

Valorization of Napier grass via intermediate pyrolysis: Optimization using response surface methodology and pyrolysis products characterization

MOHAMMED, Isah Yakub, ABAKR, Yousif Abdalla, YUSUP, Suzana and KABIR, Feroz http://orcid.org/0000-0002-3121-9086>

Available from Sheffield Hallam University Research Archive (SHURA) at:

https://shura.shu.ac.uk/14123/

This document is the Accepted Version [AM]

Citation:

MOHAMMED, Isah Yakub, ABAKR, Yousif Abdalla, YUSUP, Suzana and KABIR, Feroz (2017). Valorization of Napier grass via intermediate pyrolysis: Optimization using response surface methodology and pyrolysis products characterization. Journal of Cleaner Production, 142 (4), 1848-1866. [Article]

Copyright and re-use policy

See http://shura.shu.ac.uk/information.html

1	Valorization of Napier Grass <i>via</i> Intermediate Pyrolysis: Optimization Using Response
2	Surface Methodology and Pyrolysis Products Characterization
3	Isah Yakub Mohammed ^{1,5*} , Yousif Abdalla Abakr ² Suzana Yusup ³ and Feroz Kabir Kazi ⁴
4 5 6	¹ Department of Chemical and Environmental Engineering, the University of Nottingham Malaysia Campus, Jalan Broga, Semenyih 43500, Selangor Darul Eshan, Malaysia
7 8	² Department of Mechanical, Manufacturing and Material Engineering, the University of Nottingham Malaysia Campus, Jalan Broga, Semenyih 43500, Selangor Darul Eshan, Malaysia
9	³ Department of Chemical Engineering, Universiti Teknology Petronas (UTP), Bandar Seri
10	Iskandar, Tronoh 31750, Malaysia
11	⁴ Department of Engineering and Mathematics, Sheffield Hallam University, City Campus,
12	Howard Street, Sheffield S1 1WB, UK
13	⁵ Crops for the Future (CFF), the University of Nottingham Malaysia Campus,
14	Jalan Broga, Semenyih 43500, Selangor Darul Eshan, Malaysia

15 Abstract

16 This study presents first optimization report on pyrolysis oil derived from Napier grass. Effects 17 of temperature, heating rate and nitrogen flow rate on the intermediate pyrolysis of Napier grass biomass in a vertical fixed-bed tubular reactor were investigated collectively. Response surface 18 methodology with central composite design was used for modelling the process and optimization 19 20 of the process variables. Individual second order polynomial model was found to be adequate in 21 predicting bio-oil, bio-char and non-condensable gas yield. The optimum bio-oil yield of 50.57 wt% was recorded at 600 °C, 50 °C/min and 5 L/min nitrogen flow. The bio-oil obtained 22 throughout this study was two-phase liquid, organic and aqueous phase. The bio-oil, bio-char and 23 24 non-condensable gas were characterized using standard analytical techniques. The results 25 revealed that the organic phase consists of hydrocarbons and various benzene derivatives, which 26 can be further processed into fuels and valuable chemicals. The aqueous phase was 27 predominantly water, acids, ketones, aldehydes and some phenolics and other water-soluble organics. The non-condensable gas was made up high hydrogen/carbon monoxide ratio suitable 28

for liquid fuel synthesis via Fischer-Tropsch Synthesis. The bio-char was a porous carbonaceous material with high energy content, which can be applied as a solid fuel, adsorbent or source of biofertilizer. This study demonstrated that Napier grass biomass is a viable feedstock for production of high-value bioenergy precursors.

33 Keywords: Napier grass; pyrolysis; optimization; Response surface methodology; bio-oil
34 characterization

35 *Corresponding author: Isah Yakub Mohammed (<u>kebx3iye@nottingham.edu.my</u>)

36 1. Introduction

Lignocellulosic biomass such as forest residues, agro-wastes, energy grasses, aquatic plants, 37 38 algae, continues to again attention as a suitable alternative energy source. They are non-food 39 materials and consist carbon, which can be converted into high-grade fuel or fuel precursor, 40 biochemicals and other valuable product (Mohammed et al., 2016a; 2016b). However, biofuel 41 from these materials is generating another concern. The issue regarding land use, water and other 42 productive resources for the cultivation of such crops instead of food crops are seen as a potential 43 food scarcity. To address these challenges, production of second generation biofuels should be carefully synchronized to encourage more food production in order to develop a sustainable 44 bioenergy system. 45

Napier grass (*Pennisetum purpureum*) also known as elephant grass, Uganda grass. It is an underutilized herbaceous plant, which can be cultivated up to four times in a year with a ratio of energy output to the energy input of around 25:1 and high biomass yield between 25 and 35 oven dry tons per hectare annually, which corresponds to 100 barrels of oil energy equivalent per hectare (Samson et al., 2005; Flores et al., 2012). Comparing with other energy grasses such as 51 Miscanthus, switchgrass, the biomass yield is between 10 and 20 oven dry tons per hectare per annum and require some nutrient input during cultivation (Richter et al., 2008; Cadoux et al., 52 2012). Flores et al. (2012) reported that Napier grass can be grown without any nutrient or 53 fertilizer input. They evaluated the performance of two Napier grass species with and without 54 55 application of nitrogen fertilizer under the Cerrado climatic condition (semi-humid tropical 56 climate). Their findings revealed that the biomass yield was between 30 to 42 oven dry tons per hectare and showed no response to nitrogen fertilization. Recent study on environmental and 57 economic benefit analysis of Napier grass has been reported by Tsai and Tsai (2016). Their result 58 59 shows that Napier grass can mitigate CO₂ emissions to the order of 5 million Gg per annum. The energy equivalent was found to be 11 million barrels per 100 000 hectare annually which equals 60 to 110 barrels of oil energy equivalent per hectare. This value strongly agrees with the findings 61 reported by Samson et al. (2005) and Flores et al. (2012). Cultivation of Napier grass follows 62 conventional farming practices. It outcompetes weeds and therefore, requires lower 63 establishment costs. This makes it one of the best potential energy crops for the development of 64 efficient and economic bioenergy systems (Samson et al., 2005; Flores et al., 2012). Recently, 65 investigations have also shown that Napier grass can be incorporated into tree plantations such as 66 67 oil palm, rubber tree plantations where there are large areas of unused spaces, estimated at 26.63% of the total space (Zhou et al., 2012; Mohammed et al., 2015a). Mohammed et al. 68 (2015a) reported that intercropping Napier grass with oil palm produce thick elongated Napier 69 70 grass stem, which contained more biomass than the Napier grass grown under unshaded conditions. This therefore offers another economic benefit. 71

