

**Elucidating the Mechanistic Origin of a Spin-State
Dependent Fe_{Nx}-C Catalyst toward Organic Contaminant
Oxidation via Peroxymonosulfate Activation**

ZHANG, Bofan, LI, Xianquang, AKIYAMA, Kazuhiko, BINGHAM, Paul
<<http://orcid.org/0000-0001-6017-0798>> and KUBUKI, Shiro

Available from Sheffield Hallam University Research Archive (SHURA) at:

<https://shura.shu.ac.uk/29459/>

This document is the Accepted Version [AM]

Citation:

ZHANG, Bofan, LI, Xianquang, AKIYAMA, Kazuhiko, BINGHAM, Paul and KUBUKI, Shiro (2021). Elucidating the Mechanistic Origin of a Spin-State Dependent Fe_{Nx}-C Catalyst toward Organic Contaminant Oxidation via Peroxymonosulfate Activation. Environmental Science and Technology. [Article]

Copyright and re-use policy

See <http://shura.shu.ac.uk/information.html>

1 **Elucidating Mechanistic Origin of Spin-State Dependent FeN_x-C**

2 **Catalyst towards Organic Contaminants Oxidation via**

3 **Peroxymonosulfate Activation**

4
5 Bofan Zhang^{1,†,*}, Xianquan Li^{2,†}, Kazuhiko Akiyama¹, Paul A. Bingham³, Shiro

6 Kubuki¹

7 ¹Department of Chemistry, Tokyo Metropolitan University, Tokyo 192-0397, Japan

8 ²Dalian Institute of Chemical Physics, Chinese Academy of Sciences, Dalian 116023,

9 China

10 ³College of Business, Technology and Engineering, Sheffield Hallam University,

11 Howard Street, Sheffield S1 1WB, UK

12
13
14
15
16
17
18
19 *Corresponding author: Bofan Zhang

20 E-mail: 15054218031@163.com (B.F. Zhang)

21 [†] These authors contributed equally

22 **Abstract**

23 Atomically dispersed metal on nitrogen-doped carbon matrices have delivered
24 extensive interest in the removal of refractory organic pollutants. However, thorough
25 exploration of the particular structure for each active site and specific effects of these
26 sites still remains elusive. Herein, a Fe-pyridinic N₄ structure in a single-atom catalyst
27 (FeN_x-C) was constructed using a facile pyrolysis strategy and exhibited superior
28 catalytic activity in peroxymonosulfate (PMS) activation towards organic contaminants
29 oxidation. The various Fe-species and relative amounts of each Fe site in the FeN_x-C
30 catalyst were validated using X-ray absorption spectroscopy and ⁵⁷Fe Mössbauer
31 spectroscopy, which show critical dependencies on the precursor ratio and calcination
32 temperature. The positive correlations between relative content of high-spin state
33 species (Fe^{II} and Fe^{III}) and catalytic performance were found to determine the reactive
34 species generation and electron-transfer pathway in the FeN_x-C/PMS system. Moreover,
35 catalytic performance and theoretical calculation results revealed that the Fe^{II}-N₄ in
36 high spin state (*S*=2) tends to activate PMS to form sulfate and hydroxyl radicals via a
37 one-electron transfer process, while the Fe^{III}-N₄ moiety (*S*=5/2) is prone to high-valent
38 iron species generation with lower free energy. Benefiting from finely-tuned active sites,
39 a single-atom FeN_x-C catalyst achieved favorable applicability in actual wastewater
40 treatment with efficient resistance of common water matrix. The present work advances
41 the mechanistic understanding of spin-state-dependent persulfate activation in single-
42 atom catalysts and provides guidance to design superior catalyst based on spin-state
43 descriptions.

44 **Keywords:** refractory organic pollutant, PMS activation, single iron atom, spin state,

45 wastewater

46 **Synopsis:** A novel and precise spin-state dependent mechanism for wastewater

47 treatment via peroxymonosulfate activation using single-atom Fe–N moieties was first

48 proposed.

49 **Introduction**

50 The extensive use of antibiotics, pharmaceuticals, organic dyes and industrial
51 chemicals has released increasing amounts of organic contaminants in the environment,
52 posing a great threat to the ecological environment and human beings¹. Reactive oxygen
53 species-based advanced oxidation processes (AOPs) have exhibited significant
54 potential in the broadband abatement of refractory pollutants in aqueous systems^{2, 3}.
55 Among these, persulfate-mediated Fenton-like processes including peroxymonosulfate
56 (PMS) and peroxydisulfate (PDS) are attracting enormous interest owing to the
57 generated highly oxidative oxygen species such as sulfate radicals, hydroxyl radicals
58 and superoxide radicals⁴. By cleaving O-O bonds in PMS molecules via electron and
59 energy transfer processes, the most commonly reactive species of sulfate radicals (E^0
60 ($\text{SO}_4^{\bullet-}/\text{SO}_4^{2-}$) = 2.60~3.10 V_{NHE}) and hydroxyl radicals (E^0 ($\bullet\text{OH}/\text{OH}^-$) = 1.90~2.70
61 V_{NHE}) can be formed and further attack organic compounds⁵. Different from hydrogen
62 peroxide (H_2O_2), PMS can also directly oxidize pollutants without radical participation⁶.
63 However, the relatively low steady-state concentration of these species and rapid
64 consumption by impurity substances in water and wastewater still limit PMS activation
65 and decontamination efficiency.

66 Among various transition metals (Co, Fe, Cu and Mn), Fe-based Fenton-like
67 catalysts have developed extensive and persistent interest owing to their low toxicity
68 and cost, efficient ability and environmental sustainability⁷. Recently, the atomically
69 dispersed single-atom catalysts (SACs) have exhibited promising potential in advanced

70 oxidation processes by virtue of the high degree of dispersion of metal sites, single-
71 atom utilization, distinct electronic structure and excellent catalytic capacity⁸⁻¹⁰. For
72 instance, Gao and co-workers¹¹ prepared single iron atoms via a cascade anchoring
73 method as a superior catalyst to activate PMS toward bisphenol A degradation. Guan et
74 al.¹² also demonstrated that single Fe atoms dispersed on carbon nanotubes showed
75 excellent and selective reactivity for organic contaminants over a wide pH range for
76 which FeN₄=O intermediate was the predominant oxidant over Fe-N_x sites. Despite the
77 process of iron single atom in Fenton-like reaction, the catalytic mechanism behind
78 reactive species generation and active sites is still under clarified.

79 It is well known that catalytic ability of SACs is strongly correlated with transition
80 metal species, especially Fe-N_x moieties, which is at the heart of most Fenton-like
81 reactive sites. Very recently, the transition metal sites with different spin states were
82 considered as important aspects for explaining the excellent catalytic performance¹³.
83 For instance, Long et al¹⁴. reported that Co-N-CNTs catalyst with a high spin state
84 favored the PMS adsorption and promoted oxidation ability towards pollutant
85 degradation. Jiang et al¹⁵. demonstrated that electronic spin moment is nearly linear
86 related to the catalytic capacity of Fe-SACs during ORR reaction. Although recent
87 processes have proposed that spin states played essential roles in catalysis process, the
88 exact identity of the structure for each iron-based species with different spin states, their
89 specific effect in Fenton-like reactions and in-depth mechanisms behind Fe^{IV/V}=O and
90 radical species generation are still urgently in need of exploration and clarification.

91 Herein, a single-iron atom catalyst with Fe-N_x configuration was fabricated using
92 a one-step pyrolysis process. The single-atomically dispersed iron species on nitrogen-
93 doped carbon and Fe-pyridinic N₄ moiety were verified as the main active structure by
94 a series of analytical techniques and density functional theory (DFT) calculations. In
95 contrast to previous research, five specific iron-based species with different spin states
96 in FeN_x-C catalysts were finely tuned and clearly observed by XAS and ⁵⁷Fe Mössbauer
97 spectroscopy. During PMS activation and bisphenol A oxidation, only two of these
98 species (high-spin Fe^{II} *S*=2 and Fe^{III} *S*=5/2) significantly contributed to the efficient
99 Fenton-like catalytic performance. The exact process for free radicals and high valent
100 species generation by reactive high-spin Fe^{II}/Fe^{III} was systematically explored and
101 proposed. This work provides a novel and deeper understanding of the particular role
102 for each iron site toward PMS activation and inspires oriented regulation and design of
103 superior catalysts for future environmental remediation.

104 **Experimental Section**

105 Details of all chemicals, materials and preparation methods are listed in the
106 Supporting Information.

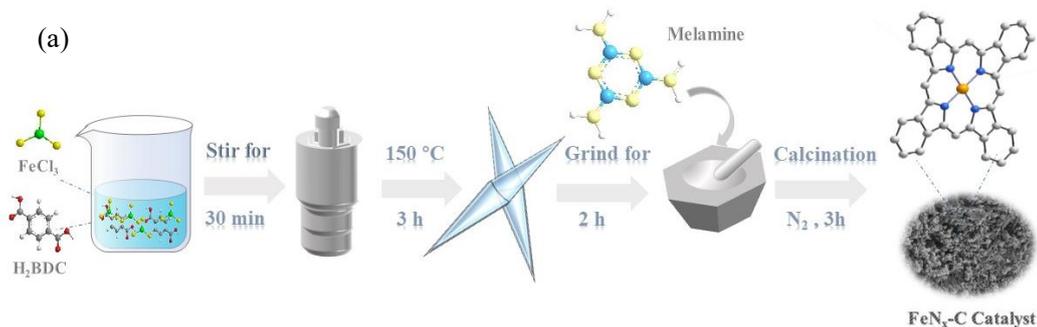
107 **Results and discussion**

108 **Structural analysis of catalysts**

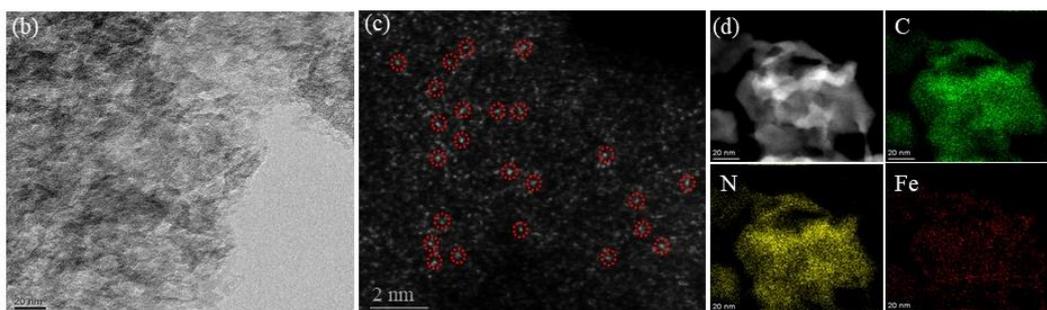
109 The synthesis procedure of FeN_x-C catalysts was demonstrated in Figure 1a via a
110 one-step pyrolysis strategy. The optimal precursor ratio was fixed at 1:5 based on the
111 catalytic capacity (vide infra). Three different temperatures (500, 600, 700°C) were

