. <i>Trop Med Int Health</i> . 2011; 16(7): 875–83.
332855	Central African Republic	Kennedy MH, Bertocchi I, Hopkins AD, Meredith SE. The effect of 5 years of annual treatment with ivermectin (Mectizan) on the prevalence and morbidity of onchocerciasis in the village of Gami in the Central African Republic. <i>Ann Trop Med Parasitol</i> . 2002; 96(3): 297–307.
332796	Congo	Brito M, Paulo R, Van-Dunem P, Martins A, Unnasch TR, Novak RJ, Jacob B, Stanton MC, Molyneux DH, Kelly-Hope LA. Rapid Integrated Clinical Survey to Determine Prevalence and Co-distribution Patterns of Lymphatic Filariasis and Onchocerciasis in a Loa Loa Co-endemic Area: The Angolan Experience. <i>Parasite Epidemiol Control</i> . 2017; 2(3): 71–84.
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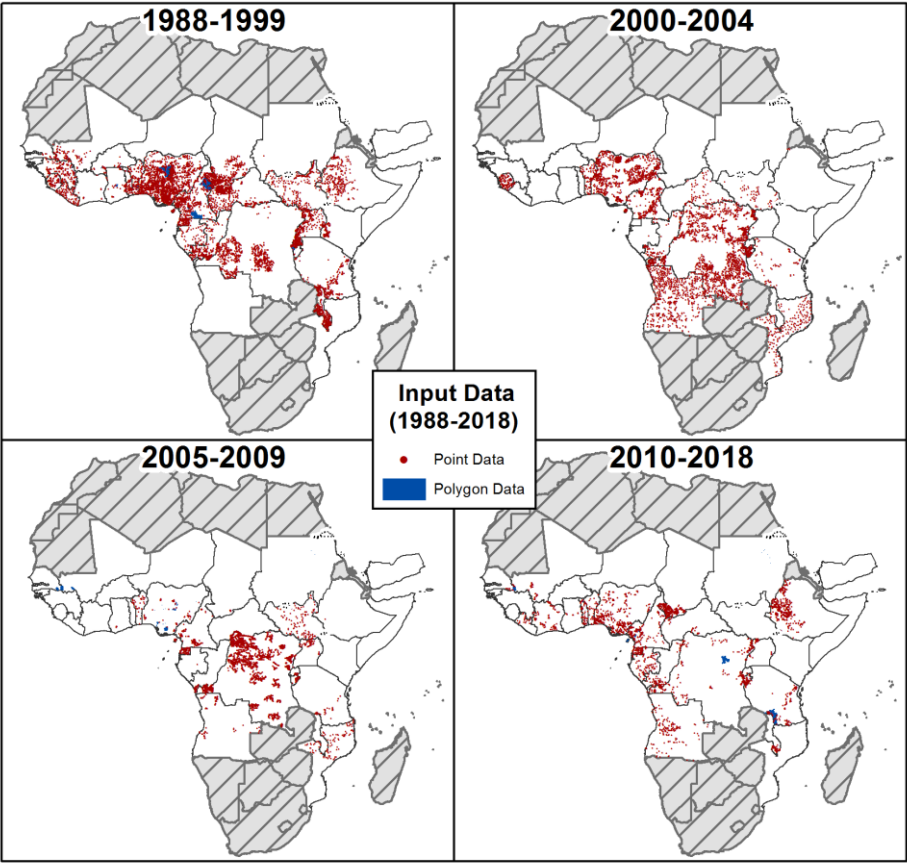
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Supplementary Figure 3. Africa and Yemen data coverage maps.

Locations of the unique point and polygon data used in modelling, grouped by years: 1988–1999, 2000–2004, 2005–2009, and 2010–2018. Hatched countries were not included in the MBG modelling region. Data that were able to be matched to point coordinates are represented by red dots and data that represented a larger area are represented by blue polygons.

3.5 APOC survey year matching

The African Programme for Onchocerciasis Control (APOC),(3) which operated from 1995 to 2015, relied on Rapid Epidemiological Mapping of Onchocerciasis (REMO) to map baseline onchocerciasis prevalence and endemicity in 19 onchocerciasis-endemic African countries outside the Onchocerciasis Control Programme (OCP) focal region (APOC was later expanded to additional countries). REMO(4,5) mapping, involving village-level surveys of nodule (onchocercoma) prevalence among adult males, was completed for 12 APOC countries by 2001.(6) We obtained these pre-2001 nodule mapping data, when available, from the ESPEN data portal, although detailed information on survey years were lacking for many (N = 4409) of these datapoints. So that these essential baseline mapping surveys could inform our spatiotemporal model of onchocerciasis prevalence, we investigated the pre-2001 REMO mapping surveys in APOC countries and attempted to narrow the possible range of survey years for these data. As widespread use of REMO by APOC began in 1996,(6) and these data were indicated by the ESPEN database as representing pre-2001 surveys, we considered 1996 and 2000 as the initial lower and upper bounds for survey years. We next identified survey locations for which the first year of onchocerciasis MDA (mass drug administration) occurred prior to 2001 (see section 4.2 below) and assumed that baseline mapping occurred at that location no later than the preceding year, treating that year as the new upper bound. Finally, we examined the available literature on pre-2001 APOC REMO surveys, by country, to further narrow the possible range of survey years. Our decisions around survey year estimates are summarised in Supplementary Table 4. For each datapoint, we specified the final survey year as the median of the lower and upper year bounds (rounding down), considering this approach to balance the potential conflicting biases from incorrect survey years on the temporal trends in the MBG model and the association with environmental and sociodemographic covariates in child models.

Survey year data were also missing for some Phase 1a and 1b monitoring and evaluation surveys in the APOC data retrieved from ESPEN. We identified survey years for most of these data by matching them to the locations, sample sizes, cases, and number of villages reported by Tekle et al.,(7) supplemented with additional sources(8,9) for some countries. We identified survey years for a total of 1001 observations. Survey years could not be reliably identified for some Phase 1a and 1b data, specifically Phase 1a data for Kasai Occidental (Democratic Republic of the Congo, DRC), Phase 1a and 1b data for Bioko Island (Equatorial Guinea), Phase 1a data for Kogi (Nigeria), and some Phase 1b data for Uganda; these data were excluded from our analysis. We assumed sample age ranges of 5–99 years of age for Phase 1a and 1b data(7) unless the ESPEN dataset indicated otherwise.

Supplementary Table 4. APOC REMO survey year investigations. CDTI: Community-directed treatment with ivermectin.

Country	Data Rows	Notes	Survey Year Strategy
Central African Republic	875	CDTI began in some areas in 1993, following REMO surveys.(10)	Use 1992 (one year prior to CDTI initiation) as lower bound.
Cameroon	389	Early proof-of-concept tests of the REMO methodology were performed in Cameroon in 1993.(4)	Assume 1993 as lower bound.
Democratic Republic of Congo	684	REMO began in 1997 in certain foci and continued until 2008;(11) CDTI areas were established in 1999 based on REMO.(12)	REMO appears to have been undertaken between 1997 and 1999 (although Bof(11) noted that it continued through 2008; ESPEN data include surveys beyond 2001). Assume 1997 as lower bound for pre-2001 DRC nodule surveys.
Republic of Congo	290	No information found except an MDA start year of 2001 for a small subset of these locations.	Assume 1996 as lower bound.
Ethiopia	283	Country-wide REMO was performed in Ethiopia in 1997 and 2001.(13)	Use 1997 as survey year for pre-2001 data.
Kenya	93	No information found.	Assume 1996 as lower bound for pre-2001 data.

Country	Data Rows	Notes	Survey Year Strategy
Nigeria	576	REMO was reported in 1995 to have been used for nationwide mapping;(3) another source indicated that REMO was carried out in Nigeria between 1994 and 1996.(14)	Assume 1994 and 1995 as lower and upper bounds of survey years for pre-2001 data.
Rwanda	90	Nationwide REMO was performed in 1999.(15)	Use 1999 as survey year for pre-2001 data.
Sudan	2	No information found.	Assume 1996 as lower bound.
South Sudan	110	REMO was performed between 1995 and 2002.(16)	Assume 1995 as lower bound.
Chad	481	REMO performed prior to 1999.(17)	Assume 1996 as lower bound, and 1998 as upper bound for surveys otherwise lacking upper bound estimates based on MDA.
Tanzania	142	Years of initiation of CDTI in Tanzania ranged from 1995 to 2004.(7)	Use 1994 as lower bound estimate.
Uganda	394	Nationwide REMO conducted between 1993 and 1997.(18)	Use 1993 and 1997 as lower and upper bounds, respectively.