Pyrolysis is the thermal decomposition of biomass in total absence of oxygen to produce solid,
liquid and gas products. This process has high improved efficiency, environmental suitability and

74 flexibility as virtually any biomass type can be handled to generate various products (Heidenreich and Foscolo, 2015; Mohammed et al., 2016c). In addition, high liquid yield (known 75 76 as bio-oil) which is of most interest can be obtained through pyrolysis process under a careful control of process variables such as inert gas flowrate, pyrolysis temperature, heating rate, vapor 77 residence time, reactor type, and the temperature regime between the reaction and cooling zone 78 79 (Bridgwater, 2012; Eom et al., 2012). Generally, pyrolysis process is classified into slow, intermediate and fast pyrolysis. In slow pyrolysis (carbonization), temperature up to 400 °C, 80 81 vapor residence time of 60 minutes to days are used and a typical product yield of about 35 % 82 bio-char, 30 % bio-oil, and 35 % non-condensable gas. For fast pyrolysis, temperature of about 500 °C, high heating rate (up 1000 °C/min) and short vapor residence time of approximately 1 s 83 are applied. Nearly 80 % bio-oil yield (wet basis), and about 13 and 12 % non-condensable gas 84 and bio-char are obtained. (Bridgwater, 2012; Mohammed et al., 2015b). On the other hand, 85 intermediate pyrolysis produces about 40-60 % bio-oil, 15-25 % bio-char and 20-30 % non-86 condensable gas under process condition of 500-650 °C and 10 to 30 s vapor residence time. 87 Intermediate pyrolysis has added advantages over the fast pyrolysis in that the oil from it is less 88 reactive and may be used directly as fuel in engines and boilers. It also produces dehydrated 89 90 char, which can be used for both agricultural and energy applications (Tripathi et al., 2016; Mohammed et al., 2016c). However, production of bio-oil from biomass via pyrolysis requires 91 92 energy at different stages. Energy is consumed during biomass collection, bio-oil production 93 stage (size reduction, feeding, pyrolysis and cooling) and bio-oil transportation. Study has shown 94 that about 82 % of the total energy requirement for a bio-oil production process is consumed at 95 the production stage mainly during size reduction and pyrolysis (high temperature is needed to 96 decompose the biomass) (Ning et al., 2013). Though, the challenge of high energy requirement 97 may be compensated through efficient utilization of other pyrolysis products (bio-char and non-98 condensable gas). Climate change mitigation is an unresolved issue between the cons and pros of 99 biofuels utilization. Production of biofuel can be said to reduce greenhouse gases only when it 100 produce residuals that render the use of fossil fuels unnecessary. Consequently, production of 101 second generation biofuels should be carefully synchronized to encourage more agricultural food 102 production and the use of marginal lands for cultivation of energy crops in order to develop a 103 sustainable bioenergy system (Caputo, 2014).

Utilization of Napier grass for the development of biofuels via pyrolysis is very limited. Strezov 104 105 et al. (2008) reported thermochemical conversion of Napier grass to bio-oil, bio-char and non-106 condensable gas using a pyro-probe type reactor. They investigated effect of heating rate (10 and 50 °C/min) on the pyrolysis product distribution. In each case, the system was ramped to a 107 maximum temperature of 900 °C. Their result showed that the bio-oil collected was made up of 108 109 organic acids, phthalate esters, and larger fraction of various benzene chemicals, phenols and pyrans. They reported that the high heating rate of 50 °C/min promoted the formation of smaller 110 acids and benzene fractions in the bio-oil compared to the low heating rate of 10 $^{\circ}$ C/min. CO₂ 111 was found to be the major component in the non-condensable gas. They stated that the bio-oil 112 collected was a two-phase liquid and sample was analyzed with GC-MS without separation into 113 individual phases. Although, a mixture of methanol and chloroform was used as solvent but there 114 is high tendency of phase separation during the GC-MS analysis and the sample injected may not 115 be the true representation of the whole bio-oil. Lee et al. (2010) reported pyrolysis of Napier 116 117 grass in an induction-heating reactor. They investigated effect of heating rate between 50 and 118 200 °C/min and biomass particle size within 2 mm at pyrolysis temperature of 500 °C. Their findings also showed that bio-oil yield increased with heating rate from 50 to 150 °C/min but 119