112 selected to prepare FeN_x-C materials and the chemical structures of this series of
113 catalysts were systematically explored. As shown in Figure S1, the X-ray diffraction
114 pattern (XRD) of the FeN_x-C-600 sample exhibited only two representative diffraction
115 peaks located at 27.5 and 44.5°, which were assigned to the graphitic carbon (JCPDS
116 No. 46-0945). It is worth noting that no additional peaks corresponding to metallic iron
117 or iron oxides were detected, excluding the existence of large crystalline iron-based
118 nanoparticles. Distinctly different from sample FeN_x-C-600, sample FeN_x-C-700
119 showed obvious diffraction peaks from Fe₃C phase. TEM and HRTEM images clearly
120 show the presence of plenty of small nanoparticles (3~5 nm) in FeN_x-C-700 catalysts
121 (Figure S5), while no visible particles or clusters can be observed in FeN_x-C-600
122 samples (Figure 1b). Numerous bright spots with high density dispersed over the N-C
123 matrix were clearly observed in high-angle annular dark-field scanning transmission
124 electron microscopy (HAADF-STEM) images, validating the formation of atomic iron
125 atom in FeN_x-C-600 sample. Energy dispersive X-ray spectrometry (EDX) elemental
126 mapping reveals the highly homogenous distribution of Fe and N elements supported
127 by the C-N substrate layer. Moreover, the Raman spectra show only two characteristic
128 bands at 1348 and 1580 cm⁻¹ corresponding to the defects and planar motion of sp²-
129 hybridized carbon¹⁶ (Figure S6). The relatively high I_D/I_G ratios demonstrate the defect-
130 rich structures in all as-synthesized catalysts¹⁷. The type-IV N₂ adsorption-desorption
131 isotherm curves and correlative pore distributions (Figure S7) corroborated the
132 presence of micro- and mesopores in the catalysts and the specific surface areas (Table

133 S2) gradually increase with increasing the pyrolysis temperature.



134



135

136 Figure 1. (a) Synthesis process of FeN_x-C catalysts; (b) TEM images; (c) HAADF-STEM image

137 and (d) energy-dispersive X-ray mapping of FeN_x-C-600 catalyst

138 **Chemical state and coordination environment of FeN_x-C catalyst**

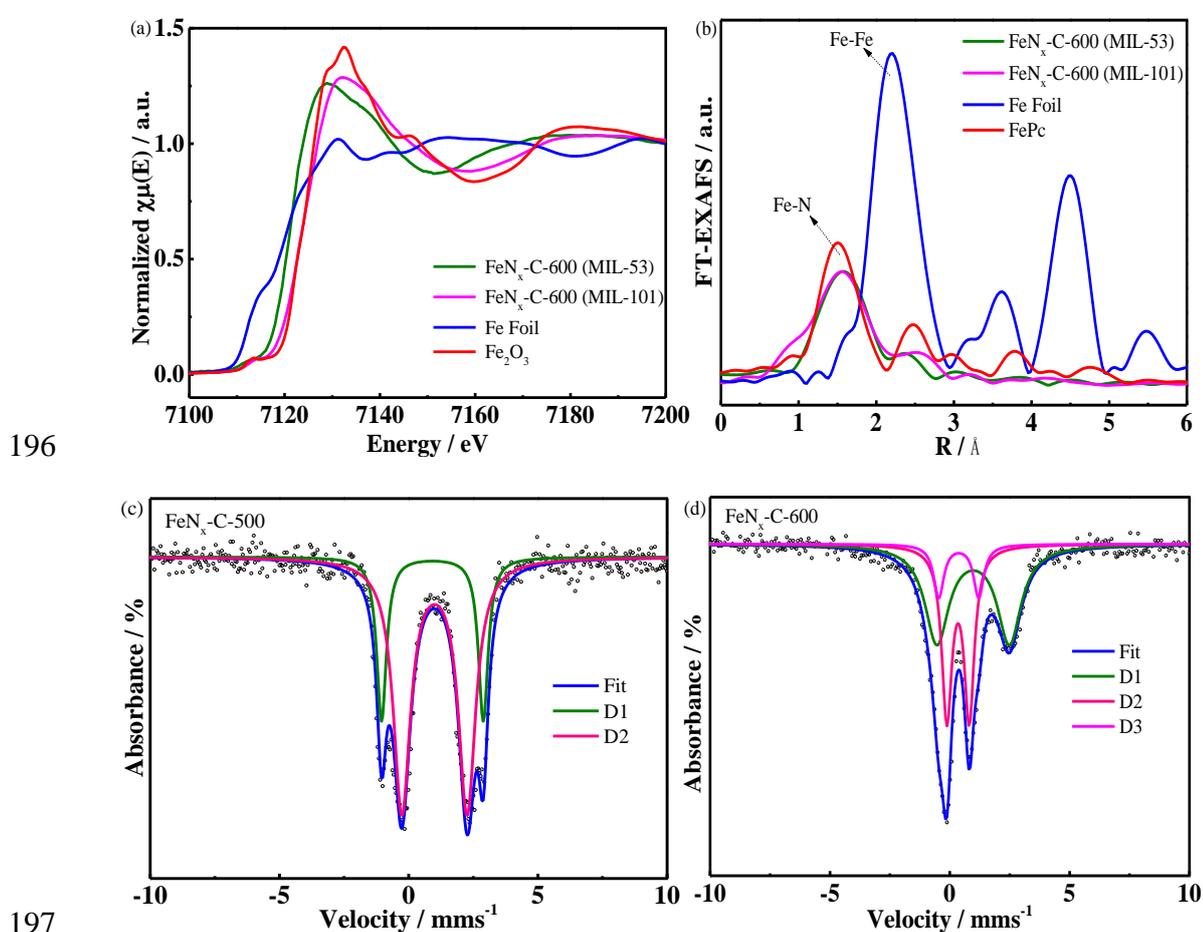
139 The chemical state and local coordination environment are essential to the
140 exploration of catalytic reaction. The Fe 2p XPS spectra of FeN_x-C catalysts (Figure
141 S8) revealed two characteristic peaks at ~710 and 724 eV, assigning to Fe 2p_{3/2} and Fe
142 2p_{1/2}¹⁸. As depicted in Figure 2a and Figure S10, the Fe K-edge X-ray absorption near
143 edge structure (XANES) spectrum of sample FeN_x-C-600 indicated the energy
144 absorption threshold (edge) energy is located between those of Fe foil and Fe₂O₃
145 references, indicating that the Fe atoms in the FeN_x-C-600 sample possess a positive
146 charge between Fe⁰ and Fe³⁺. It can be found that XANES spectrum of sample FeN_x-
147 C-600 was very similar to the FePc reference, which might be derived from the

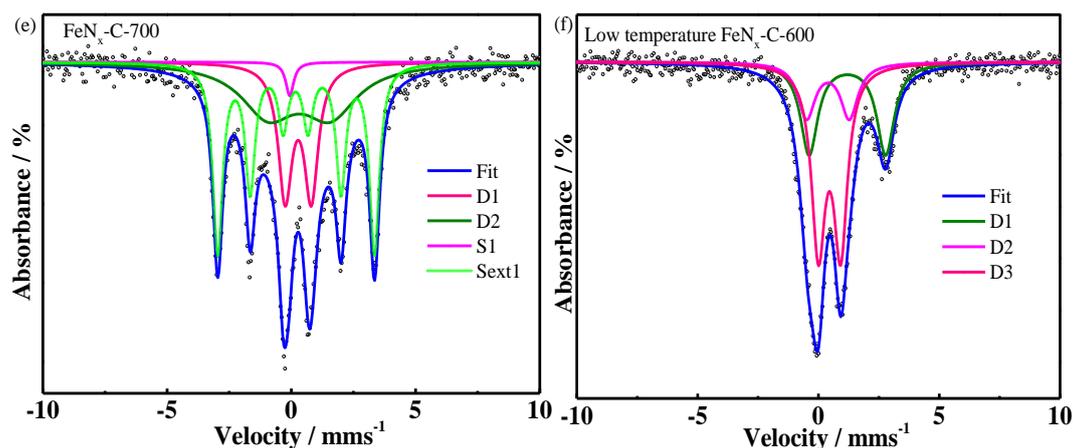
148 predominance of Fe-N₄ coordination sites in the FeN_x-C-600 sample. The Fourier
149 transform (FT)- extended X-ray absorption fine structure (EXAFS) curve of sample
150 FeN_x-C-600 exhibited only a strong peak at ~1.5 Å (Figure 2b), corresponding with the
151 Fe-N configuration¹⁹. The fitting parameters of the first coordination shell illustrated an
152 average Fe-N coordination number of 4.1 and an average Fe-N bond length of 2.06 Å.
153 To further confirm the combination environment, possible Fe-N_x coordination models
154 including pyridinic N, pyrrolic N and graphitic N were explored using DFT calculations.
155 As shown in Figure S11, the lowest formation energy was found in the structure of Fe-
156 pyridinic N₄, signifying that pyridinic N is more conducive to occupying the isolated
157 Fe atoms with a four-coordinated site. Based on the DFT calculations and EXAFS
158 results, Fe single atoms coordinated with an average of four pyridinic N atoms is the
159 main configuration in FeN_x-C catalysts.