4.0 Supplementary covariates

4.1 Pre-existing covariates considered for analysis

A variety of environmental and sociodemographic variables were used to predict all-age onchocerciasis prevalence. Where available, the finest spatiotemporal resolution of gridded datasets was used. Geospatial covariate rasters were resampled to ~ 5 km in GeoTiff format, for consistent modelling. Where data coverage was inconsistent between our standard mask and the input data, either a local average or nearest neighbour method (depending on data type) was used to fill spatial data gaps. Data from the nearest year available were used if covariate coverage did not include all model years. Supplementary Table 5 contains a full list of covariates considered in our analysis.

4.2 Creation of MDA covariate

Data used in the creation of the onchocerciasis MDA covariate were downloaded from the ESPEN Portal in April 2019.(19) The ESPEN data included the MDA start year and cumulative number of MDA rounds for each IU. ArcGIS software was then used to join the MDA data to the NTD (Neglected Tropical Disease) Implementation Unit Shapefile maintained by the Task Force for Global Health. A custom polygon was created for Sudan by referencing annual reports from The Carter Center. The MDA covariate value was set calculated as the cumulative number of rounds (defined as total number of rounds implemented) for the IU, and then converted to a raster for use in geospatial analysis. Data for lymphatic filariasis MDA were obtained from WHO(20) for the years during which LF (lymphatic filariasis) MDA was conducted, by IU, and joined onto the IU shapefile maintained by ESPEN. So that the effects of MDA for onchocerciasis and lymphatic filariasis were not double-counted in locations where both were indicated, a composite MDA covariate was produced which indicated the cumulative number of years for which either onchocerciasis or lymphatic filariasis MDA was indicated. This composite covariate was used in modelling.

4.3 Covariate reduction

High collinearity among covariates may lead to unstable model coefficients and unreliable predictions.³⁷ To reduce this problem, we derived a reduced covariate set using analysis of variance inflation factors (VIF).(22) Starting with the full list of covariates, we iteratively removed covariates with the highest VIF values until all remaining covariates had a VIF below 3.0. The reduced covariate set was used in fitting the MBG model and for spatiotemporal predictions. Supplementary Table 5 indicates the covariates that were retained in the final model, with representative plots provided for each covariate in Supplementary Figure 4. All variables were matched to the

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274 year of reported prevalence data, without any temporal lags (eg, temperature values for the year 2000 were joined to
275 prevalence data for the year 2000).

Supplementary Table 5. Covariates considered or retained for modelling, 1988–2018.

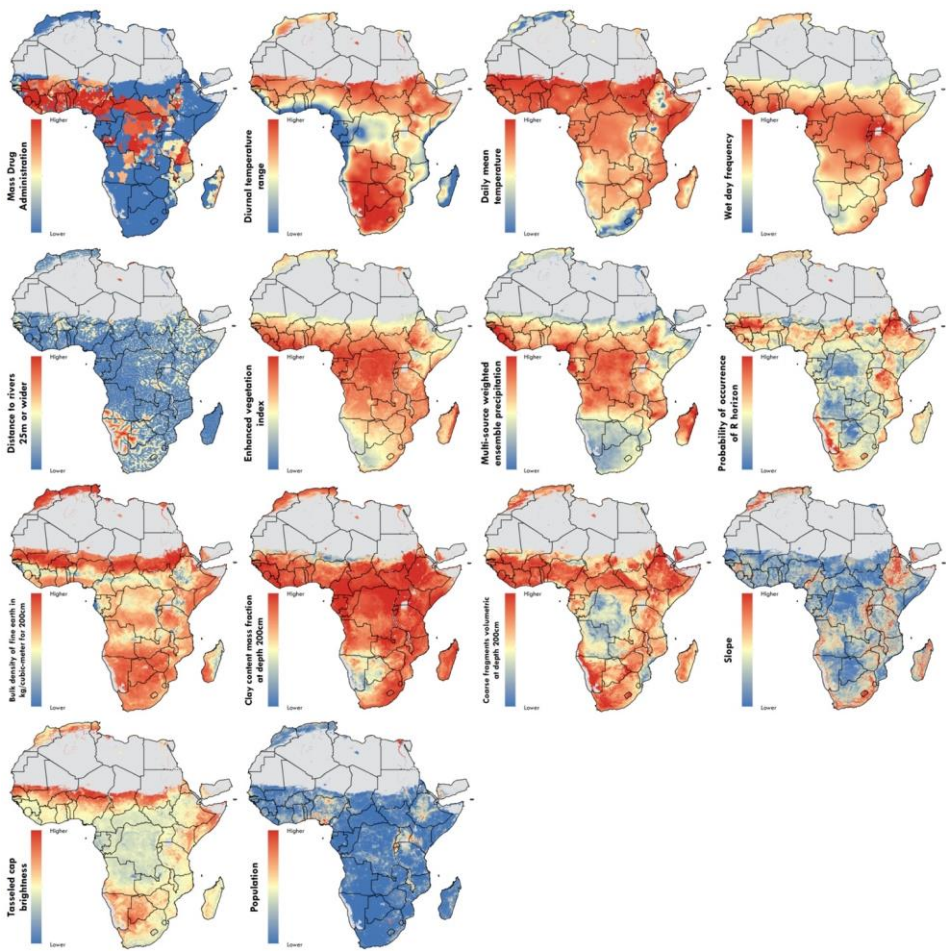
All covariates considered for inclusion as predictors of onchocerciasis prevalence. Short names are provided for reference in later figures and tables. Covariates retained after VIF analysis are indicated with a (+).

Covariate	Short Name	Retained in Final Model	Years Available	Source	Reference
Aridity	crutsard	-	1988–2018	Climatic Research Unit Time-Series (CRUTS)	Harris, I., Osborn, T.J., Jones, P. et al. Version 4 of the CRU TS monthly high-resolution gridded multivariate climate dataset. Sci Data 7, 109 (2020). https://doi.org/10.1038/s41597-020-0453-3 University of East Anglia. Climatic Research Unit TS v. 4.04 dataset. Available at: https://crudata.uea.ac.uk/cru/data/hrg/
Average daily mean temperature	crutstmp	+	1988–2018	CRUTS	Harris, I., Osborn, T.J., Jones, P. et al. Version 4 of the CRU TS monthly high-resolution gridded multivariate climate dataset. Sci Data 7, 109 (2020). https://doi.org/10.1038/s41597-020-0453-3 University of East Anglia. Climatic Research Unit TS v. 4.04 dataset. Available at: https://crudata.uea.ac.uk/cru/data/hrg/
Average daily minimum temperature	crutstmn	-	1988–2018	CRUTS	Harris, I., Osborn, T.J., Jones, P. et al. Version 4 of the CRU TS monthly high-resolution gridded multivariate climate dataset. Sci Data 7, 109 (2020). https://doi.org/10.1038/s41597-020-0453-3 University of East Anglia. Climatic Research Unit TS v. 4.04 dataset. Available at: https://crudata.uea.ac.uk/cru/data/hrg/
Average daily maximum temperature	crutstmx	-	1988–2018	CRUTS	Harris, I., Osborn, T.J., Jones, P. et al. Version 4 of the CRU TS monthly high-resolution gridded multivariate climate dataset. Sci Data 7, 109 (2020). https://doi.org/10.1038/s41597-020-0453-3 University of East Anglia. Climatic Research Unit TS v. 4.04 dataset. Available at: https://crudata.uea.ac.uk/cru/data/hrg/
Diurnal temperature range	crutsdtr	+	1988–2018	CRUTS	Harris, I., Osborn, T.J., Jones, P. et al. Version 4 of the CRU TS monthly high-resolution gridded multivariate climate dataset. Sci Data 7, 109 (2020). https://doi.org/10.1038/s41597-020-0453-3 University of East Anglia. Climatic Research Unit TS v. 4.04 dataset. Available at: https://crudata.uea.ac.uk/cru/data/hrg/