120 declined thereafter. Maximum bio-oil yield of 36 wt% was reported at 150 °C/min with 0.224 121 mm biomass particle size. The bio-oil yield was lower than those generally obtained from a fast pyrolysis system (50-60 wt%) and was attributed to the low level of heating rate used compared 122 123 to the fast pyrolysis where up to 1000 °C/min is normally employed. This conclusion may not be valid as the study was conducted under a constant temperature of 500 °C, which may not be the 124 125 optimum temperature for complete devolatilization of various structural components in the biomass. Pyrolysis product distribution is strongly related to the pyrolysis temperature and vapor 126 residence time in the reactor compared to the heating rate (Bridgwater, 2012; Mohammed et al., 127 128 2015b). Their findings also reveal that the bio-oil was predominantly acetic acid, phenols, ketones, and nitrogenated hydrocarbons. These chemical species are polar compounds that are 129 generally present in the aqueous phase of the bio-oil. The major composition of the non-130 131 condensed gas reported were furans, olefins and mono-aromatic hydrocarbons (benzene, toluene and xylene), which are normal component of the condensate. This means that the cooling system 132 used during the study was inefficient. Thus, the composition of the bio-oil reported does not 133 represent the whole condensables from the Napier grass. 134

A study recently conducted by Sousa et al. (2016) on pyrolysis of Napier grass in a fluidized bed 135 reactor with a feed rate between 20 and 35 kg/h, temperature at around 500 °C. Maximum oil 136 yield of 28.2 % was recorded. This yield was also far lower relative to a typical bio-oil yield (40-137 60 %) from a fluidized bed pyrolysis system. The authors attributed it to the accumulation of 138 heavier fractions within the system before the cooling zone. The bio-oil composition reported 139 140 had principally phenolics, organic acids and traces of levoglucosan and mono-aromatic 141 hydrocarbons. H₂, CO, CO₂, CH₄ and C₂H₆ were the main component detected in the noncondensable gas. However, the yield and composition of the bio-oil collected may have been 142

143 affected as the authors stated that air was employed as the fluidizing gas instead of an inert gas. Effect of other temperature levels and variations in the fluidizing gas velocity were not 144 investigated. Similarly, De Conto et al. (2016) studied pyrolysis of Napier grass in a rotary kiln 145 146 reactor where the effect of pyrolysis temperature and rotating speed of the reactor were investigated. Information regarding the bio-oil collected by the authors was only limited to the 147 148 yield. Physicochemical properties and chemical compositions of the bio-oil were not determined. The non-condensable gas analyzed particularly at 600 °C was made up of high hydrogen 149 (H_2) /carbon monoxide (CO) ratio which was attributed to the temperature only. However, this 150 151 may be a combined effect of temperature and vapor residence time in the reactor. From the 152 reactor size and carrier gas flowrate reported by the authors, vapor residence was around 1.425 min (85 sec), which is long enough to result to severe secondary cracking of the pyrolysis vapor 153 154 at such a high temperature. Some classical studies on the pyrolysis of Napier grass have also been reported by our research group recently (Mohammed et al., 2015b; Lim et al., 2015, 155 Mohammed et al., 2016a, 2016b, 2016d, Lim et al., 2016). To date, no study on pyrolysis of 156 157 Napier grass is available in the literature that deals with collective examination of pyrolysis 158 temperature, heating rate, inert gas flowrate on the products distribution and composition. This 159 investigation is needed in order to appropriately evaluate the energy potential of the biomass.

160 The objective of this study was to evaluate collective effects of heating rate, pyrolysis 161 temperature and nitrogen flow rate on the pyrolysis products distribution and optimize bio-oil 162 yield from intermediate pyrolysis of Napier grass using response surface methodology (RSM) 163 based on a central composite design (CCD).

164 2. Materials and Methods

165 2.1 Material and characterization

Fresh locally grown Napier grass stem (NGS) was collected from the Crops for the Future (CFF) 166 Field Research Centre. The material was oven dried at 105 °C for moisture content determination 167 and thereafter shredded in a Retsch[®] rotor beater mill. The ground biomass sample (0.2 - 2.5 mm) 168 was collected and preserved in airtight plastic bags for further analysis. Proximate and ultimate 169 170 analyses on dry basis were carried out according to the relevant standard procedures. Major inorganic elements in the ash were determined using atomic absorption spectrometer (AAS) 171 Perkin Elmer analyst 400 (Perkin Elmer Sdn Bhd, Selangor, Malaysia). Higher heating value 172 (HHV) was determined using a Parr 6100 oxygen bomb calorimeter. Elemental compositions 173 were determined using a Perkin Elmer 2400 Series II CHNS/O analyzer (Perkin Elmer Sdn Bhd, 174 Selangor, Malaysia). The structural analysis of the biomass was performed using High-175 performance liquid chromatography (HPLC-1260 infinity, Agilent Technologies Sdn Bhd, 176 Selangor, Malaysia). Thermogravimetric analysis was conducted with the thermal analyzer 177 178 (STA) 6000 (TGA) (Perkin Elmer Sdn Bhd, Selangor, Malaysia) in a nitrogen atmosphere, flow rate 20 mL/min at temperatures between 30-950 °C and heating rate of 10 °C/min. About 10.0 179 mg (particle size of 0.2 mm) of sample was used. 180

181 2.2 Pyrolysis experiment

Intermediate pyrolysis study was carried in a vertical tube fixed bed pyrolysis system as shown in Figure 1. The system consists of a fixed bed reactor made of stainless steel (115 cm long, 6 cm inner diameter), a distribution plate with 1.5 mm hole diameter which sits at 25 cm from the bottom of the tube, two nitrogen preheating sections, a cyclone, water chiller operating at 3 °C 186 attached to a coil condenser, oil collector and gas scrubbers. 200 g of NGS (bone dry, 2.5 mm 187 particle size) was placed on the distribution plate inside the reactor tube and pyrolysis was conducted under nitrogen atmosphere. The reaction temperature was monitored with a K-type 188 189 thermocouple connected to a computer through data logger. The reaction time was kept at 60 min after the temperature reaches the desired value. The pyrolysis vapor was condensed by passing 190 through a condenser connected to a chiller at 3 °C and the oil was collected. The bio-oil, bio-char 191 and non-condensable gas yield was computed according to equation (1), (2) and (3) respectively. 192 193 20 sets of pyrolysis experiments were conducted based on the number generated by the statistical software where the pyrolysis temperature (450-750 °C), nitrogen flow rate (5-25 L/min) and 194 heating rate (10-50 °C/min) were selected in this study. 195