160 ⁵⁷Fe Mössbauer spectroscopy is an excellent technique for identifying different
161 iron-based materials via recoil free absorption of γ -rays using ⁵⁷Fe nuclei. Our previous
162 study showed that the ⁵⁷Fe Mössbauer spectrum for pristine MOF precursor was
163 composed of one doublet consistent with 100% Fe^{III} species in an octahedral geometry²⁰.
164 After calcination at different temperatures, the ⁵⁷Fe Mössbauer spectra for FeN_x-C
165 catalysts indicated the presence of different Fe species and coordination environment
166 (Figure 2c-f and Table S7). Specifically, the Mössbauer spectra for both samples FeN_x-
167 C-500 and FeN_x-C-600 can be only fitted with doublet and no singlet and sextet,
168 excluding the existence of zero-valence Fe crystalline phases. On the contrary, not only

169 doublets but also singlets and sextets assigned to Fe^0 and Fe_3C species²¹ were clearly
170 detected in the $\text{FeN}_x\text{-C-700}$ sample, which is in good accordance with the results of
171 XRD and XPS. For the $\text{FeN}_x\text{-C-500}$ sample, the two doublets with larger IS (isomer
172 shift, > 0.80 mm/s) and QS (quadrupole splitting) values, can be assigned to the high-
173 spin Fe^{II} species ($S=2$)²¹. The spectrum for sample $\text{FeN}_x\text{-C-600}$ can be fitted robustly
174 with three different doublets (D1, D2, D3). Among these, the D1 exhibited similar
175 parameters to the doublet in the spectrum for sample $\text{FeN}_x\text{-C-500}$, and is assigned to
176 high-spin $\text{Fe}^{\text{II}}\text{-N}_4$ sites ($S=2$). D2, with a relatively low QS value (0.3~1.0 mm/s) and
177 an IS value of ~ 0.34 mm/s was ascribed to high spin $\text{Fe}^{\text{III}}\text{-N}_4$ sites ($S=5/2$) with an
178 octahedral structure due to the pulling force between Fe atoms and N_4 plane^{22, 23}.
179 Different from D2, D3 exhibited a similar isomer shift value (~ 0.37 mm/s) but a much
180 higher QS value (~ 1.67 mm/s), which can be assigned to low spin $\text{Fe}^{\text{III}}\text{-N}_4$ species
181 ($S=1/2$)²⁴. In the spectrum for the $\text{FeN}_x\text{-C-700}$ sample, approximately 22.0% high-spin
182 Fe^{III} ($S=5/2$) was still preserved, suggesting the relative stability of this structure at
183 higher pyrolysis temperatures. Additionally, a similar IS value and larger QS value (D2)
184 than for high-spin Fe-N_4 structures attributed to the appearance of medium-spin ferrous
185 species ($S=1$), which is close to the configuration of ferrous iron phthalocyanine.
186 Concurrently, iron carbide and zero valent iron nanoparticles are formed and aggregated,
187 accompanying the destruction of the low-spin ferric structure, which is consistent with
188 TEM images and XRD patterns. To further explore the coordination of iron and nitrogen
189 in sample $\text{FeN}_x\text{-C-600}$, low temperature ^{57}Fe Mössbauer spectroscopy was performed

190 at 77K (Figure 2f). Notably, both room temperature and low temperature ^{57}Fe
191 Mössbauer spectra were devoid of any singlet or sextet assignable to nanometer-sized
192 iron particles, ruling out the existence of iron oxides, iron carbide and zero-valent irons
193 species in the $\text{FeN}_x\text{-C-600}$ catalyst²⁵. On the basis of these results, it can be concluded
194 that atomically dispersed iron atoms coordinated with four pyridinic nitrogen atoms in
195 the $\text{FeN}_x\text{-C}$ catalyst was successfully fabricated.





198

199

200

201

202

Figure 2. (a) K-edge XANES spectra of sample FeN_x-C-600, Fe₂O₃, Fe foil, FePc, (b) Fourier transforms of the *k*³-weighted K-edge EXAFS spectra of sample FeN_x-C-600, Fe foil, FePc, (c-f) Room temperature and low temperature ⁵⁷Fe Mössbauer spectra of samples FeN_x-C-500/600/700

Evaluation of catalytic activity of FeN_x-C catalysts

203

204

205

206

207

208

209

210

211

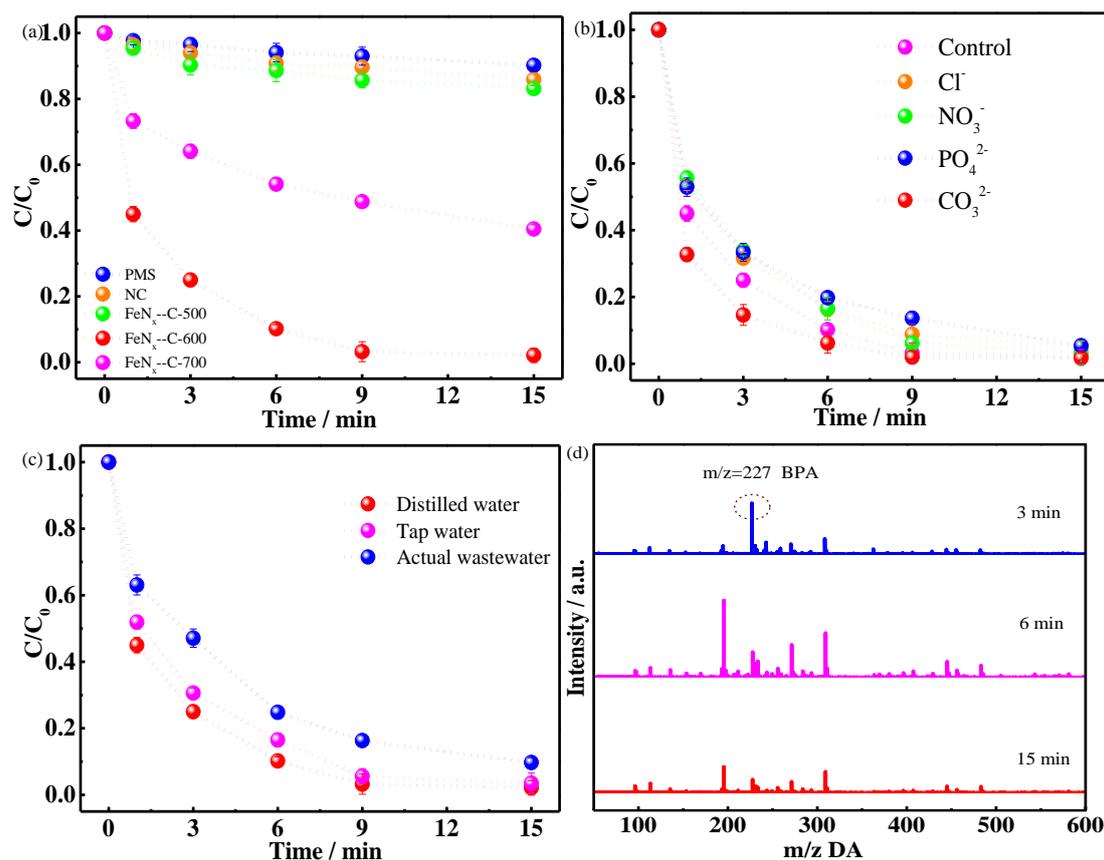
212

213

214

The adsorption and decomposition toward bisphenol A (BPA) were evaluated using as-synthesized FeN_x-C catalysts. As depicted in Figure 3a, PMS alone and NC catalyst exhibited negligible activity toward BPA decomposition. In comparison with sample FeN_x-C-600 (99.5%), only 16.9% and 59.5% of BPA were removed by FeN_x-C-500 and FeN_x-C-700, respectively. The pseudo-first-order kinetic rate constant of sample FeN_x-C-600 exceeded those of samples FeN_x-C-500 and FeN_x-C-700 by factors of 16.23 and 9.92, respectively (Figure S15a). Moreover, the precursor ratio (Fe-MOF/melamine) at 1:5 exhibited a more rapid BPA removal than the other two mixing ratios (1:3 and 1:7, Figure S14c). The distinct difference in the Fenton-like ability between these FeN_x-C catalysts may be ascribed to the differences in surface area and pore structure, different Fe-species (Fe⁰, Fe₃C, Fe-N, Fe-O, etc.) and iron coordination environments²⁶. To better understand the relationship between reactivity and surface properties, a specific activity

215 (k -specific), through normalizing the rate constant by the catalyst concentration and
216 specific surface area¹¹, is applied to reveal the truly intrinsic capacity of FeN_x-C
217 catalysts. Notably, the k -specific value of sample FeN_x-C-600 (0.013 L m⁻² min⁻¹) is 2.0
218 and 22.8 times higher than for samples FeN_x-C-500 (0.0065 L m⁻² min⁻¹) and FeN_x-C-
219 700 (0.00057 L m⁻² min⁻¹), as well as those previously reported Fenton-like catalysts
220 (Table S5), highlighting the superior catalytic performance of single iron atom
221 decorated catalyst for PMS activation. Furthermore, more than 90% of BPA and phenol
222 degradation can be removed, while only 14.3% of carbamazepine (CBZ), 48.5% of
223 methylene blue (MB) and 44.2% Rhodamine B (RhB) were totally decomposed into
224 carbon dioxide and water (Figure S13b-d), implying a certain level of selectivity
225 towards different aromatic organic pollutants decontamination. Moreover, Figure 3b-c
226 showed that the FeN_x-C-600/PMS system could maintain a stable and superior catalytic
227 ability not only in the presence of various anion ions but also for the degradation of
228 BPA in tap water and actual wastewater, indicating a robust resistance to the background
229 matrix. These phenomena indicated that there might be other dominantly reactive
230 oxidizing species instead of radicals in the FeN_x-C-600/PMS system. Apart from
231 degradation efficiency, approximately 75% of total organic carbon (TOC) can be totally
232 removed within 30 min, indicating outstanding decomposition capacity. The detailed
233 decomposition intermediates of BPA in treated solution were detected using mass
234 spectrometry and the possible degradation pathway was correspondingly proposed
235 (Figure S16-17), implying the superior catalytic capacity of FeN_x-C-600 catalyst.