Covariate	Short Name	Retained in Final Model	Years Available	Source	Reference
Vapor pressure	crutsvap	-	1988–2018	CRUTS	Harris, I., Osborn, T.J., Jones, P. et al. Version 4 of the CRU TS monthly high-resolution gridded multivariate climate dataset. Sci Data 7, 109 (2020). https://doi.org/10.1038/s41597-020-0453-3 University of East Anglia. Climatic Research Unit TS v. 4.04 dataset. Available at: https://crudata.uea.ac.uk/cru/data/hrg/
Wet day frequency	crutswet	+	1988–2018	CRUTS	Harris, I., Osborn, T.J., Jones, P. et al. Version 4 of the CRU TS monthly high-resolution gridded multivariate climate dataset. Sci Data 7, 109 (2020). https://doi.org/10.1038/s41597-020-0453-3 University of East Anglia. Climatic Research Unit TS v. 4.04 dataset. Available at: https://crudata.uea.ac.uk/cru/data/hrg/
Distance to rivers >25 m wide	distrivers25m	+	Synoptic	Natural Earth Data (derived)	Andreadis KM, Schumann GJ-P, Pavelsky T. A simple global river bankfull width and depth database. Water Resources Research. 2013;49(10):7164–8.
River width (largest river within 5x5 km cell)	river_size	+	Synoptic	Natural Earth Data (derived)	Andreadis KM, Schumann GJ-P, Pavelsky T. A simple global river bankfull width and depth database. Water Resources Research. 2013;49(10):7164–8.
Elevation	elevation	-	Synoptic	NOAA/NCEI	Young, A. H., K. R. Knapp, A. Inamdar, W. B. Rossow, and W. Hankins, 2017: “The International Satellite Cloud Climatology Project, H-Series Climate Data Record Product”, Earth System Science Data, in preparation.
Enhanced Vegetation Index (EVI)	evi_v6	+	2000–2018	MODIS	Huete, A., Justice, C. & van Leeuwen, W. MODIS vegetation index (MOD 13) algorithm theoretical basis document. (1999). USGS & NASA. Vegetation indices 16-Day L3 global 500m MOD13A1 dataset. Available at: https://lpdaac.usgs.gov/dataset_discovery/modis/modis_products_table/mod13a1 . (Accessed: 25th July 2017) Weiss, D. J. et al. An effective approach for gapfilling continental scale remotely sensed timeseries. Isprs J. Photogramm. Remote Sens. 98, 106–118 (2014). C. Schaaf, Z. Wang. (2015). MCD43A1 MODIS/Terra+Aqua BRDF/Albedo Model Parameters Daily L3 Global - 500m V006. NASA EOSDIS Land Processes DAAC. http://doi.org/10.5067/MODIS/MCD43A1.006
Population	worldpop	+	1988–2018	WorldPop	Lloyd, C. T., Sorichetta, A. & Tatem, A. J. High resolution global gridded data for use in population studies. Sci. Data 4, sdata20171 (2017). World Pop. Get data. Available at: http://www.worldpop.org.uk/data/get_data/ . (Accessed: 25th July 2017)

Covariate	Short Name	Retained in Final Model	Years Available	Source	Reference
Growing season length	growingseason	-	Synoptic	FAO	FAO. GAEZ - Global Agro-Ecological Zones data portal. Available at: http://www.fao.org/nr/gaez/about-data-portal/en/ . (Accessed: 25th July 2017) FAO. GAEZ - Global Agro-Ecological Zones users guide. (2012).
Tassled cap brightness	tcb_v6	+	2000–2017	MODIS	USGS & NASA. Nadir BRDF- Adjusted Reflectance Reflectance 16-Day L3 Global 1km dataset. Available at: https://lpdaac.usgs.gov/dataset_discovery/modis/modis_products_table/mcd43b4 . (Accessed: 25th July 2017) Strahler, A. H. & Muller, J.-P. MODIS BRDF/Albedo product: algorithm theoretical basis document version 5.0. (1999). Weiss, D. J. et al. An effective approach for gapfilling continental scale remotely sensed timeseries. <i>Isprs J. Photogramm. Remote Sens.</i> 98, 106–118 (2014). C. Schaaf, Z. Wang. (2015). MCD43A1 MODIS/Terra+Aqua BRDF/Albedo Model Parameters Daily L3 Global - 500m V006. NASA EOSDIS Land Processes DAAC. http://doi.org/10.5067/MODIS/MCD43A1.006
Precipitation (Multi-source Weighted-Ensemble)	mswep	+	1988–2016	Princeton Climate Analytics	Beck, H.E., A.I.J.M. van Dijk, V. Levizzani, J. Schellekens, D.G. Miralles, B. Martens, A. de Roo: MSWEP: 3-hourly 0.25 global gridded precipitation (1979-2015) by merging gauge, satellite, and reanalysis data, <i>Hydrology and Earth System Sciences</i> , 21(1), 589-615, 2017.
Slope for land surfaces	slope	+	Synoptic	NOAA/NCEI	Young, A. H., K. R. Knapp, A. Inamdar, W. B. Rossow, and W. Hankins, 2017: “The International Satellite Cloud Climatology Project, H-Series Climate Data Record Product”, <i>Earth System Science Data</i> , in preparation.
Soil: Saturated water content (200 cm depth)	sgawcts	-	Synoptic	SoilGrid	Hengl T, Mendesde Jesus J, MacMillan RA, Batjes NH, Heuvelink GBM, Ribeiro E, Samuel-Rosa A, Kempen B, Leenaars JGB, Walsh MG, Gonzalez MR. “SoilGrids1km — Global soil information based on automated mapping.” <i>PLOS ONE</i> 9(8): e105992. 29 Aug 2014. doi:10.1371/journal.pone.0105992
Soil: Probability of bedrock exposure	sgbdrlog	+	Synoptic	SoilGrid	Hengl T, Mendesde Jesus J, MacMillan RA, Batjes NH, Heuvelink GBM, Ribeiro E, Samuel-Rosa A, Kempen B, Leenaars JGB, Walsh MG, Gonzalez MR. “SoilGrids1km — Global soil information based on automated mapping.” <i>PLOS ONE</i> 9(8): e105992. 29 Aug 2014. doi:10.1371/journal.pone.0105992
Soil: Bulk density (200 cm depth)	sgbldfie	+	Synoptic	SoilGrid	Hengl T, Mendesde Jesus J, MacMillan RA, Batjes NH, Heuvelink GBM, Ribeiro E, Samuel-Rosa A, Kempen B, Leenaars JGB, Walsh MG, Gonzalez MR. “SoilGrids1km — Global soil information based on automated mapping.” <i>PLOS ONE</i> 9(8): e105992. 29 Aug 2014. doi:10.1371/journal.pone.0105992
Soil: Clay content mass fraction (200 cm depth)	sgclyppt	+	Synoptic	SoilGrid	Hengl T, Mendesde Jesus J, MacMillan RA, Batjes NH, Heuvelink GBM, Ribeiro E, Samuel-Rosa A, Kempen B, Leenaars JGB, Walsh MG, Gonzalez MR. “SoilGrids1km — Global soil

Covariate	Short Name	Retained in Final Model	Years Available	Source	Reference
Soil: Coarse fragments (200 cm depth)	sgcrfvol	+	Synoptic	SoilGrid	information based on automated mapping.” <i>PLOS ONE</i> 9(8): e105992. 29 Aug 2014. doi:10.1371/journal.pone.0105992
Cumulative years of mass drug administration (MDA) for lymphatic filariasis or onchocerciasis	allmda	+	1988–2018	WHO (rasters produced in current study)	<p>Hengl T, Mendesde Jesus J, MacMillan RA, Batjes NH, Heuvelink GBM, Ribeiro E, Samuel-Rosa A, Kempen B, Leenaars JGB, Walsh MG, Gonzalez MR. “SoilGrids1km — Global soil information based on automated mapping.” <i>PLOS ONE</i> 9(8): e105992. 29 Aug 2014. doi:10.1371/journal.pone.0105992</p> <p>WHO. Expanded Special Project for Elimination of Neglected Tropical Diseases (ESPEN). https://www.espen.org/ (accessed May 20, 2020).</p> <p>WHO. Global Programme to Eliminate Lymphatic Filariasis. http://www.who.int/lymphatic_filariasis/elimination-programme/en/ (accessed May 29, 2019).</p>



Supplementary Figure 4. Africa and Yemen covariate values.

15 environmental or sociodemographic variables were used as inputs for the modelling process for Africa and Yemen (river size is not shown in this figure). Time-varying covariates are presented here for the year 2015. Please refer to Supplementary Table 5 for the corresponding citations for each covariate.

5.0 Supplementary methods

5.1 Age and diagnostic crosswalks

Surveys of onchocerciasis prevalence have varied in both their sampled age ranges and in the diagnostic tests they utilised. For example, epidemiological surveys during the OCP program relied on skin snip (microfiladermia) examinations of individuals aged 5 years or older,(23,24) while APOC relied on the REMO (Rapid Epidemiological Mapping of Onchocerciasis)(5,6,25,26) methodology to identify endemic areas requiring MDA, employing nodule palpation in adults. To harmonise these data sources, we adjusted survey data to represent all-age microfiladermia

prevalence by developing and applying an age and diagnostic crosswalk model that simultaneously accounts for typical age-dependent trends in prevalence and differential detection sensitivity between skin snip examination and nodule palpation. We did not crosswalk results from antibody (eg, Ov16) surveys as these data were not used in fitting the MBG model.