196
$$\text{Yield}\left[\text{wt\%}\right]_{\text{bio - oil}} = \left[\frac{\text{weight of the bio - oil collected}}{\text{weight of biomass feed}} \times 100\right] - \dots - \dots - (1)$$

197 Yield
$$[wt\%]_{bio - char} = \left[\frac{weight of the bio - char collected}{weight of biomass feed} \times 100\right] - - - - (2)$$

198 Yield
$$[wt\%]_{non - condensable gas} = 100 - (Yield $[wt\%]_{bio - oil} + Yield [wt\%]_{bio - char})^{----}$ (3)
199$$

200



202 pyrolysis section, (4) furnace controller, (5) heater, (6) insulator, (7) thermocouples, (8) data

logger, (9) computer, (10) water chiller, (11) cyclone, (12) condenser, (13) bio-oil collector, (14)

204 gas scrubber, (15) gas sampling bag, (16) gas venting

205

The response surface methodology (RSM) is a statistical and mathematical technique for 207 208 designing experiments to provide reliable measurements of the desired response which may be affected by many variables. A mathematical model with the best fit is usually developed using 209 210 data from the experimental design and the optimal value of the variables that produces maximum or minimum response are determined. In this study, RSM was used following the face central 211 212 composite design (FCCD) method with the aid of Design expert software (Version 6.0.6, Stat-Ease Inc., MN, USA). The CCD consists of 8 cube (factorial) points, 6 axial points and 6 213 replicates at centre points which translate to a number of experiments and 20 experiments were 214 generated according to equation (4). The variables and the experimental domain in this design 215 216 are presented in Table 1. At the end of the experiment, the results were fitted to second-degree 217 polynomial design as shown in the equation (5) below.

218
$$N = 2^{n} + 2n + n_{c} = 2^{3} + 2 \times 3 + 6 = 20 - - - - - - - (4)$$

219 Where N is the number of experiments, n is the number of factors and n_c is the number of 220 replicates at the centre point.

$$Y = \beta_0 + \sum_{i=1}^n \beta_i X_i + \sum_{i=1}^n \beta_{ii} X_i^2 + \sum_{i=1}^n \sum_{j>1}^n \beta_{ij} X_i X_j - \dots - \dots - \sum_{j=1}^{221} (5)$$

222

Where Y is the predicted response, n is the number of experiments, β_0 , β_i , β_{ii} and β_{ij} is regression coefficients for the constant, linear, quadratic and interaction terms, respectively. X_i and X_j are the coded independent factors. The adequacy of the final model was tested using both graphical and numerical analysis and the experimental data were analysed statistically using analysis ofvariance (ANOVA).

228

Table 1: Range of independent variables and experimental levels

229

230 2.4 Pyrolysis products characterizations

Pyrolysis oil collected at optimized condition was further analyzed. Physicochemical properties 231 of the bio-oil such as pH, water content, calorific value, density and viscosity were analyzed with 232 233 a WalkLAB microcomputer pH meter (TI9000, Trans Instruments, Singapore), Karl Fischer V20 volumetric titrator (Mettler Toledo, Columbus, OH, USA), oxygen bomb calorimeter (Parr 6100, 234 Parr Instruments, Molin, IL, USA), Anton Paar density meter (DMA 4500 M, Ashland, VA, 235 USA) and Brookfield viscometer (DV-E, Hamilton, NJ, USA) respectively. Carbon, hydrogen, 236 nitrogen, sulfur and oxygen contents of the bio-oil were determined with a CHNS/O Analyzer 237 238 (2400 Series II CHNS/O analyzer, Perkin Elmer Sdn Bhd, Selangor, Malaysia). Functional group analysis was performed with fourier transform infrared-FTIR (Spectrum RXI, PerkinElmer, 239 240 Selangor, Malaysia) using a pair of circular demountable potassium bromide (KBr) cell windows 241 (25mm diameter and 4mm thickness). Spectra were recorded with Spectrum V5.3.1 software within wave number range of 400-4000 cm⁻¹ at 32 scans and a resolution of 4 cm⁻¹. Fractionation 242 of the organic phase bio-oil was simulated using TGA in nitrogen atmosphere at 20mL/min, 243 10° C/min from ambient to 500°C to examine the volatile fractions and the result was compared with 244 the simulated distillation of fossil-gasoline, kerosene and diesel. Detail of the chemical composition 245 of the bio-oil was determined using a gas chromatograph-mass spectrometer (GC-MS) 246 (PerkinElmer Clarus[®] SQ 8, Akron, OH, USA) with a quadruple detector and column (30m x 247 0.25mm x 0.25µm) (PerkinElmer-EliteTM-5ms, Akron, OH, USA). The oven was programmed at 248 an initial temperature of 40 °C, ramp at 5 °C /min to 280 °C and held there for 20 min. The 249

injection temperature, volume, and split ratio were 250 °C, 1 µl, and 50:1 respectively. Helium
was used as carrier gas at a flow rate of 1 mL/min. The peaks of the chromatogram were
identified by comparing with standard spectra of compounds in the National Institute of
Standards and Technology library (NIST, Gaithersburg, MD, USA).