236

237

238 Figure 3. (a) Kinetics of BPA degradation via PMS activation with as-synthesized catalysts; (b)

239 Influence of various anions on BPA degradation in the FeN_x-C-600/PMS system; (c) Kinetics of

240 BPA degradation in tap water and actual wastewater via PMS activation with FeN_x-C-600 catalyst;

241 (d) Mass spectrometry chromatogram for BPA degradation in FeN_x-C-600/PMS system. Reaction

242 conditions: catalyst dosage = 0.2 g/L, [PMS]₀ = 0.50 mM, [BPA]₀ = 20 mg/L, pH₀ = 6.5, [Cl⁻]₀ =

243 $[\text{NO}_3^-]_0 = [\text{PO}_4^{2-}]_0 = [\text{CO}_3^{2-}]_0 = 10.0 \text{ mM}$

244 Identification of active species and sites

245 To explore the contribution of radical species in the FeN_x-C-600/PMS system,

246 scavenging experiments and electron paramagnetic resonance (EPR) spectroscopy

247 measurements were conducted. Herein, methanol (MeOH) and tert-butyl alcohol (TBA)

248 were chosen as quenchers for scavenging $\cdot\text{OH}/\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ species, respectively²⁷.

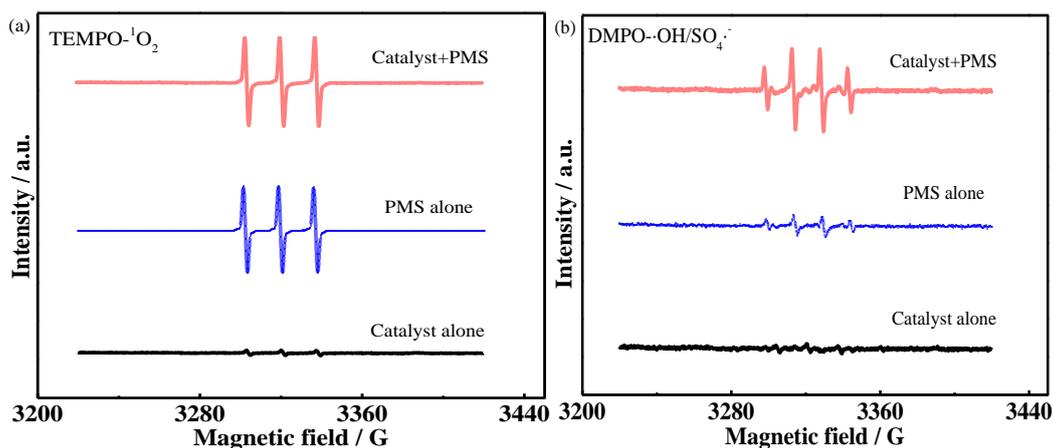
249 As depicted in Figure S18, the MeOH, at the concentration 400 times of PMS, exhibited
250 a slight inhibition of BPA degradation, demonstrating that $\cdot\text{OH}/\text{SO}_4^{\cdot-}$ are not the major
251 active species. Moreover, the insignificant inhibition effect of TBA further supported
252 this hypothesis. As shown in Figure 4, no EPR signal was detected with catalyst alone.
253 In contrast, an intensive signal from 5,5-dimethyl-1-pyrroline N-oxide DMPO- $\cdot\text{OH}$
254 adduct and DMPO- $\text{SO}_4^{\cdot-}$ adduct appeared after 5 min in the $\text{FeN}_x\text{-C-600/PMS}$ system
255 compared with PMS alone, implying that the generated $\cdot\text{OH}/\text{SO}_4^{\cdot-}$ radicals do
256 contribute to BPA decomposition but they are not the dominant oxidant species during
257 the Fenton-like reaction. In addition to the singlet oxygen, triplet signals with similar
258 intensity of TEMPO adduct (2,2,6,6-tetramethyl-4-piperidine-N-oxyl) were observed
259 for PMS alone and the $\text{FeN}_x\text{-C-600/PMS}$ system after 5min, excluding the formation
260 of $^1\text{O}_2$ during PMS activation and BPA degradation. Moreover, the weak intensity of
261 DMPO- $\text{O}_2^{\cdot-}$ signals in EPR spectra (Figure S18c) was similar to that of PMS alone
262 system, further excluding the formation of $\text{O}_2^{\cdot-}$ radicals in $\text{FeN}_x\text{-C-600/PMS}$ system.

263 The N-doped carbon and Fe atoms coordinated with N-doped carbon material, have
264 been reported to play essential roles in Fenton-like reactions²⁸. As expected, the
265 addition of oxalate evidently depressed the catalytic performance (Figure S19a),
266 verifying the important role of iron sites for PMS activation and pollutant
267 decomposition. Furthermore, DFT calculations were performed to explore PMS
268 adsorption on these possible catalytic sites. As shown in Figure 4c, the adsorption
269 energy on Fe-pyridinic $\text{N}_4\text{-C}$, pyridinic N-C , Fe-pyrrolic $\text{N}_4\text{-C}$, pyrrolic N-C and Fe-

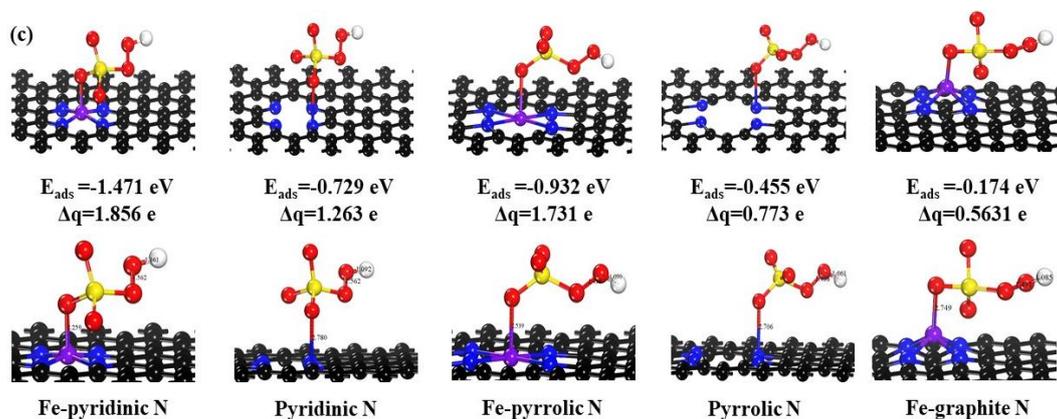
270 graphitic N₄-C were calculated to be -1.471, -0.729, -0.932, -0.455 and -0.174 eV,
271 respectively. The O-O bond length (*l*_{O-O}) in PMS was 1.562, 1.438 and 1.490 Å after
272 adsorbing on the Fe-pyridinic N₄-C, Fe-pyrrolic N₄-C, and Fe-graphitic N₄-C,
273 respectively, which were obviously elongated compared with pure PMS molecules
274 (1.410 Å)¹¹. It is noted that Fe-pyridinic N₄-C interacts more strongly with HSO₅⁻ in
275 PMS molecules compared with other configuration models due to the larger adsorption
276 energy and longer O-O bond length. Moreover, the order of the Bader charge followed
277 Fe-pyridinic N₄ > Fe-pyrrolic N₄ > pyridinic N₄ > pyrrolic N₄ > graphitic N₄ > Fe-
278 graphitic N₄ (Figure 4c), further confirming that PMS molecule is prone to bond with
279 Fe-pyridinic N₄ moiety rather than nitrogen doped carbon material, which obviously
280 accelerated the charge transfer in the FeN_x-C/PMS system. The control homogenous
281 experiments with the addition of FeCl₂ and Fe(NO₃)₃ (1.87 μmol, the equal Fe content
282 of sample FeN_x-C-600) were carried out (Figure S19b). As expected, the negligible
283 effects of FeCl₂ and Fe(NO₃)₃ completely validate the reactive sites for this Fenton-like
284 reaction are Fe-pyridinic N₄ moiety and not Fe ions. Additionally, acid etching
285 treatment of FeN_x-C-600 catalyst was performed to remove the possible existence of
286 surface Fe clusters or tiny nanoparticles rather than Fe-N_x species. With soaking in 2M
287 HCl 60 °C for 3h, the negligible change in XRD pattern, TEM images and insignificant
288 suppression of BPA removal rate (Figure S20) validate the dominant composition and
289 active site of Fe-N_x in the FeN_x-C-600/PMS system.