We identified published within-study comparisons reporting results from skin snip biopsies (microfiladermia), nodule (onchocercoma) palpation, or both, for more than one age group in a given study population. Study cohorts with zero reported cases were excluded from the crosswalk training dataset because they do not contribute information with which to derive age trends in endemic settings. Studies were eligible for inclusion in the crosswalk training dataset if they reported data from a country in our MBG modelling region; all eligible studies reported data from surveys conducted in our geospatial modelling timeframe (1988–2018). We identified 133 unique survey populations from 36 studies reporting skin snip prevalence in multiple age groups, and 126 unique survey populations from 22 studies reporting nodule prevalence in multiple age groups; among these surveys, a total of 100 unique survey populations from 19 studies reported both skin snip and nodule prevalence from the same study populations and age groups. Supplementary Table 6 summarises these sources.

Supplementary Table 6. Data used in estimation of age and diagnostic crosswalk.

Note that studies may have reported data for additional diagnostic tests that were not used in crosswalk model fitting.

Countries	Survey Years	Diagnostic Data
Benin(27)	1991	skin snip, nodule
Cameroon(28)	1992	skin snip
Cameroon(29)	1992	skin snip, nodule
Cameroon(30)	1996, 2005, 2006, 2011	skin snip, nodule
Cameroon(31)	2000	nodule
Cameroon(32)	2008	skin snip, nodule
Cameroon(33)	2009	skin snip, nodule
Cameroon(34)	2009	skin snip, nodule
Cameroon(35)	2012	skin snip, nodule
Cameroon(36)	2015	skin snip, nodule
Cameroon(37)	2015	skin snip, nodule
Cameroon(38)	2015	skin snip, nodule
Cameroon, Uganda(39)	1993 (Uganda), 1996 (Cameroon), 2005 (both countries)	skin snip, nodule
Central African Republic(40)	1990, 1995	skin snip
Ethiopia(41)	1994	skin snip, nodule
Ethiopia(42)	1997	skin snip

Ethiopia(43)	2005	skin snip
Ethiopia(44)	2006	skin snip
Ethiopia(45)	2012	skin snip
Gabon(46)	2004	nodule
Nigeria(47)	1992	skin snip
Nigeria(48)	1994	skin snip, nodule
Nigeria(49)	1994	skin snip
Nigeria(50)	1997	skin snip
Nigeria(51)	2000	skin snip, nodule
Nigeria(52)	2005	skin snip
Nigeria(53)	2007	skin snip
Nigeria(54)	2008	skin snip, nodule
Nigeria(55)	2008	skin snip
Nigeria(56)	2008	skin snip, nodule
Nigeria(57)	2009	skin snip
Nigeria(58)	2009	skin snip, nodule
Nigeria(59)	2010	nodule
Sierra Leone (60)	2010	skin snip
Togo(61)	1992	skin snip
Togo(62)	2014	skin snip
Togo(63)	2015	skin snip
Uganda(64)	1993	skin snip, nodule
Uganda(65)	2012	skin snip, nodule

We first retrieved population estimates by single age-year from the Global Burden of Disease (GBD)(66) for the country and year of each survey in the crosswalk training set. We assumed that the age distribution ($P(A)$, or probability of age A) within a survey sample matched that in the country and year of the survey, and estimated prevalence-by-age models ($P(D|A)$, or the probability of disease, D , at age A) from birth through age 94 years, the maximum individual age-year modelled by GBD. We used the GBD age distributions to estimate the likelihood of observing the reported number of cases in each surveyed age bin, given a logistic (binomial) regression model of the average prevalence-by-age relationship across surveys:

$$\text{logit}(P(D|A)) = \beta_0 + I_{ss}f_{ss}(A) + I_{nod}f_{nod}(A) + \alpha_i + I_{ss}\beta_{ss}$$

Onchocerciasis prevalence at a given age ($P(D|A)$) was modelled in logit space as a linear combination of an intercept, β_0 , which was set at $\text{logit}(0.00001)$ to drive prevalence at birth toward zero; basis splines (fda R package(67)) on age, $f(A)$, to accommodate non-linear age trends; α_i , cohort-level fixed effects (indicator variables

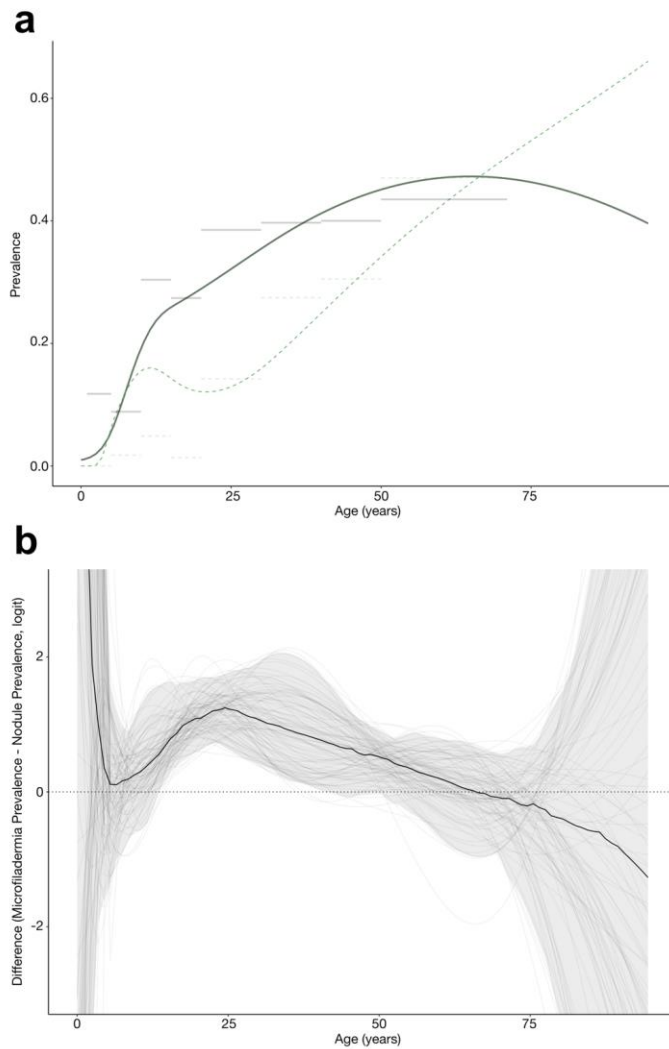
identifying study cohorts in the training dataset) to account for differences in study populations and survey designs; and a fixed diagnostic effect for skin snip surveys, β_{ss} , to model overall differences between skin snip and nodule surveys. Diagnostic indicator variables, I_{ss} and I_{nod} , were set to 1 for skin snip and nodule observations, respectively, and set to 0 otherwise, to select the appropriate age and diagnostic effects for a given survey. Study cohort effects were assumed to be shared across diagnostic tests. The likelihood function was evaluated as the sum, across studies (i) and age bins ($A1 \leq A < A2$), of the binomial likelihood of observing the reported numbers of cases given the reported sample sizes, age ranges, and diagnostic tests, using population-weighted mean prevalence values evaluated at the midpoints of each age year.

Spline knot placements were identified by spacing four internal knots evenly by quantile in the training data, determined separately for skin snip and nodule surveys, with additional knots placed at ages 3, 6, and 65 to help stabilise model behaviour in early childhood and older adulthood. Starting values for all spline, cohort and diagnostic coefficients were randomly drawn from uniform distributions in the interval $[-5, 5]$. Models were estimated with maximum likelihood optimisation using the box constraint quasi-Newton algorithm⁶⁶ (without constraints) in the optim function (R stats package(69)).

Prior to running the MBG model, all survey data that derived from a restricted age range (ie, anything other than 0–94 years), or that represented nodule prevalence, were adjusted using the fitted crosswalk model. For each individual survey population, the crosswalk model was refit via maximum likelihood to the prevalence data for that survey, with all coefficients other than the cohort-level effect fixed, α_i , fixed to their estimates from the full crosswalk model. This newly estimated value of α_i was then re-inserted into the crosswalk model with the other coefficients fixed, to calculate all-age skin snip prevalence for that survey sample. Crosswalked estimates were then used as the outcome measure in the MBG model.

Surveys that reported prevalence of 0% or 100% were particularly difficult to interpret for crosswalk purposes. A report of 0% prevalence could reflect true absence of infection, insufficient diagnostic sensitivity, or sampling variance in small samples, while a report of 100% prevalence masks any implicit linkage between prevalence and the scale of infection intensity. While many of these concerns also apply to surveys reporting intermediate prevalence, the particular uncertainty involved with surveys reporting 0% or 100% prevalence increases the risk of inappropriate crosswalking, and for this reason we did not crosswalk such surveys.

Uncertainty in the crosswalk models was estimated using bootstrap analysis, with 100 replicates generated by sampling, with replacement, an equal number of study cohorts as in the full crosswalk training dataset. Resampling was conducted by cohort rather than age bin to better estimate variation among surveys. Spline knot locations and model coefficients were estimated for each bootstrap replicate independently of other replicates and of the full model. To visualise bootstrapped results for a given survey, cohort effects were calculated for each replicate using maximum likelihood optimisation with all other coefficients fixed to the estimated values for the replicate.