254 Samples of the non-condensable gas were collected in SKC polypropylene fitted gas sampling 255 bag and analyzed offline with a gas chromatography (PerkinElmer Clarus 500, Akron, OH, USA) equipped with stainless steel column (Porapak R 80/100) and thermal conductivity 256 detector (TCD). Helium was used as a carrier gas and the GC was programed at 60 °C, 80 °C and 257 258 200 °C for oven, injector and TCD temperature respectively. Bio-char proximate and ultimate analyses were performed following the same analytical procedure adopted for the feedstock 259 260 characterization above. Scanning electron microscopy (SEM, FEI Quanta 400 FE-SEM, Hillsboro, OR, USA) and physisorption analyzer (ASAP 2020 Micrometrics, Norcross, GA, 261 262 USA) were used to evaluate the surface and structural characteristics, and specific surface area (BET) and pore properties of the bio-char respectively. 263

264 **3**.

Results and Discussion

265 *3.1 Feed stock characteristics*

Characteristic of NGS used in this study is summarized in Table 2. The proximate analysis result obtained showed significant difference relative to similar properties of Napier grass (NG) reported in the literature. Higher volatile matter and lower ash contents were recorded compared to the values reported by Strezov et al. (2008), Lee et al. (2010), Braga et al. (2014) and De Conto et al. (2016). Higher heating value (HHV) was 18.05 MJ/kg relative to 15.61 MJ/kg and 15.77 MJ/kg reported by Braga et al. (2014) and De Conto et al. (2016) respectively. These variations in the proximate analysis result is attributed to the post-harvest treatment of the NG 273 sample used. The results of ultimate and structural analyses (Table 2) showed good agreement 274 with the literature values. Comparing the structural characteristics of NGS with that of switchgrass and miscanthus, from the work of Imam and Capareda (2012) and Rena et al. (2016), 275 276 switchgrass has lower cellulose (32-34 wt%) and lignin (18.8 wt%) contents. Similarly, 277 miscanthus has lower lignin content (12-12.58 wt%) but higher cellulose (50.34-52.13 wt%) and hemicellulose (24.83-25.76 wt%) contents as reported by Brosse et al. (2012) and Shemfe et al. 278 (2016). Studies on the detail mineral composition of NG biomass is rarely reported. This 279 characteristic is very important for biomass thermochemical conversion. For pyrolysis, it has 280 281 been reported that mineral composition of biomass has great influence on both product yield and 282 bio-oil composition (Mohammed et al., 2016b). Most researchers employed x-ray fluorescence (XRF), energy dispersive x-ray (EDX) in determining the mineral composition of biomass. 283 284 These techniques provide elemental composition at a specific point within the sample instead of the mineral distribution in the whole sample. Strezov et al. (2008) reported silicon (Si), 285 potassium (K), magnesium (Mg), calcium (Ca), iron (Fe), aluminum (Al) and sodium (Na) using 286 287 XRF as the major elements present in the Napier grass biomass. In this study, the major elements recorded in the feedstock using atomic absorption spectrometer (AAS) showed similar 288 289 mineralogical composition (Na, K, Ca, Al, Fe and Si).

290

Table 2: Characteristics of Napier grass biomass used

292

291

The thermogravimetric profile of NGS is presented in Figure 2. The sample was heated in nitrogen atmosphere (20 mL/min) at 10 °C/min from ambient temperature to 100 °C and held there for 30min to eliminate all the physically absorbed moisture and thereafter, ramped to 900

^oC. Pyrolysis of a lignocellulosic biomass proceeds through simultaneous decomposition of 296 297 hemicellulose, cellulose and lignin of the biomass with each component displaying separate mechanism, which makes the process so complex (Kan et al., 2016; Anca-Couce, 2016). Most 298 299 studies have focused on the classical evaluation of degradation of the individual components as a foundation of the likely decomposition pathway. However, research have shown that there is 300 301 interaction between the structural components of the biomass during pyrolysis, which is also affected by the mineral content of the biomass (Kan et al., 2016; Anca-Couce, 2016). In this 302 study, pyrolysis profile of NGS from the TG/DTG showed four distinct regions at temperature 303 around 200 °C, 229-285 °C, 326 °C and 373-540 °C. Visible peak at about 200 °C was observed, 304 which is due to the decomposition of extractives. This could also be due further dehydration of 305 the biomass through cleavage of hydroxyl groups in the hemicellulose, cellulose and lignin (Van 306 307 de Velden et al., 2010; Collard and Blin, 2014). The total weight loss recorded under this region was 12.58 wt%. A shoulder is observed at temperature between 229 and 285 °C and is ascribed 308 to the fast decomposition of hemicellulose as a result of cleavage of the glyosidic bond between 309 310 the monomeric units. Depolymerization of cellulose to active cellulose is also expected within this temperature range (Van de Velden et al., 2010; Collard and Blin, 2014). The corresponding 311 weight loss of 13.66 wt% was recorded. The main characteristic peak was observed at 326 °C. 312 This is attributed to cellulose decomposition and had a maximum weight loss of 35.74 wt%. 313 Temperature between 373 and 540 °C represent degradation of lignin due to carbon-carbon 314 scission and disruption of aromatic rings within the lignin structure, which resulted in weight 315 loss of 10.78 wt%. Temperature beyond 540 °C represent transformation of the remaining solid 316 (char) through aromatization and demethylation (Patwardhan et al., 2011; Collard and Blin, 317 318 2014). Similar decomposition profile of NG has also been reported by Braga et al. (2014). They

ascribed peaks observed at temperature range of 180-300 °C, 300-380 °C and 380-530 °C to the decomposition of hemicellulose, cellulose and lignin respectively. Similarly, De Conto et al. (2016) studied thermogravimetrics of NG and reported decomposition of cellulose at 318 °C and lignin with 350-500 °C. These decomposition temperature ranges has also been reported in the literature for other lignocellulosic biomass (Gómez et al., 2016)

324

Figure 2: Thermogravimetric profile of Napier grass stem (TG and DTG)

325

326 *3.2 Central composite design (CCD) and statistical analysis*

Experiments were conducted according to the experimental designed matrix and the responses (bio-oil, bio-char and non-condensable gas yields) are presented in Table 3. The experimental results were fitted to second-degree polynomial model as shown in the (6), (7) and (8) below. Where A, B and C is the temperature (°C), nitrogen flow rate (L/min) and heating rate (°C/min). Analysis of variance (ANOVA) for each response model was carried to establish its significance and nature of interactions between the process variables (Table 4a-c).