290 The previous studies reported Fe^{III}/Fe^{II} single atom attached to various carbon

291 materials can efficiently activate PMS or H₂O₂ to generate high-valent iron species
292 (marked as HFeS)²⁹⁻³¹. Considering the faint contribution of SO₄^{•-}/•OH radicals, the
293 confirmation of HFeS becomes indispensable in the FeN_x-C-600/PMS/BPA system. To
294 quantitatively analyze the HFeS species, methyl phenyl sulfoxide (PMSO) and methyl
295 phenyl sulfone (PMSO₂) MS spectrometry were conducted and the corresponding
296 spectra are illustrated in Figure S21. It can be clearly observed that abundant PMSO₂
297 was detected and the intensity of PMSO₂ gradually increases with time, and finally
298 levels out at a yield of around 75%. Furthermore, ¹⁸O isotope-labeled products were
299 detected using ultra-performance liquid chromatography coupled with high resolution
300 time-of-flight mass spectrometry (UPLC-TOF-MS) via the oxygen atom exchange
301 process between H₂¹⁸O and high valent iron species (Figure S21 g-j). Two characteristic
302 peaks observed at m/z 157.0321 and 159.0361 in the MS spectrum are assigned to the
303 PMS¹⁶O¹⁶O and PMS¹⁶O¹⁸O. Moreover, two prominent peaks with m/z 78.9853 and
304 80.9894 represented the characteristic fragments of PMS¹⁶O¹⁶O ([M-C₆H₅]⁺) and
305 PMS¹⁶O¹⁸O ([M-C₆H₅]⁺) can be also detected in FeN_x-C/PMS system. Contrarily, only
306 fragment of PMS¹⁶O¹⁶O but no PMS¹⁶O¹⁸O was detected in PMS alone system,
307 verifying the successful integration of ¹⁸O to PMSO₂ using as-prepared FeN_x-C
308 catalyst^{32, 33}. On the basis of these analysis, the relatively high selectivity towards
309 phenolic organic pollutants and the significant inhibition of various anion ions can be
310 reasonably explained by the electrophilic addition reaction via dominant HFeS species.



311



312

313 Figure 4. (a) TEMP-trapped EPR spectra for $^1\text{O}_2$; (b) DMPO-trapped EPR spectra of $\text{SO}_4^{\bullet-}/\bullet\text{OH}$;

314 (c) The adsorption energy on different $\text{FeN}_x\text{-C}$ and N-C sites and the corresponding O-O bond

315 length in PMS and Bader charge after PMS adsorption. Reaction conditions: catalyst dosage = 0.2

316 g/L, $[\text{PMS}]_0 = 0.50$ mM, $[\text{BPA}]_0 = 20$ mg/L, $\text{pH}_0 = 6.5$

317 To understand the specific reaction process in the $\text{FeN}_x\text{-C-600/PMS}$ system, *in situ*

318 Raman spectra were carried out. As depicted in Figure 5a, the peak located at 1060 cm^{-1}

319 1 is assigned to the HSO_5^- in PMS, while the peak at 982 cm^{-1} represents the S=O bond

320 in SO_4^{2-} with symmetric stretching vibration³⁴. Notably, the I_{1060}/I_{982} value significantly

321 decreased from 1.02 to 0.68 in the corresponding $\text{FeN}_x\text{-C/PMS}$ system, demonstrating

322 the rapid conversion of HSO_5^- into SO_4^{2-} via the intense interaction between PMS and

323 Fe-N_4 sites. Moreover, the dynamic PMS concentration was detected via potassium

324 iodide spectrophotometry⁴. It can be observed that the PMS concentration gradually
325 declined with increasing reaction time and the tendency was almost synchronous with
326 the BPA degradation process (Figure 5b), amply validating the surface chemical
327 evolution of Fe-N sites for PMS decomposition rather than adsorption.

328 Considering that ⁵⁷Fe Mössbauer spectroscopy can particularly distinguish various
329 Fe phases with similar structure but in different oxidation and spin states, the effects of
330 different Fe sites in FeN₄ moiety for radical generation, high-valent species formation,
331 PMS activation and BPA decomposition were systematically investigated in this work.
332 According to the specific catalytic activity as well as their relative content in Mössbauer
333 spectra, the intermediate-spin Fe^{II} species (*S*=1) can be first excluded from being
334 dominant species. The medium-spin configuration was caused by the interaction of
335 central Fe ions (*S*=1) with four nitrogen atoms, which are bonded with the surrounding
336 carbon atoms yielding in pseudo-6-fold-octahedral coordination of iron³⁵. The fully
337 filled 3d_{z²} orbitals in Fe^{II} sites make it difficult to approach peroxymonosulfate and
338 pollutant molecules. Furthermore, this intermediate-spin Fe^{II} coordination (relative
339 amount of 32.0%) only exists in the FeN_x-C-700 catalyst, for which the specific activity
340 was just 0.00057 L m⁻² min⁻¹, excluding a major contribution of this specie to the
341 Fenton-like activity. For the high-spin Fe^{II} species (*S*=2), its larger IS value
342 demonstrates a higher density of 3d electrons. Although its relative amounts in samples
343 FeN_x-C-5-500, FeN_x-C-3/5/7-600, were highest, here it did not contribute the most to
344 BPA degradation. This is because decreasing the pyrolysis temperature to 500°C

345 improved the relative Fe content present in high-spin Fe^{II} state, without obviously
346 increasing the catalytic performance. However, considering the higher specific activity
347 compared with sample FeN_x-C-700, the high-spin Fe^{II} species in sample FeN_x-C-500
348 did provide a certain enhancement in PMS activation and BPA decomposition. This
349 might be ascribed to high-spin Fe^{II} configuration with $S=2$ and $3d_z^2$ orbital occupied by
350 single electron being conducive to directly activate PMS via electron transfer
351 pathways³⁶. Notably, by comparing the Fenton-like activity with the relative abundance
352 of each Fe component, the superior catalytic activity of FeN_x-C catalyst closely
353 correlates with its relatively larger amount of high-spin Fe^{III} species ($S=5/2$), with a
354 coefficient of determination R^2 of 0.995 (Figure S22e-f). In comparison with sample
355 FeN_x-C-3-600, the higher HS Fe^{III} and lower LS Fe^{III} phase in sample FeN_x-C-5-600
356 resulted in excellent catalytic efficiency toward BPA removal, which can be also
357 identified by the FeN_x-C-7-600 catalyst. When the calcination temperature increased to
358 700 °C, the relative content of high-spin Fe^{III} in sample FeN_x-C-700 was only 22.0%,
359 far below the content in sample FeN_x-C-5-600. The relative content of high-spin
360 Fe^{III}/Fe^{II} was positively correlated with the catalytic capacity, indicating that the
361 efficiency of PMS activation and BPA oxidation was governed by the spin state of Fe-
362 N_x sites owing to the space electronic configuration differentiation in d orbitals.
363 Generally, the high-spin state of Fe-N_x sites with large magnetic moments is conducive
364 to participating in σ -bonding with PMS molecule via overlapping with O 2p orbitals³⁶,
365 thereby improving PMS adsorption ability, promoting the oxidation potential and

366 accelerating electron transfer during the Fenton-like reaction. In contrast, the structure
367 of low-spin Fe^{III} is highly saturated and the number of unpaired electrons with the same
368 spin direction is the smallest²³, making it difficult to be a reactive site. Moreover, the
369 newly formed Fe₃C (44.4%) and Fe⁰ (Super-paramagnetic Fe, 1.6%) in FeN_x-C-700 are
370 unlikely to significantly improve the pollutants removal efficiency, excluding the
371 dominant role of these species in the FeN_x-C/PMS/BPA system.

372 Furthermore, DFT calculations were undertaken to shed light on the pathway of
373 PMS activation into SO₄^{·-}, ·OH radicals and Fe(V)=O complex via the proposed Fe-
374 pyridinic N₄ configurations with different spin-state. The energy profiles and
375 corresponding intermediate structure were shown in Figure 5c-d. Considering that
376 single-electron, two-electron and oxygen atom transfer pathways³⁷ may occur
377 simultaneously in the FeN_x-C/PMS system, multiple binding configurations and the
378 corresponding free energies were calculated. When the PMS molecule was adsorbed on
379 the Fe-pyridinic N₄ sites, it can be rapidly activated and subsequently generate
380 intermediate (II) and transition state (TS), and finally form reactive oxygen species
381 (SO₄^{·-}, ·OH, Fe(V)=O). As depicted in Figure 5c, the Gibbs free energies for formation
382 of SO₄^{·-}, ·OH radicals were much lower than the reactant, demonstrating that it is
383 thermodynamically favorable. Among them, the most favorable active site for SO₄^{·-},
384 ·OH radical formation with the lowest Gibbs free energy was high-spin Fe^{II} rather than
385 the Fe^{III}-N₄ moiety. Contrarily, the free energy for generating high valent iron species
386 Fe(V) by Fe^{II}-N₄ is about 0.59 and 0.64 eV, which is an endothermic process and cannot

387 happen automatically under ambient conditions. Different from Fe^{II}-N₄ moiety, the
388 energy barrier for Fe(V)=O complex formation is easier to overcome by high-spin Fe^{III}-
389 N₄ sites, which is an exothermic reaction and favorable in thermodynamics and kinetics.
390 In other words, the high spin Fe^{III} $S=5/2$ and Fe^{II} $S=2$ sites are much more reactive to
391 bond with PMS to generate Fe(V)=O complexes and SO₄⁻, ·OH radicals, respectively,
392 which is consistent with the experimental results.

393 Low temperature EPR measurement, which could provide detailed information of
394 iron species with half-integer spin and the potential presence of high valent iron
395 components in the redox reaction³⁸, was carried out at 77K. As shown in Figure 5e and
396 Figure S23a, an isotropic signal at $g=2.0034$ was detected in the FeN_x-C-600/H₂O
397 system, ascribed to the partial 3d⁵ orbitals of Fe^{III} with large zero-field splitting and
398 unpaired electrons³⁹. Notably, after mixing PMS with FeN_x-C-600 catalyst, the
399 intensity of this signal was enhanced and exhibited a positive shift to a lower value.
400 This phenomenon was consistent with the previous report by Dashdorj et al.,⁴⁰ likely
401 originating the generation of Fe^V=O via oxidizing high-spin Fe^{III} sites ($S=5/2$) during
402 PMS conversion process. Additionally, a broad isotropic signal at $g=4.282$ originating
403 from rhombic high-spin ferric species was also clearly detected³⁸. In comparison with
404 the deionized water system, the increased intensity of ferric sites after addition of PMS
405 demonstrated more production of Fe^{III} species, which are mainly derived from the redox
406 characteristics of Fe^{II}/Fe^{III} and Fe^{III}/Fe^V during activation of PMS via atomically
407 dispersed iron-N sites. Moreover, the current density displayed an obvious drop with

408 the addition of PMS (Figure 5f), likely ascribed to electron redistribution owing to the
 409 interaction between PMS and the surface of the catalyst. The subsequent introduction
 410 of BPA provoked a remarkable decline in current density, indicating the occurrence of
 411 electron transfer between BPA, PMS and FeN_x-C-600 catalyst.