Supplementary Figure 5. Diagnostic and age crosswalk model.

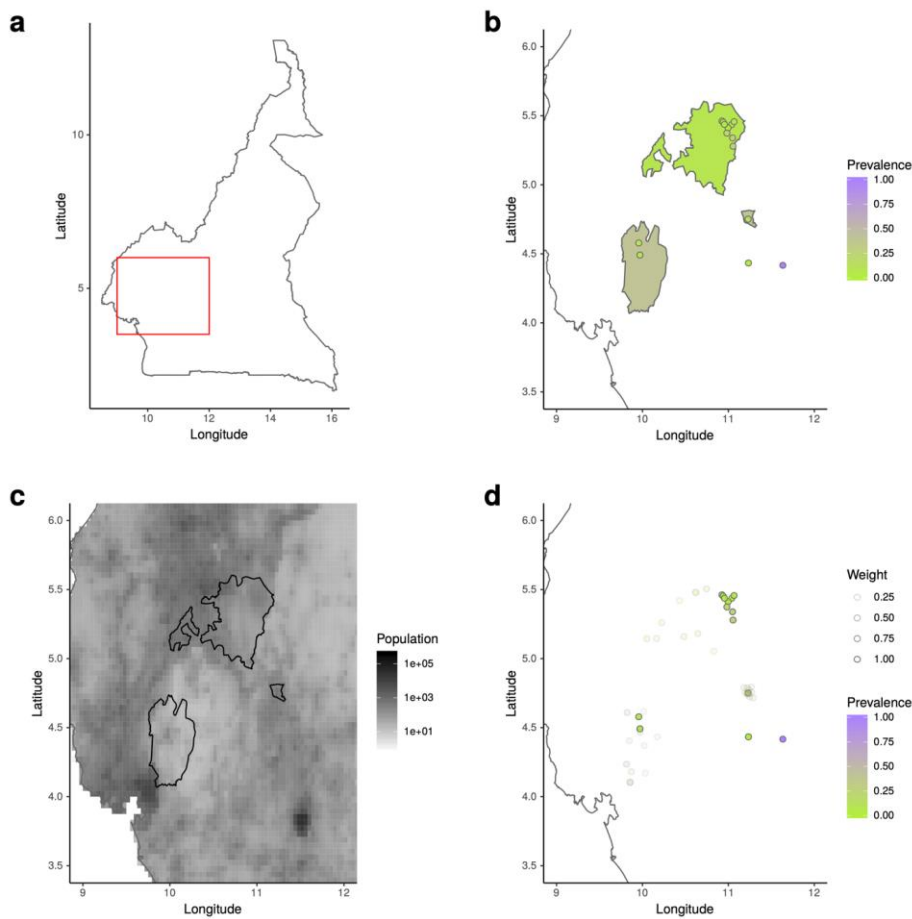
Outputs from the diagnostic and age crosswalk model. (a) Example prevalence-by-age curves fit to a study from Benin(27) that reported both skin snip and nodule age-binned data and was included in the crosswalk model training dataset. Estimates are shown for skin snip microfiladermia prevalence (solid black line) and nodule prevalence (dotted green line); horizontal bars indicate the age-binned prevalence data reported by that study (solid gray: skin snip; dotted green: nodule). (b) Bootstrapped estimates of differences between microfiladermia (skin snip) and nodule prevalence by age, in logit space. Each of 100 bootstrap samples is shown (faint grey line) along with the median estimate (solid black line) and central 95% uncertainty interval (shaded area). Positive values indicate higher skin snip prevalence than nodule prevalence.

In an example location with moderately high microfiladermia prevalence, our final crosswalk model (Supplementary Figure 5) suggests a sharp increase in prevalence through childhood and adolescence, with a slower increase between approximately ages 15 and 65 years. This general pattern is qualitatively consistent with the range of onchocerciasis microfiladermia age trends reported by other studies,(70–72) although substantial site-specific variation exists. The relationship between microfiladermia and nodule prevalence within onchocerciasis-endemic communities was previously modelled by Coffeng et al.,(73) using pre-control data from a broad sampling of sites in OCP and APOC regions. Their multivariate logistic regression model, relating nodule prevalence in adult males (aged 20 years or older) to microfiladermia prevalence in individuals aged 5 years or older, did not explicitly model changes in prevalence with age, but revealed higher prevalence estimates from skin snip biopsies than those from nodule palpation surveys in the same communities. Our full crosswalk model and bootstrapped uncertainty estimates (Supplementary Figure 5) are qualitatively consistent with their results, estimating that skin snip prevalence exceeds nodule prevalence, on average, from adolescence until about age 50, with poorer resolution of this relationship in children and older adults.

While crosswalking survey data that are based on differing diagnostics and age coverage enabled us to leverage a more comprehensive geospatial dataset than is otherwise tractable, we identify several limitations to our crosswalk approach. Our model assumes that, for a given diagnostic approach, changes in prevalence by age follow a consistent pattern across locations, years, sexes, and programmatic contexts. However, actual and reported microfiladermia and nodule prevalence patterns are influenced by local factors including ecological conditions; infection intensity; vector identity, density, exposure, and control history; MDA coverage; variations in survey sampling designs and in diagnostic specificity and sensitivity; and sex.(70,72–79) We have not modelled the age- and diagnostic-specific effects of MDA on prevalence in the crosswalk models (although MDA is included as a predictor during the MBG modelling stage). We were also limited by the variable reporting of age information among data sources. Some surveys did not report age ranges, and for these surveys we assumed that their data represented all-age prevalence, risking possible misclassification. The absence of individual-level data on onchocerciasis prevalence also precluded full age-standardisation, as we could only assume that the survey sample matched the age structure of the general population. Our crosswalks do not currently account for the sensitivity and specificity of skin snip and nodule diagnostics, and crosswalk uncertainty is therefore likely to be underestimated. We also do not currently have a computationally feasible method to propagate uncertainty from the crosswalk models into the MBG models in a way that accounts for the inconsistent reporting of sampled age ranges among studies.

5.2 Polygon resampling

Prevalence records are representative of either georeferenced point locations or polygonal areas (eg, as defined by the borders of administrative or programmatic units). As our modelling framework relies on coordinate-referenced data in order to fit the continuous spatial random fields, we converted areal data to a representative collection of point data. This “polygon resampling” process, described previously for geospatial modelling of under-5 mortality,(80) generates candidate locations based on the underlying population density of the resampled area, implicitly assuming that surveys employed population-based designs, and is illustrated in Supplementary Figure 6 using Cameroon in 2015 as an example. For each polygonal observation in our dataset, 10 000 points were randomly sampled from within the polygon, with weighting by the WorldPop total population raster. Candidate points were clustered using k-means clustering, generating a set of final points with a density of 1 per 1000 grid cells, except for small polygons, in which case density was iteratively increased by a factor of 10 until a minimum threshold of 10 points was achieved. Weights were assigned to each point proportionally to the number of candidate points that entered into the k-means cluster. The points generated by this resampling process were assigned the prevalence of onchocerciasis reported for the survey for that polygon. These sample weights were used in MBG model fitting within INLA.



Supplementary Figure 6. Polygon resampling.

The process of polygon resampling is illustrated using reported onchocerciasis prevalence data for Cameroon in 2015. (a) National map of Cameroon, showing the inset area (red frame) that is featured in the remaining panels. (b) Reported areal (irregular polygons) and point-level (circles) prevalence data. (c) Underlying population surface from WorldPop (displayed on a log₁₀ scale), with survey polygons overlaid. Polygons are resampled proportionally to 5 x 5 km population density. (d) Locations of the final datapoints for geospatial modelling, showing both those data originally reported with coordinates and those derived by resampling the polygon data. The opacity of resampled datapoints in (d) reflects their relative weights, which sum to 1.0 within an individual areal survey.