333
$$Y_{\text{bio-oil}} = 49.37 + 2.03A - 0.70B + 2.04C - 13.31A^2 + 0.75B^2 - 0.83C^2 - 0.67AB - 1.88AC + 0.09BC - ----(6)$$

334
$$Y_{bio-char} = 21.78 - 11.89.A + 0.66B - 0.73C + 10.32A^2 - 0.14B^2 - 0.72C^2 - 0.61AB + 0.27AC - 0.52BC - (7)$$

335
$$Y_{\text{Non-condensable-gas}} = 28.85 + 9.85.A + 0.04B - 1.31C + 2.99A^2 - 0.61B^2 + 1.55C^2 + 1.28AB + 1.61AC + 0.43BC - -- (8)$$

336

Table 3: CCD Experimental Design Matrix and Response

338

339 For the bio-oil yield, the Model F-value of 171.60 (Table 4a) implies the model is significant. 340 Lack of Fit F-value of 3.18 implies that it is not significant relative to the pure error. Nonsignificant lack of fit is desirable because it is wanted to fit. For the bio-char and non-341 342 condensable gas yield, similar trends are observed. Model F-value of 151.22 (Table 4b) and 69.74 (Table 4c) with the corresponding lack of fit F-value of 1.47 and 3.41 are recorded. 343 Values of "Prob > F" less than 0.0500 indicates model terms are significant while values greater 344 0.1000 indicate significant. than the model not 345 terms are In the case of bio-oil, the significant model terms are A, B, C, A², AB, AC, with the 346 corresponding F-value of 62.92, 7.46, 63.59, 741.20, 5.52, and 43.12. It can also be observed 347 that the most significant model term on the bio-oil yield obeyed the following order 348 $A^2 > A > C > AC > B > AB$ while the quadratic terms B^2 , C^2 and the interaction BC does not have a 349 significant impact on the bio-oil yield. For bio-char, the significant model terms are A and A^2 350 351 with the linear term having the largest significance due higher F-value of 1006.27 (Table 4b). The significant model terms for the non-condensable gas are A, C, A^2 , AB, AC, with A being the 352 most significant term having F-value of 557.41 (Table 4c). The coefficient of determination, R^2 353 for the bio-oil, bio-char and non-condensable gas model is 0.9936, 0.9927 and 0.9843 354 respectively (Table 4a-c). Another regression parameter considered is the adjusted R2, which 355 improves the coefficient of determination (R^2) in relation to the sample size and the model terms, 356 and the corresponding value is 0.9878, 0.9861 and 0.8903. The R^2 and adjusted R^2 values for the 357 individual model are high enough and comparable, which indicate that the selected quadratic 358 359 response surface model for the pyrolysis products sufficiently describe the experimental data within the selected operating conditions. Also, the predicted R^2 value of the bio-oil, bio-char and 360 the non-condensable gas model is 0.9518, 0.9411 and 0.8903 respectively which is in good 361

agreement with the respective adjusted R^2 values. Adequate precision (Adeq Precision) is a 362 measure of signal to noise ratio and a ratio greater than 4 is desirable. In this study, the value of 363 Adeq precision is 36.4925, 32.8072 and 27.3167 for the bio-oil, bio-char and non-condensable 364 gas model respectively. These high values indicate adequate signal and the models can be used to 365 navigate the design space. Furthermore, the coefficient of variation (CV) which is a measure of 366 the reliability of the experiment, it expresses the overall experimental error as a percentage of the 367 overall mean. The CV value recorded for the pyrolysis products is less than 4.5, hence, this 368 experiment can be said to be highly reliable since the lower the CV value, the higher is the 369 370 reliability of the experiment.

371

Table 4: ANOVA test for (a) bio-oil, (b) bio-char, (c) non-condensable gas response model and respective model term

374

375 *3.3 Model validation and response surface plot*

To further evaluate the adequacy of the quadratic models in fitting the experimental data, 376 diagnostic plots (normal % probability against studentized and outlier T against a number of 377 runs) were established and are presented in Figure 3 (a and b). In the Figure 3a, the points are 378 379 distributed approximately on a straight line along the diagonal for each of the pyrolysis product 380 models. This trend depicts that the error terms are normally distributed and independent of each. 381 Furthermore, from Figure 3b, the points are randomly distributed around zero on the outlier T 382 axis and between +3.5 and -3.5 in all the cases, which connote homoscedasticity and therefore 383 suggest that the respective model is suitable and successfully establish the relationship between the pyrolysis process variables studied and the product distribution. 384

385 386 Figure 3: Diagnostics of models (a) Normal % probability versus studentized residuals (b) Outliers T versus run number