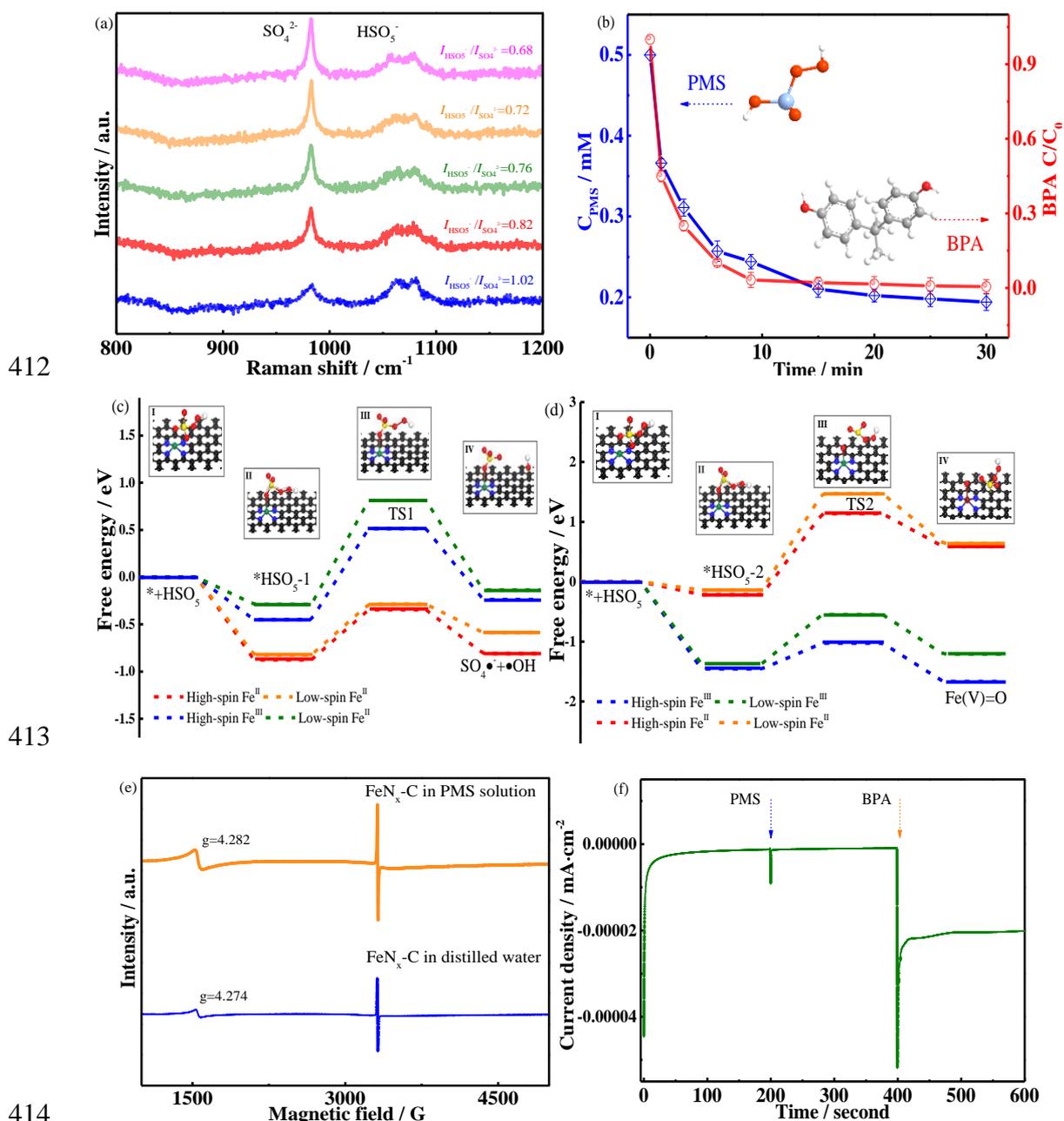
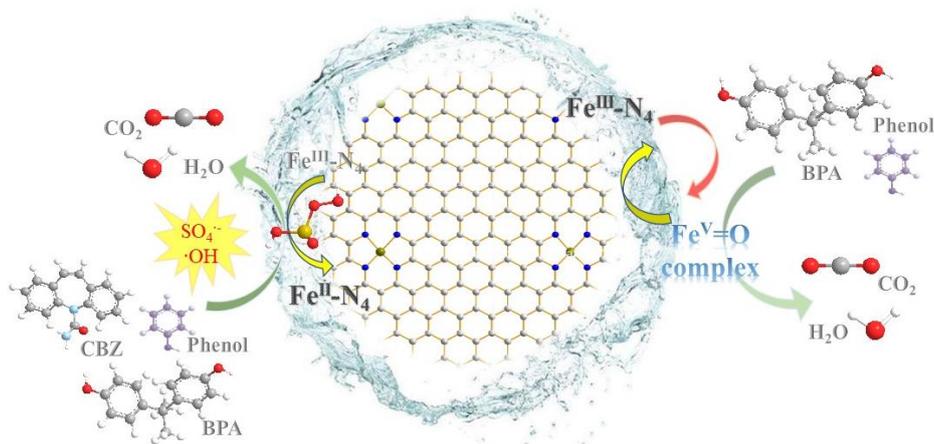


Figure 5. (a) *In situ* Raman spectra of FeN_x-C-600/PMS system measured at different reaction
 times; (b) Variation of PMS concentration and BPA removal rate in the FeN_x-C-600/PMS/BPA

417 system; (c-d) Reaction pathways of $\text{SO}_4^{\cdot-}$, $\cdot\text{OH}$ radicals and $\text{Fe}(\text{V})=\text{O}$ generation at different spin
418 state Fe sites; (e) EPR spectra of $\text{FeN}_x\text{-C-600}$ in distilled water and PMS solution under 77K; (f)
419 Current response at the $\text{FeN}_x\text{-C-600}$ coated working electrode; Reaction conditions: catalyst
420 dosage = 0.2 g/L, $[\text{PMS}]_0 = 0.50$ mM, $[\text{BPA}]_0 = 20$ mg/L, $\text{pH}_0 = 6.5$

421 On the basis of experimental results and theoretical calculations, the overall
422 mechanism for PMS activation and pollutant oxidation using $\text{FeN}_x\text{-C}$ catalyst was
423 proposed in Figure 6. The PMS molecule could first react with Fe-pyridinic N_4 moieties
424 including $\text{Fe}^{\text{II}}\text{N}_4$ and $\text{Fe}^{\text{III}}\text{N}_4$ sites to form $\text{FeN}_x\text{-(HO)OSO}_3^-$ intermediates. Due to that
425 PMS can be decomposed via cleavage of the O-O bond, and the formed $\text{Fe}^{\text{II}}\text{N}_x\text{-}$
426 $(\text{HO)OSO}_3^-$ was further activated to generate $\text{SO}_4^{\cdot-}$ radicals via a one-electron transfer,
427 while the $\text{Fe}^{\text{III}}\text{N}_x\text{-(HO)OSO}_3^-$ could form high-valent species ($\text{Fe}^{\text{V}}=\text{O}$ complex) via a
428 two-electron and O-atom transfer⁴¹. The $\text{SO}_4^{\cdot-}$ radicals can further react with H_2O in the
429 solution to yield $\cdot\text{OH}$. Benefiting from the effect of predominant high-valent iron
430 species and supporting $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$ radicals, bisphenol A can be efficiently
431 decomposed into small organic substrates, CO_2 and H_2O . Owing to the consumption of
432 transient $\text{Fe}^{\text{V}}=\text{O}$ complex by bisphenol oxidation, the Fe^{V} can be reduced and
433 transformed into Fe^{III} , achieving redox circulation between Fe^{V} and Fe^{III} . After the
434 Fenton-like reaction, the relative content of Fe^{II} and Fe^{III} remains almost unchanged, as
435 demonstrated by ^{57}Fe Mössbauer spectroscopy (Figure S22c) and Fe 2p XPS
436 spectroscopy (Figure S25a), demonstrating an efficient conversion of $\text{Fe}^{\text{III}}/\text{Fe}^{\text{II}}$ and
437 $\text{Fe}^{\text{V}}/\text{Fe}^{\text{III}}$ in the $\text{FeN}_x\text{-C/BPA/PMS}$ system.



438

439 Figure 6. Proposed mechanism behind PMS activation over $\text{FeN}_x\text{-C-600}$ catalyst toward organic
 440 pollutants decomposition

441 To furnish evidence of the universal applicability of this preparation strategy and
 442 mechanistic exploration, a similarly structured Fe-MOF, named MIL-101(Fe) was
 443 chosen as a precursor to synthesize single atom catalyst using a procedure being parallel
 444 to that of the MIL-53 based $\text{FeN}_x\text{-C}$ sample. The XRD, HRTEM, XAS results indicated
 445 the successful construction of an atomic structure model of the $101\text{FeN}_x\text{-C}$ catalyst
 446 (Figure S24). The ^{57}Fe Mössbauer spectra of $101\text{FeN}_x\text{-C}$ can be deconvoluted into three
 447 doublets based on IS and QS values (Figure S22d), which are assigned to the Fe- N_4
 448 moieties with different local coordination structures (high-spin Fe^{II} $S=2$, high-spin Fe^{III}
 449 $S=5/2$ and low-spin Fe^{III} species $S=1/2$). The coordination of these three components
 450 was similar to the MIL-53 based $\text{FeN}_x\text{-C}$ sample, indicating the existence of similar Fe
 451 single atomic structures. Notably, the relative amount of high-spin Fe^{III} is close to that
 452 in the MIL-53 based $\text{FeN}_x\text{-C}$, while high-spin Fe^{II} site is lower and low-spin Fe^{III} is
 453 larger than for the MIL-53 based $\text{FeN}_x\text{-C}$ sample (Table S7). This phenomenon could
 454 explain the slightly decreased removal efficiency of BPA by sample $101\text{FeN}_x\text{-C}$

455 ($k=0.268 \text{ min}^{-1}$) compared with MIL-53 based $\text{FeN}_x\text{-C}$ ($k=0.357 \text{ min}^{-1}$), further
456 validating the dominant role of high-spin Fe^{III} and supporting effect of high-spin Fe^{II}
457 during the Fenton-like reaction.