5.3 Geostatistical model

5.3.1 Model geographies and time period

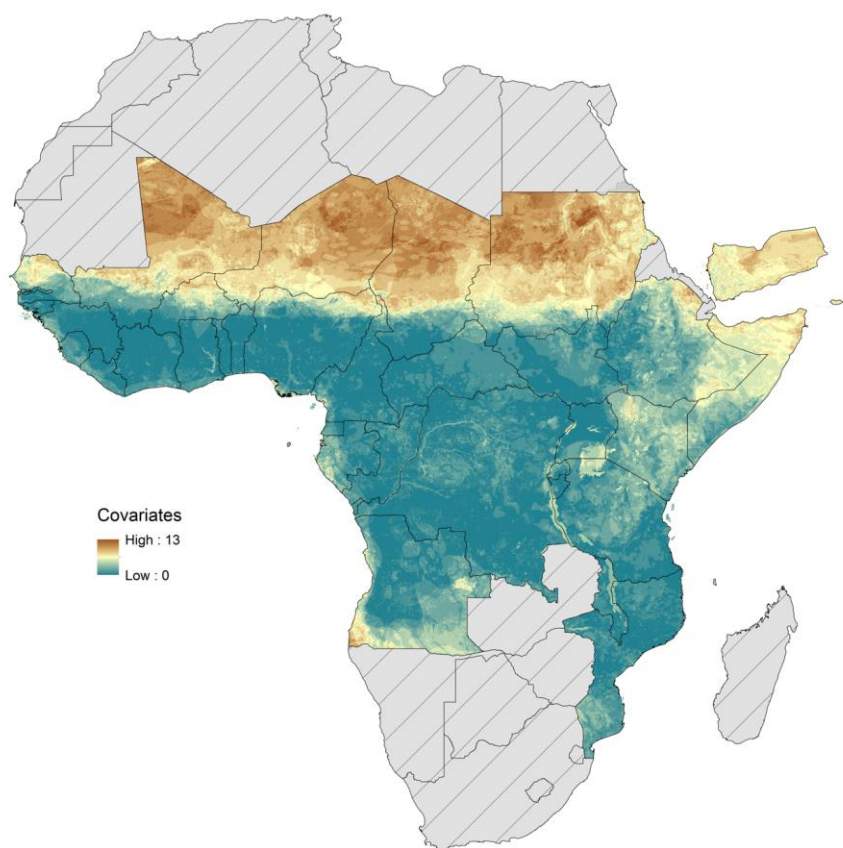
Model-based geostatistical (MBG) methods were used to generate estimates of all-age onchocerciasis microfiladermia prevalence for onchocerciasis-endemic countries of Africa, plus Yemen (listed in Supplementary Table 2). A single model was fit to this geographical region. We did not model countries in the Western Hemisphere that were formerly or are currently endemic for onchocerciasis, due to their highly localised onchocerciasis foci. We were principally concerned with estimates for the time period 2000–2018, but fit the model using data from 1988–2018 in order to leverage information from pre-2000 OCP and APOC surveys, and thereby improve “baseline” (ie, year 2000) model estimates in countries covered by those programmes. Reporting of results in the main text focuses on estimates for 2000–2018.

5.3.2 Covariate coverage

As with any regression model, the reliability of predictions from our model is affected by the overlap between covariate values in training and prediction datasets. Predictions in regions with a range of covariate values that fall outside the range of values in the training set may be prone to extrapolation errors. Supplementary Figure 7 illustrates the number of covariates whose mean values (averaged over 1988–2018) fall outside the central 95% quantile interval of values in the training set, for each 5 x 5-km pixel. Child model predictions in areas with poor covariate representation (ie, a large number of covariates with values outside the central interval) should be considered with special caution. These areas include the Sahel and Sahara, as well as Yemen, Kenya, Somalia, eastern Ethiopia, and southern Angola.

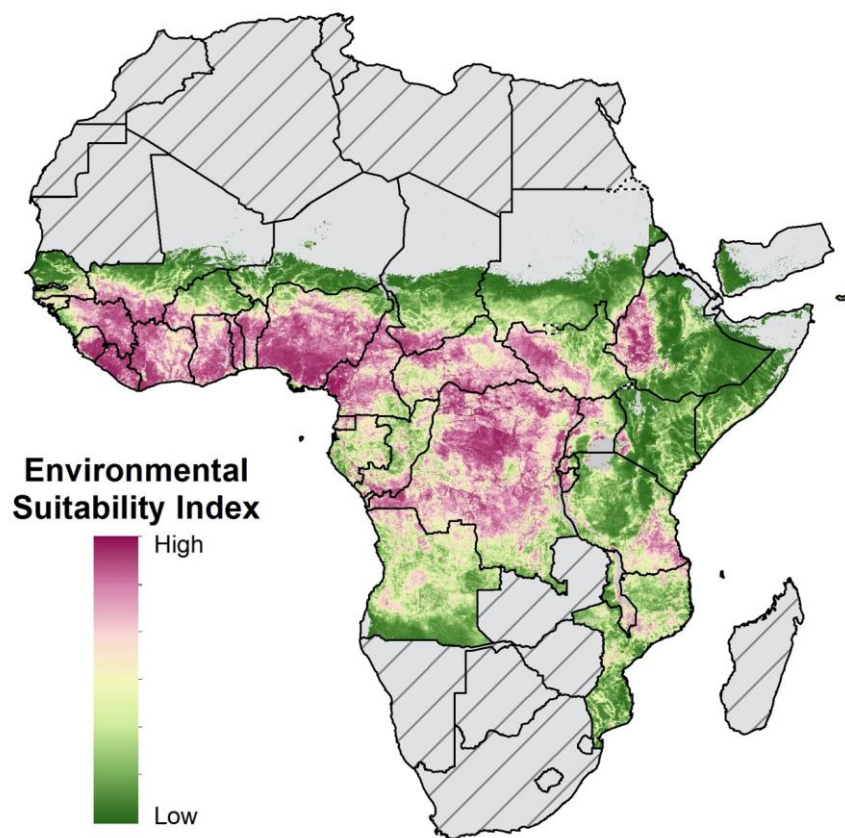
5.3.3 Environmental suitability

Cromwell et al. recently produced an environmental suitability model for onchocerciasis, using a boosted regression tree (BRT) model trained with space- and time-referenced data on onchocerciasis occurrence. (81) In contrast to our present MBG model, the Cromwell suitability model leveraged data from not only skin snip and nodule palpation surveys, but also Ov16 serosurveys, onchocerciasis-derived eye or skin morbidity surveys, and other diagnostics. The model was trained using an overlapping but non-identical set of environmental covariates as those used in the present study, including climatic, topographic, hydrologic, vegetative, and urbanicity variables. Pseudo-absence records were generated using background sampling. Importantly for our present purposes, this suitability model does not incorporate the influence of programmatic interventions for onchocerciasis and does not use data on onchocerciasis absence, providing a reflection of *O. volvulus* endemicity apart from temporal changes in infection prevalence. The outputs of the BRT model represent a spatially explicit index from 0–1, reflecting environmental suitability for onchocerciasis occurrence in a given 5 x 5-km cell. We incorporated mean predictions from this suitability model, using covariate values for 2016 (Supplementary Figure 8), to improve the behaviour of our MBG model in areas of poor prevalence data coverage, particularly in areas where covariates lie outside the range of values in sampled locations and MBG predictions are therefore subject to extrapolation uncertainty.



Supplementary Figure 7. Number of covariates with values outside the central 95% interval of values at survey sites.

For each 5 x 5-km pixel, the number of covariates (out of a total of 16) whose average value lies outside the central 95% interval of values at observed survey sites is indicated. Localities with a large number of such covariates represent areas in which predictions are less reliable, due to potential extrapolation error.



Supplementary Figure 8. Onchocerciasis suitability model predictions (Cromwell et al., 2021) for 2016.

Mean environmental suitability for onchocerciasis, from a previous boosted regression tree (BRT) analysis (81) of onchocerciasis presence data, using covariate values for 2016.

5.3.4 Model description

We modelled *Onchocerca volvulus* infection prevalence using a spatially explicit Bayesian generalised linear mixed effects regression model in R-INLA:

$$Y_{i,t} \sim \text{Binomial}(p_{i,t}, N_{i,t})$$

$$\text{logit}(p_{i,t}) = \beta_0 + \beta X_{i,t} + \gamma_{c[i]} + U_i + Z_i + \epsilon_{i,t}$$

$$Z_i \sim \text{GP}(0, \Sigma_{\text{space}})$$

$$\epsilon_{i,t} \sim N(0, \sigma_{\text{nug}}^2)$$

The number of infected individuals ($Y_{i,t}$) among a sample ($N_{i,t}$) in location i and year t was modelled as a binomial variable. This model specifies logit-transformed infection prevalence ($p_{i,t}$) as a linear combination of an intercept for the modelling region (β_0); covariate fixed effects (coefficients, β , and values, $X_{i,t}$); country random effects ($\gamma_{c[i]}$); a second-order random walk model on estimates of onchocerciasis environmental suitability (U_i); a spatially correlated random field (Z_i); and an uncorrelated error term or nugget effect ($\epsilon_{i,t}$). The spatial random field (Z_i) was modelled as a Gaussian process with mean 0 and a Matérn covariance function (Σ_{space}). Modelled outputs from the BRT analysis of onchocerciasis environmental suitability (U_i), described in section 5.3.3, were modelled with a second-order random walk, with values grouped into 25 bins by quantile. This random walk model accommodates non-linearity in the relationship between suitability (based on presence-absence data) and the linear predictor. The INLA (integrated nested Laplace approximation) model was fit using an “empirical Bayes” integration strategy, which relies on mode values to approximate hyperparameter posterior distributions during estimation and enabled us to achieve tractable model computation times.(82)

5.3.5 Priors

We specified minimally informative priors for INLA hyperparameters, as detailed in Supplementary Table 7. Priors for the spatial hyperparameters τ and κ were derived automatically by R-INLA based on the finite elements mesh.