387

388 In order to examine the effect of the pyrolysis process variables on the response values, interaction and 3D surface plots were used. Although, it is not possible to present the effects of 389 all the parameters studied on the same 3D, as such, the response surface plots are presented by 390 391 varying two factors and keeping one factor constant as shown in Figure 4(a-f), 5(a-f) and 6(a-f) 392 for bio-oil, bio-char and non-condensable gas yields respectively. The interaction between the nitrogen flow and temperature at a constant heating rate (30 °C/min) on the bio-oil yield (Figure 393 4a-b) shows that increase in the nitrogen flow rate from 5 L/min to 25 L/min and temperature 394 from 450°C to 600 °C increased the bio-oil yield. The oil yield became maximum at 600 °C for 395 396 both nitrogen levels but at a different amount. Lower nitrogen level (5 L/min) produced 52.06 wt% oil compared to 49.20 wt% oil yield at 25 L/min nitrogen flow at the same 600 °C. The 397 decreased oil yield at the higher nitrogen flow rate could be as a result of more uncondensed 398 399 volatiles leaving as part of non-condensable gas due to short vapor residence time in the condenser. Declines in the oil yields were observed at temperature above 600 °C. The bio-oil 400 yield trend recorded with temperature from 450 °C to 600 °C and above 600 °C can be 401 respectively ascribed to degradation of more lignin, and secondary reactions of pyrolysis vapor 402 403 and more decomposition of bio-char at the elevated temperature (Soetardji et al., 2014). The effect of heating rate and temperature on the bio-oil yield at a constant nitrogen flow rate (15 404 L/min) is presented in Figure 4c-d. Increasing temperature from 450 to 600 °C and heating rate 405 between 10 and 50 °C/min increased the oil yield from 29.26 to 46.13 wt% and 37.12 to 50.89 406 wt% respectively. The yield of more bio-oil at 50 °C/min relative to 10 °C/min under the same 407 pyrolysis temperature is attributed to rapid depolymerization of the biomass to primary volatiles 408

409 at the higher heating rate. (Yorgun and Yildz, 2015). The oil yield decreased to 37.10 wt% at both heating rates under the same temperature (750 °C) and nitrogen flow (15 L/min) due to 410 secondary reactions at such a high temperature. The combined effects of heating rate and 411 nitrogen flow rate at a pyrolysis temperature of 600 °C on the bio-oil yield is shown in Figure 4e-412 f. As the nitrogen flow increased from 5 L/min to 25 L/min, bio-oil yield at 50 °C/min heating 413 rate remained higher compared to the oil collected at 10 °C/min but no significant impact of 414 nitrogen flow was observed in both cases. Bio-oil yield at 50 °C/min was between 50.72 and 415 51.94 wt% relative to 46.45-48.03 wt% oil recorded at 10 °C/min. 416

417 Figure 4: Interaction graphs and the corresponding surface response plots for the combined
418 effects of process variables on bio-oil yield. (a) and (b) effect of nitrogen flow and
419 temperature at 30°C/min; (c) and (d) effect of heating rate and temperature at
420 15L/min N₂; (e) and (f) effect heating rate and nitrogen flow at 600°C

421

422 Impacts of pyrolysis temperature, heating rate and nitrogen flow rate on the yield of bio-char are 423 shown in Figure 5 (a-f). Interactions between temperature and nitrogen flow Figure 5 (a-b) 424 showed that the temperature had great influence on the bio-char yield compared to the nitrogen 425 flow rate. A general decline in the bio-char yield was recorded with increasing temperature at 426 both 5 L/min and 25 L/min nitrogen flow rates, which is attributed to devolatilization more organic materials as dehydration of hydroxyl groups and decomposition of lignocellulose 427 structure progresses with increasing temperature (Mohammed et al., 2015b). This clearly 428 demonstrated that bio-char yield is governed by the pyrolysis temperature. The impact of 429 430 temperature also dominated interaction between temperature and heating rates on the bio-char yield (Figure 5c-d). There was no considerable difference between the bio-char yields at 10 431 °C/min and 50 °C/min throughout the investigated temperature range. Bio-char yields recorded 432

were 44.27 wt% and 42.27 wt% at 450 °C, and 20.59 wt% and 19.03 wt% at 750 °C for 10 433 434 °C/min and 50 °C/min heating rate respectively. The slightly lower value of bio-char recorded at 50 °C/min could also be attributed to rapid degradation of biomass. Influence of heating rate and 435 nitrogen flow on the bio-char yield is presented in Figure 5(e-f). Both factors did not 436 significantly impacted on the char yield. As nitrogen flow increased from 5 L/min to 25 L/min, 437 bio-char yield at 50 °C/min remained the same (around 20 wt %) while between 20.48 to 22.84 438 wt% char yield was recorded at 10 °C/min under the same nitrogen flow regime. These 439 observations indicated that the range of nitrogen flow rates used in this study was sufficient in 440 441 preventing secondary reaction such as condensation which normally favored char formation (Yorgun and Yildz, 2015). 442

Figure 5: Interaction graphs and the corresponding surface response plots for the combined effects of process variables on bio-char yield. (a) and (b) effect of nitrogen flow and temperature at 30°C/min; (c) and (d) effect of heating rate and temperature at 15L/min N₂; (e) and (f) effect heating rate and nitrogen flow at 600°C