458 In view of the durability and practicability, a continuous batch experiment was
459 conducted. With the number of cycles increased, approximately 89.8% BPA
460 degradation efficiency can be still achieved by sample $\text{FeN}_x\text{-C-600}$ (Figure S26). The
461 total leached iron (0.37 mg/L) did not give rise to significant catalytic ability towards
462 PMS activation and BPA degradation (Figure S26d). The XPS spectrum of sample
463 $\text{FeN}_x\text{-C-600}$ indicated that the relative amounts of N and Fe slightly decreased (Figure
464 S25), while O content increased after cyclic experiments. These results illustrate that
465 the attenuation of removal efficiency might be ascribed to the slight loss of Fe and N
466 content¹¹, resulting in a reduced number of active sites for PMS activation and pollutant
467 degradation, which is consistent with the result of ^{57}Fe Mössbauer spectra of used
468 catalysts (Figure S22). Furthermore, XRD pattern and FTIR spectra (Figure S26b-c)
469 signify the highly stable crystal structure and no other Fe species aggregation on the
470 catalyst surface. The practicability of sample $\text{FeN}_x\text{-C-600}$ toward BPA decontamination
471 in tap water and realistic industrial wastewater was also tentatively evaluated.
472 Wastewater was collected from the effluent of the Tokyo Sewerage system in Tokyo,
473 Japan. The parameters of the wastewater were summarized in Table.S6. As depicted in
474 Figure S26d-e, sample $\text{FeN}_x\text{-C-600}$ still fulfilled high removal efficiency of BPA in
475 actual wastewater and tap water via PMS activation. Considering the extremely low

476 concentration of metallic ions in the wastewater sample, the outstanding catalytic
477 performance for BPA oxidation was mainly contributed to the catalyst itself, which
478 confirm the distinct advantage of catalyst FeN_x-C-600 in engineering to remediate
479 actual wastewater.

480 **Environmental Implications**

481 A single-atom iron catalyst with Fe-pyridinicN₄ configuration was successfully
482 synthesized using a facile pyrolysis strategy in this work. As a PMS activator, catalyst
483 FeN_x-C-600 delivered superior catalytic activity for refractory organic pollutant
484 decomposition, for which the specific activity was much higher than the control
485 experiments and various previously reported heterogeneous catalysts. Moreover, the
486 robust stability and efficient applicability in practical wastewater remediation clearly
487 demonstrate its potential to be a candidate in future environmental remediation. The
488 positive correlation between relative contents of high-spin state species (Fe^{II} *S*=2 and
489 Fe^{III} *S*=5/2) and catalytic performance was strongly evidenced by ⁵⁷Fe Mössbauer
490 spectroscopy and DFT calculations, which determined the reactive species generation
491 and electron-transfer pathway in the FeN_x-C/PMS/BPA system. Therefore, a thorough
492 understanding of the spin-state dependent mechanism for PMS activation over single
493 Fe atoms was discerned in this work. These findings highlight in detail the exact
494 identifies of the various iron coordination sites and the specific catalytic pathways of
495 these sites during PMS conversion and compounds decontamination, providing
496 valuable guidance to design strategies toward promoting catalytic capacity in future

497 real-world applications.

498 **Supporting Information**

499 Details of chemicals, synthesis and characterization of catalysts; analytic methods;
500 XRD patterns; TG curve; SEM and TEM images; Raman spectra; N₂ adsorption-
501 desorption isotherm; XPS, XAS and ⁵⁷Fe Mössbauer spectra; DFT calculation result;
502 catalytic performance of pollutants; influence of different parameters; UPLC/MS
503 spectra and BPA degradation pathway; scavenging test result; PMSO consumption and
504 PMSO₂ generation; ERP spectra; cycling experiment; characterization of 101FeN_x-C-
505 600 catalyst; TOC removal efficiency; BPA degradation in actual wastewater;

506 **Acknowledgements**

507 This work was supported by the Tokyo Human Resources Fund for City Diplomacy
508 (H29-1). The authors wish to acknowledge Dr Németh Zoltán, Institute of Wigner
509 Research Centre for Physics, for his help in guiding the analysis of the X-ray absorption
510 spectra (XAS) results of this study.

511 **References**

- 512 1. Rojas, S.; Horcajada, P. Metal–Organic Frameworks for the Removal of Emerging Organic
513 Contaminants in Water. *Chem. Rev.* **2020**, *120*, 8378-8415.
- 514 2. Zhen, Y.; Zhu, S.; Sun, Z.; Tian, Y.; Li, Z.; Yang, C.; Ma, J. Identifying the Persistent
515 Free Radicals (PFRs) Formed as Crucial Metastable Intermediates during Peroxymonosulfate (PMS)
516 Activation by N-Doped Carbonaceous Materials. *Environ. Sci. Technol.* **2021**, *55*, 9293-9304.
- 517 3. Bae, G.; Chung, M. W.; Ji, S. G.; Jaouen, F.; Choi, C. H. pH Effect on the H₂O₂-Induced
518 Deactivation of Fe-N-C Catalysts. *ACS Catal.* **2020**, *10*, 8485-8495.
- 519 4. Lee, J.; von Gunten, U.; Kim, J. H. Persulfate-Based Advanced Oxidation: Critical
520 Assessment of Opportunities and Roadblocks. *Environ. Sci. Technol.* **2020**, *54*, 3064-3081.
- 521 5. Peng, Y.; Tang, H.; Yao, B.; Gao, X.; Yang, X.; Zhou, Y. Activation of
522 peroxymonosulfate (PMS) by spinel ferrite and their composites in degradation of organic pollutants:
523 A Review. *Chem. Eng. J.* **2021**, *414*, 128800.
- 524 6. Wu, L.; Li, B.; Li, Y.; Fan, X.; Zhang, F.; Zhang, G.; Xia, Q.; Peng, W. Preferential
525 Growth of the Cobalt (200) Facet in Co@N–C for Enhanced Performance in a Fenton-like Reaction.
526 *ACS Catal.* **2021**, *11*, 5532-5543.
- 527 7. Tang, Z.; Zhao, P.; Wang, H.; Liu, Y.; Bu, W. Biomedicine Meets Fenton Chemistry.
528 *Chemical Reviews* **2021**, *121*, 1981-2019.
- 529 8. Jiao, L.; Jiang, H.-L. Metal-Organic-Framework-Based Single-Atom Catalysts for Energy
530 Applications. *Chem* **2019**, *5*, 786-804.
- 531 9. Xu, J.; Zheng, X.; Feng, Z.; Lu, Z.; Zhang, Z.; Huang, W.; Li, Y.; Vuckovic, D.;
532 Li, Y.; Dai, S.; Chen, G.; Wang, K.; Wang, H.; Chen, J. K.; Mitch, W.; Cui, Y. Organic
533 wastewater treatment by a single-atom catalyst and electrolytically produced H₂O₂. *Nat Sustain.*
534 **2020**, *4*, 233-241.
- 535 10. Shang, Y.; Xu, X.; Gao, B.; Wang, S.; Duan, X. Single-atom catalysis in advanced
536 oxidation processes for environmental remediation. *Chem. Soc. Rev.* **2021**, *50*, 5281-5322.
- 537 11. Gao, Y.; Zhu, Y.; Li, T.; Chen, Z.; Jiang, Q.; Zhao, Z.; Liang, X.; Hu, C. Unraveling
538 the High-Activity Origin of Single-Atom Iron Catalysts for Organic Pollutant Oxidation via

539 Peroxymonosulfate Activation. *Environ. Sci. Technol.* **2021**, *55*, 8318-8328.

540 12. Qian, K.; Chen, H.; Li, W.; Ao, Z.; Wu, Y. N.; Guan, X. Single-Atom Fe Catalyst
541 Outperforms Its Homogeneous Counterpart for Activating Peroxymonosulfate to Achieve Effective
542 Degradation of Organic Contaminants. *Environ. Sci. Technol.* **2021**, *55*, 7034-7043.

543 13. Shen, G.; Zhang, R.; Pan, L.; Hou, F.; Zhao, Y.; Shen, Z.; Mi, W.; Shi, C.; Wang,
544 Q.; Zhang, X.; Zou, J.-J. Regulating the Spin State of Fe^{III} by Atomically Anchoring on Ultrathin
545 Titanium Dioxide for Efficient Oxygen Evolution Electrocatalysis. *Angew. Chem., Int. Ed.* **2020**, *59*,
546 2313-2317.

547 14. Miao, J.; Zhu, Y.; Lang, J.; Zhang, J.; Cheng, S.; Zhou, B.; Zhang, L.; Alvarez, P.
548 J. J.; Long, M. Spin-State-Dependent Peroxymonosulfate Activation of Single-Atom M–N Moieties
549 via a Radical-Free Pathway. *ACS Catal.* **2021**, *11*, 9569-9577.

550 15. Zhong, W.; Qiu, Y.; Shen, H.; Wang, X.; Yuan, J.; Jia, C.; Bi, S.; Jiang, J. Electronic
551 Spin Moment As a Catalytic Descriptor for Fe Single-Atom Catalysts Supported on C₂N. *J. Am.*
552 *Chem. Soc.* **2021**, *143*, 4405-4413.