Supplementary Table 7. INLA model priors.

Parameter	Description	Prior
β_0, β	Intercept, covariate fixed effects	$N(\mu = 0, \sigma^2 = 3^2)$
$\left(\frac{1}{\sigma_{country}^2}\right)$	Precision for country random effects ($\gamma_{c[i]}$, i.i.d.)	$\text{gamma}(a = 1, b = 0.00005)$
$\left(\frac{1}{\sigma_{nug}^2}\right)$	Precision for nugget effect ($\epsilon_{i,t}$, i.i.d)	$\text{pc.prec}(u = 0.5, \alpha = 0.05)^1$
$\left(\frac{1}{\sigma_U^2}\right)$	Precision for suitability model (U_i , RW2 model)	$\text{gamma}(a = 1, b = 0.00005)$
σ_{SPDE}	Standard error for SPDE model	$\text{pc}(3.0, 0.05)^2$
Range	Range for SPDE model	$\text{pc}(0.06171913, 0.05)^3$

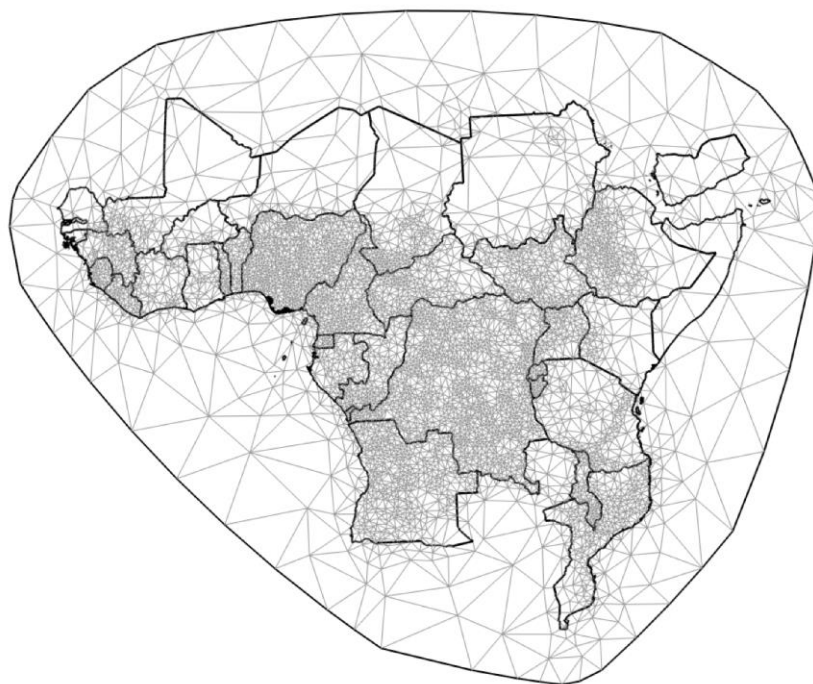
¹PC prior for precision.

²PC prior for SPDE σ , indicating a 5% probability that σ is greater than 3.

³PC prior for SPDE range, indicating a 5% probability that range is less than 0.06161412 (5% of the maximum extent of the spatial mesh).

5.3.6 Mesh construction

We modelled continuous spatial random effects using stochastic partial differential equations (SPDE) representations of Gaussian-Markov random field (GMRF) approximations of a spatially autocorrelated Gaussian process, using triangular finite element meshes as implemented in the R-INLA R package.(83–85) Due to the large geographical size of the model region, a spherical (S2) mesh was constructed in order to minimise distance distortions. Minimum and maximum edge lengths were set to 25 and 500 km, respectively, and a 1000-km external buffer was used to avoid artifacts at the edges of the spatial domain.(84) These values were chosen to yield denser mesh vertices in data-rich areas while maximising the overall spatial field resolution, within computational constraints. The spatial mesh is illustrated in Supplementary Figure 9.



Supplementary Figure 9. Spatial mesh construction.

Two-dimensional projection of spherical refined Delaunay triangulation mesh used in estimating spatial random fields in Africa and Yemen, with national boundaries (bold lines). The mesh features greater vertex density in data-rich locations.

5.3.7 Model fitting and estimation generation

Models were fit using the integrated nested Laplace approximation (INLA) algorithm in R-INLA. Fitted models for each region were used to generate 1000 random samples from the joint posterior distributions of model parameters, yielding mean and uncertainty estimates for onchocerciasis prevalence.

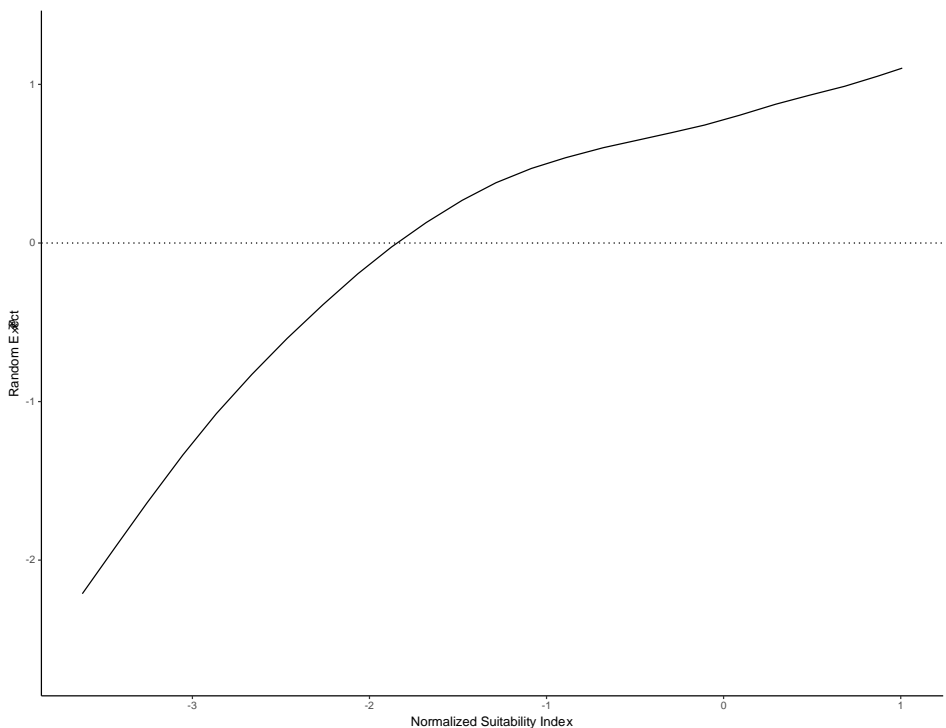
5.3.8 Model results

Model parameter estimates from the MBG model are summarised in Supplementary Table 8. Nominal range is the distance (in km) at which spatial correlation has declined to about 0.1. Estimated random effects for the environmental suitability covariate are displayed in Supplementary Figure 10.

Supplementary Table 8. Parameter estimates from in-sample onchocerciasis MBG model.

GP: Gaussian process. i.i.d.: Independent and identically distributed random effects. RW2: Second-order random walk model.

Parameter	Median (95% UI)
Intercept	-3.656 (-4.214, -3.099)
GP nominal range	189.6 (176.7, 207.1)
GP nominal variance	3.177 (2.892, 3.568)
Nugget precision (i.i.d.)	0.719 (0.683, 0.742)
Country precision (i.i.d.)	0.505 (0.268, 0.992)
Environmental suitability precision (RW2)	32.269 (5.731, 125.009)
Covariate coefficients	
MDA (allmda)	-1.106 (-1.136, -1.077)
Diurnal temperature range (crutsdr)	-0.363 (-0.537, -0.199)
Temperature (crutstmp)	-0.172 (-0.325, -0.018)
Wet day frequency (crutswet)	-0.163 (-0.315, -0.012)
Distance to rivers >25 m wide (distrivers25m)	-0.021 (-0.070, 0.028)
Enhanced Vegetation Index (evi_v6)	0.192 (0.129, 0.256)
Precipitation (mswep)	0.137 (0.057, 0.217)
Soil: Probability of bedrock exposure (sgbdrlog)	-0.063 (-0.119, -0.007)
Soil: Bulk density (200 cm depth) (sgbldfie)	0.100 (0.038, 0.163)
Soil: Clay content mass fraction (200 cm depth) (sgclyppt)	-0.078 (-0.157, 0.002)
Soil: Coarse fragments (200 cm depth) (sgcrfvol)	0.077 (0.008, 0.146)
Slope (slope)	0.008 (-0.021, 0.038)
Tasseled cap brightness (tcb_v6)	-0.333 (-0.404, -0.262)
Population (worldpop)	-0.014 (-0.037, 0.010)
River width (river_size)	0.072 (0.039, 0.106)



Supplementary Figure 10. Estimated random effects for environmental suitability.

Random effect estimates (mean and 95% UI) are shown for the environmental suitability layer, fit within INLA using a second-order random walk (RW2) model with standard-normalised suitability values (x-axis) grouped into 25 bins by quantile; y-axis values indicate effects in logit space.

5.3.9 *Loa loa* endemicity

The co-occurrence of *O. volvulus* and the filarial nematode *Loa loa* complicates onchocerciasis control in some parts of central Africa,(86,87) due to potentially severe complications from ivermectin treatment in individuals with high *L. loa* microfilariae loads.(88,89) Previous modelling studies have estimated the scale of co-infections between *O. volvulus* and *L. loa* and have suggested a substantial burden of co-infection or populations at risk.(90,91) We complemented these studies by calculating the mean number of onchocerciasis cases estimated by our geospatial model in 2018, in areas considered meso- or hyper-endemic for loiasis according to endemicity classifications from ESPEN. We obtained data from ESPEN for loiasis endemicity in 2015 at the level of Implementation Units and combined this with the WorldPop population raster and our onchocerciasis mean prevalence estimates, deriving a mean estimate of 7 146 618 onchocerciasis cases in loiasis-endemic areas (excluding hypo-endemic regions) in 2018.

5.4 Model validation

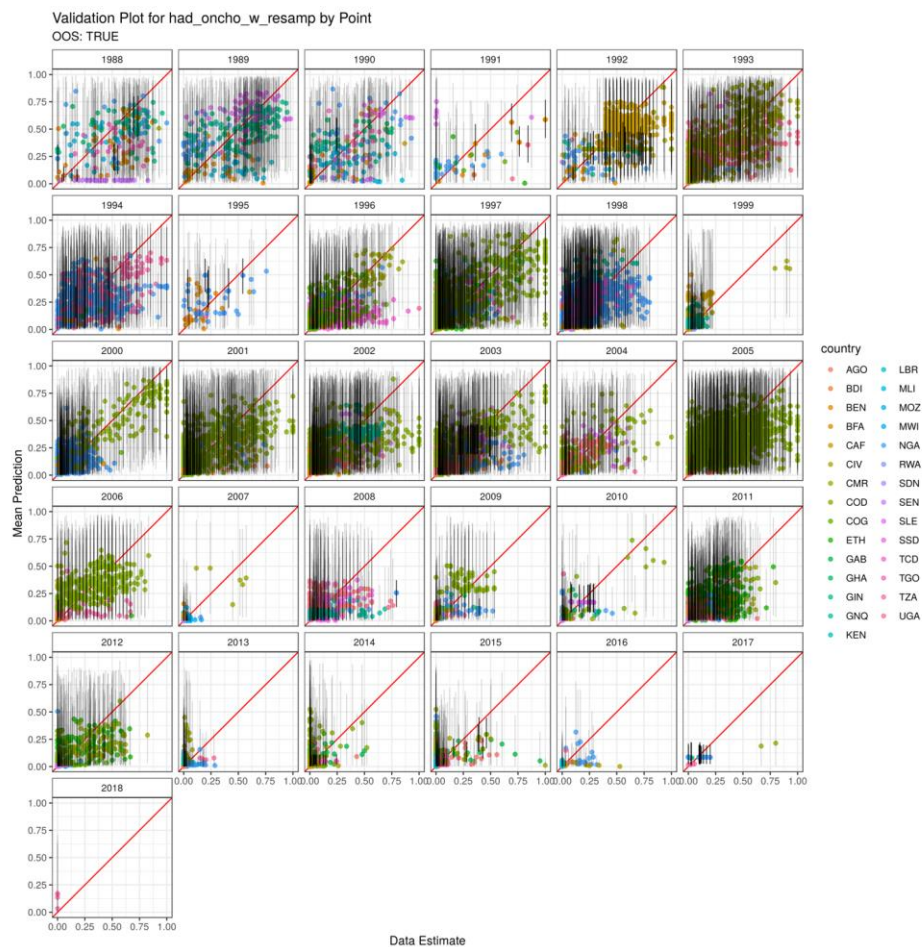
5.4.1 Metrics of predictive validity

In order to assess the predictive validity of our estimates, we validated our models using spatially stratified five-fold out-of-sample cross-validation. To construct each spatial fold, we used a modified bi-tree algorithm to spatially aggregate datapoints. This algorithm recursively partitions two-dimensional space, alternating between horizontal

and vertical splits on the weighted data sample size medians, until the data contained within each spatial partition are of a similar sample size. The depth of recursive partitioning is constrained by the target sample size within a partition and the minimum number of clusters or pseudo-clusters allowed within each spatial partition (in this case, a minimum sample size of 500 was used). These spatial partitions are then allocated to one of five folds for cross-validation. Temporal partitioning was unstructured (random).

For validation, each geostatistical model was run five times, each time holding out data from one of the folds. A set of out-of-sample predictions were generated by sampling from the posterior predictive distribution for each held-out datapoint. A full suite of out-of-sample predictions over the entire dataset was calculated by combining the out-of-sample predictions from the five cross-validation runs. Using these out-of-sample predictions, we computed mean error (bias), mean absolute error, 95% coverage of the predictive intervals (the proportion of observed out-of-sample data that fall within the predicted 95% uncertainty intervals), root-mean squared error (RMSE, which summarises error variance), and the correlation of predicted versus observed prevalence at the level of individual datapoints. A scatterplot of reported prevalence versus mean out-of-sample predictions is provided in Supplementary Figure 11, and validation metrics are summarised in Supplementary Table 9.

In addition to performing cross-validation, we also evaluated over-dispersion in the model input data by performing posterior predictive checks. Briefly, a binomial count was simulated from each of 1,000 model draws of predictions for each input data row. The distribution of predictions provided a good approximation of the observed distribution, suggesting that the model is adequately addressing possible over-dispersion in the data.



Supplementary Figure 11. Model validation scatterplots for Africa and Yemen.

Reported prevalence versus mean out-of-sample predictions for individual datapoints, by year and country. Vertical bars represent 95% UI; red lines indicate equivalence.

Supplementary Table 9. Out-of-sample validation metrics at the level of individual datapoints, from five-fold cross-validation

Out-of-sample performance was aggregated over 1988–2018 and is also provided for individual model years. Values were computed in prevalence space. N observations: Number of data rows in the full dataset from a given year. Mean abs. error: Mean absolute error. 95% cov.: 95% coverage. Corr.: Correlation.

Year	N Observations	Mean error	Mean abs. error	95% cov.	RMSE	Corr.
1988	216	0.158	0.257	0.837	0.298	0.298
1989	310	0.017	0.155	0.934	0.205	0.645
1990	232	-0.081	0.171	0.919	0.209	0.565
1991	58	0.159	0.317	0.639	0.407	-0.003
1992	983	0.040	0.130	0.961	0.174	0.658
1993	866	-0.005	0.175	0.920	0.233	0.554
1994	770	0.060	0.159	0.941	0.210	0.359
1995	58	-0.037	0.124	0.908	0.161	0.500
1996	1 104	0.021	0.093	0.967	0.144	0.709
1997	1 148	-0.008	0.138	0.959	0.189	0.745
1998	1 620	-0.023	0.109	0.970	0.152	0.615
1999	370	-0.052	0.074	0.942	0.126	0.575
2000	958	-0.026	0.089	0.978	0.123	0.814
2001	1 163	-0.010	0.112	0.954	0.163	0.707
2002	885	0.012	0.147	0.913	0.200	0.547
2003	1 082	0.017	0.119	0.953	0.173	0.615
2004	397	0.009	0.113	0.930	0.153	0.658
2005	1 930	0.005	0.144	0.965	0.188	0.533
2006	383	-0.002	0.104	0.727	0.170	0.611
2007	117	-0.017	0.046	0.899	0.083	0.631
2008	367	0.046	0.112	0.902	0.195	0.666
2009	193	-0.010	0.055	0.816	0.092	0.566
2010	133	0.020	0.066	0.904	0.097	0.722
2011	764	0.009	0.097	0.903	0.140	0.620

Year	N Observations	Mean error	Mean abs. error	95% cov.	RMSE	Corr.
2012	543	0.014	0.063	0.934	0.115	0.659
2013	477	-0.014	0.021	0.965	0.046	0.207
2014	397	-0.035	0.051	0.914	0.096	0.344
2015	472	-0.043	0.077	0.853	0.129	0.182
2016	96	0.004	0.033	0.955	0.075	0.138
2017	20	0.062	0.096	0.995	0.189	0.881
2018	4	-0.138	0.138	0.524	0.145	NA
1988–2018	18 116	0.003	0.111	0.924	0.168	0.706

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