553 16. Mineva, T.; Matanovic, I.; Atanassov, P.; Sougrati, M.-T.; Stievano, L.; Clémancey,
554 M.; Kochem, A.; Latour, J.-M.; Jaouen, F. Understanding Active Sites in Pyrolyzed Fe–N–C
555 Catalysts for Fuel Cell Cathodes by Bridging Density Functional Theory Calculations and ⁵⁷Fe
556 Mössbauer Spectroscopy. *ACS Catal.* **2019**, *9*, 9359-9371.

557 17. Peng, L.; Duan, X.; Shang, Y.; Gao, B.; Xu, X. Engineered carbon supported single iron
558 atom sites and iron clusters from Fe-rich Enteromorpha for Fenton-like reactions via nonradical
559 pathways. *Appl. Catal., B* **2021**, *287*, 119963.

560 18. Zhou, Y.; Zhou, L.; Zhou, Y.; Xing, M.; Zhang, J. Z-scheme photo-Fenton system for
561 efficiency synchronous oxidation of organic contaminants and reduction of metal ions. *Appl. Catal.,*
562 *B* **2020**, *279*, 119365.

563 19. Weon, S.; Huang, D.; Rigby, K.; Chu, C.; Wu, X.; Kim, J.-H. Environmental Materials
564 beyond and below the Nanoscale: Single-Atom Catalysts. *ACS ES&T Engg* **2021**, *1*, 157-172.

565 20. Zhang, B.; Zhang, L.; Akiyama, K.; Bingham, P. A.; Zhou, Y.; Kubuki, S. Self-
566 Assembly of Nanosheet-Supported Fe-MOF Heterocrystals as a Reusable Catalyst for Boosting

567 Advanced Oxidation Performance via Radical and Nonradical Pathways.
568 *ACS Appl. Mater. Interfaces* **2021**, *13*, 22694-22707.

569 21. Xing, Y.; Yao, Z.; Li, W.; Wu, W.; Lu, X.; Tian, J.; Li, Z.; Hu, H.; Wu, M. Fe/Fe₃C
570 Boosts H₂O₂ Utilization for Methane Conversion Overwhelming O₂ Generation. *Angew. Chem., Int.*
571 *Ed.* **2021**, *60*, 8889-8895.

572 22. Sahraie, N. R.; Kramm, U. I.; Steinberg, J.; Zhang, Y.; Thomas, A.; Reier, T.;
573 Paraknowitsch, J.-P.; Strasser, P. Quantifying the density and utilization of active sites in non-
574 precious metal oxygen electroreduction catalysts. *Nat. Commun.* **2015**, *6*, 8618.

575 23. Li, J.; Ghoshal, S.; Liang, W.; Sougrati, M.-T.; Jaouen, F.; Halevi, B.; McKinney,
576 S.; McCool, G.; Ma, C.; Yuan, X.; Ma, Z.-F.; Mukerjee, S.; Jia, Q. Structural and
577 mechanistic basis for the high activity of Fe–N–C catalysts toward oxygen reduction. *Energy*
578 *Environ. Sci.* **2016**, *9*, 2418-2432.

579 24. Zhu, Y.; Zhang, B.; Liu, X.; Wang, D.-W.; Su, D. S. Unravelling the Structure of
580 Electrocatalytically Active Fe–N Complexes in Carbon for the Oxygen Reduction Reaction. *Angew.*
581 *Chem., Int. Ed.* **2014**, *53*, 10673-10677.

582 25. Collings, I. E.; Vasiukov, D. M.; McCammon, C. A.; Dubrovinsky, L.; Cerantola, V.;
583 Petitgirard, S.; Hübschle, C. B.; Schönleber, A.; Chernyshov, D.; van Smaalen, S.;
584 Dubrovinskaia, N. Local Structure of Ferrous Iron Formates at Low Temperature and High Pressure
585 Studied by Mössbauer Spectroscopy. *J. Phys. Chem. C* **2019**, *123*, 21676-21684.

586 26. Chen, Z.; Song, J.; Peng, X.; Xi, S.; Liu, J.; Zhou, W.; Li, R.; Ge, R.; Liu, C.;
587 Xu, H.; Zhao, X.; Li, H.; Zhou, X.; Wang, L.; Li, X.; Zhong, L.; Rykov, A. I.; Wang,
588 J.; Koh, M. J.; Loh, K. P. Iron Single Atom Catalyzed Quinoline Synthesis. *Adv. Mater.* **2021**,
589 e2101382.

590 27. Li, H.; Zhao, Z.; Qian, J.; Pan, B. Are Free Radicals the Primary Reactive Species in Co(II)-
591 Mediated Activation of Peroxymonosulfate? New Evidence for the Role of the Co(II)-
592 Peroxymonosulfate Complex. *Environ. Sci. Technol.* **2021**, *55*, 6397-6406.

593 28. Jiang, N.; Xu, H.; Wang, L.; Jiang, J.; Zhang, T. Nonradical Oxidation of Pollutants with
594 Single-Atom-Fe(III)-Activated Persulfate: Fe(V) Being the Possible Intermediate Oxidant. *Environ.*

595 *Sci. Technol.* **2020**, *54*, 14057-14065.

596 29. Liang, J.; Duan, X.; Xu, X.; Chen, K.; Zhang, Y.; Zhao, L.; Qiu, H.; Wang, S.;
597 Cao, X. Persulfate Oxidation of Sulfamethoxazole by Magnetic Iron-Char Composites via
598 Nonradical Pathways: Fe(IV) Versus Surface-Mediated Electron Transfer. *Environ. Sci. Technol.*
599 **2021**, *55*, 10077-10086.

600 30. Dong, H.; Li, Y.; Wang, S.; Liu, W.; Zhou, G.; Xie, Y.; Guan, X. Both Fe(IV) and
601 Radicals Are Active Oxidants in the Fe(II)/Peroxydisulfate Process. *Environ. Sci. Technol. Lett.*
602 **2020**, *7*, 219-224.

603 31. Jin, Q.; Kang, J.; Chen, Q.; Shen, J.; Guo, F.; Chen, Z. Efficiently enhanced Fenton-
604 like reaction via Fe complex immobilized on silica particles for catalytic hydrogen peroxide
605 degradation of 2,4-dichlorophenol. *Appl. Catal., B* **2020**, *268*, 118453.

606 32. Zong, Y.; Guan, X.; Xu, J.; Feng, Y.; Mao, Y.; Xu, L.; Chu, H.; Wu, D., Unraveling
607 the Overlooked Involvement of High-Valent Cobalt-Oxo Species Generated from the Cobalt(II)-
608 Activated Peroxymonosulfate Process. *Environ. Sci. Technol.* **2020**, *54*, 16231-16239.

609 33. Zong, Y.; Zhang, H.; Zhang, X.; Liu, W.; Xu, L.; Wu, D., High-valent cobalt-oxo
610 species triggers hydroxyl radical for collaborative environmental decontamination. *Appl. Catal., B*
611 **2022**, *300*, 120722.

612 34. Zhang, T.; Zhu, H.; Croue, J. P. Production of sulfate radical from peroxymonosulfate
613 induced by a magnetically separable CuFe₂O₄ spinel in water: efficiency, stability, and mechanism.
614 *Environ. Sci. Technol.* **2013**, *47*, 2784-91.

615 35. Kramm, U. I.; Herrmann-Geppert, I.; Behrends, J.; Lips, K.; Fiechter, S.; Bogdanoff, P.
616 On an Easy Way To Prepare Metal-Nitrogen Doped Carbon with Exclusive Presence of MeN₄-type
617 Sites Active for the ORR. *J. Am. Chem. Soc.* **2016**, *138*, 635-40.

618 36. Kramm, U. I.; Herranz, J.; Larouche, N.; Arruda, T. M.; Lefevre, M.; Jaouen, F.;
619 Bogdanoff, P.; Fiechter, S.; Abs-Wurmbach, I.; Mukerjee, S.; Dodelet, J. P. Structure of the
620 catalytic sites in Fe/N/C-catalysts for O₂-reduction in PEM fuel cells. *Phys. Chem. Chem. Phys.*
621 **2012**, *14*, 11673-88.

622 37. Zhao, Y.; Deng, C.; Tang, D.; Ding, L.; Zhang, Y.; Sheng, H.; Ji, H.; Song, W.;

623 Ma, W.; Chen, C.; Zhao, J. α -Fe₂O₃ as a versatile and efficient oxygen atom transfer catalyst in
624 combination with H₂O as the oxygen source. *Nat. Catal.* **2021**, *4*, 684-691.

625 38. Saveleva, V. A.; Ebner, K.; Ni, L.; Smolentsev, G.; Klose, D.; Zitolo, A.; Marelli,
626 E.; Li, J.; Medarde, M.; Safonova, O. V.; Nachtegaal, M.; Jaouen, F.; Kramm, U. I.;
627 Schmidt, T. J.; Herranz, J. Potential-Induced Spin Changes in Fe/N/C Electrocatalysts Assessed by
628 In Situ X-ray Emission Spectroscopy. *Angew. Chem., Int. Ed.* **2021**, *60*, 11707-11712.

629 39. Kramm, U. I.; Herrmann-Geppert, I.; Behrends, J.; Lips, K.; Fiechter, S.; Bogdanoff, P.
630 On an Easy Way To Prepare Metal–Nitrogen Doped Carbon with Exclusive Presence of MeN₄-type
631 Sites Active for the ORR. *J. Am. Chem. Soc.* **2016**, *138*, 635-640.

632 40. Dashdorj, J.; Zvanut, M. E.; Stanley, L. J. Iron-related defect levels in SrTiO₃ measured by
633 photoelectron paramagnetic resonance spectroscopy. *J. Appl. Phys.* **2010**, *107*, 083513.

634 41. Wang, Z.; Jiang, J.; Pang, S.; Zhou, Y.; Guan, C.; Gao, Y.; Li, J.; Yang, Y.; Qiu,
635 W.; Jiang, C. Is Sulfate Radical Really Generated from Peroxydisulfate Activated by Iron(II) for
636 Environmental Decontamination? *Environ. Sci. Technol.* **2018**, *52*, 11276-11284.