447

Effects of process variables on the production of non-condensable gas are shown in Figure 6 (a-448 449 f). The interaction between temperature and nitrogen flow (Figure 6a-b) revealed that noncondensable gas yield increased with temperature at both nitrogen flow levels. As the 450 temperature progressed from 450 to 600 °C, the increase in the gas yield was not significant 451 under 5 L/min nitrogen flow. This indicates the most of the volatiles generated during the 452 pyrolysis are effectively captured in the condenser, which eventually ended up in the bio-oil as 453 rightly identified in the bio-oil yield section above. The significant change in the gas yield 454 recorded under 25 L/min nitrogen flow between 450 and 600 °C is mainly due to shorter vapor 455 residence time in the condenser compared to the 5 L/min nitrogen flow rate. With increasing 456

temperature above 600 °C, the gas yield increased rapidly in both nitrogen flow regimes, which 457 can be ascribed to further thermal decomposition of pyrolysis vapor in addition to some 458 uncondensed volatiles leaving the pyrolysis system as part of non-condensable gas, particularly 459 at the higher flow rate. 30.21-42.40 wt% gas yield was recorded with 25 L/min nitrogen flow 460 relative to 26.05-39.77 wt% gas yield using 5 L/min. Furthermore, the combined effects of 461 temperature and heating rate on the yield of non-condensable gas (Figure 6c-d) also revealed that 462 the temperature has great influence on the gas yield. With increasing temperature from 450-750 463 °C, the non-condensable gas yield increased from 26.46 to 42.94 wt% and 20.62 to 43.55 wt% at 464 10 °C/min and 50 °C/min respectively. This trend is believed to be as a result of secondary 465 cracking of volatiles and further decomposition of bio-char at higher temperatures (Yorgun and 466 Yildz, 2015). On the other hand, nitrogen flow and the heating rate seem to have no considerable 467 impact on the yield of non-condensable gas (Figure 6e-f). As the nitrogen flow rate increased 468 from 5L/min to 25L/min, the change in the yield of non-condensable gas was insignificant at 469 both 10°C/min and 50°C/min. Gas yield of 28.02-28.95 wt% and 30.71-31.50 wt% were recorded 470 with 10 and 50° C/min heating rate respectively throughout the carrier gas flow range used in this 471 study. 472

473

474 Figure 6: Interaction graphs and the corresponding surface response plots for the combined
475 effects of process variables on non-condensable gas yield (a) and (b) effect of
476 nitrogen flow and temperature at 30°C/min; (c) and (d) effect of heating rate and
477 temperature at 15L/min N₂; (e) and (f) effect heating rate and nitrogen flow at 600°C

478

479 *3.4 Optimization and validation of result*

In order to maximize the bio-oil yield, the process variables considered in this study wereoptimized. Pyrolysis temperature, heating rate, and nitrogen flow rate were maintained within the

482 range of experimental condition studied while the bio-char and non-condensable gas responses were minimized (Table 5a). Based on these conditions, eight (8) solutions were generated by the 483 software as presented in Table 5(b). Solution number one (1) with the highest desirability was 484 chosen. This indicates that the pyrolysis temperature of around 600 °C, 5 L/min nitrogen flow 485 rate, and 50 °C/min heating rate was sufficient to produce optimum bio-oil of 51.94 wt%. This 486 487 result was validated by performing further experiments under the optimized process condition. The experiment was repeated in triplicates and the results are shown in Table 6. The average 488 value of bio-oil yield recorded (50.57 wt%) sufficiently agree with the predicted value. 489 490 Consequently, this validation confirms the adequacy of the developed quadric model for bio-oil vield. 491 492

 Table 5: (a) Optimization condition (constraints) (b): Optimized solutions

 493

494

495 **Table 6**: Bio-oil yield predicated at the optimized condition and experimental value

496

The bio-oil yield from NGS observed at the optimized condition (600 °C, 5 L/min N₂ and 50 497 ^oC/min) in this study is similar to the liquid product reported by Strezov et al. (2008) during the 498 499 pyrolysis of NG. They observed 54.37 wt% bio-oil yield at 50 °C/min and 500 °C pyrolysis 500 temperature under argon atmosphere. Similarly, the result of pyrolysis study of NG by De Conto et al. (2016) showed a total bio-oil yield of 52.99 wt% at 700 °C, 25 °C/min and 1 L/min nitrogen 501 502 flow. Sousa et al. (2016) also reported a total liquid product of 48.2 wt% from the pyrolysis of NGS at 540°C. Report by Lee et al. (2010) on NG pyrolysis showed a lower oil yield of 36 wt% 503 at 500 °C and 150 °C/min higher heating rate under nitrogen atmosphere. These variations in the 504

505 yield of bio-oil from NG biomass can be linked to the characteristic of the feedstock used, 506 heating and heat transfer rates of the reactor, and the difference in the pyrolysis condition 507 (temperature, inert gas flow).

508 3.5 Physicochemical properties of bio-oil produced at optimized condition

509 The bio-oil product collected throughout this study was two-phase liquid, the organic phase (high 510 molecular weight component) and aqueous phase (low molecular weight component). The organic fractions are usually water-insoluble derived from lignin component of the feed biomass 511 while the light fractions consist predominantly water (originates from both moisture in the feed 512 biomass and pyrolysis reaction), acids, ketones, aldehydes, small fraction of phenols and other 513 water-soluble organics (Resende et al., 2015). The physicochemical characteristics of the oil 514 515 produced at optimized condition are summarized in Table 7. Both the organic and aqueous phases present an acidic characteristics with corresponding pH values of 3.71 and 2.09. This 516 517 property is attributed to presence of organic acids and phenolics in the bio-oil. The presence of water in the pyrolysis oil is mainly from moisture in the feed biomass and product of dehydration 518 during the pyrolysis reaction (Resende et al., 2015; Mohammed et al., 2015b). In this study, the 519 520 biomass feedstock used was bone dry and, therefore, the water content of the oil could be said to 521 have originated from the pyrolysis reaction. Despite a careful separation of the aqueous phase, 7.24 wt% water remained in the organic phase. Solid content of the organic phase was 522 determined as ethanol insoluble. Bio-oil concentration of 1.25 wt% (1 g of bio-oil sample in 80 g 523 of ethanol) was used. The mixture was agitated and then filtered using 0.1µm filter. The filter 524 and retentates was thereafter oven dried at 105 °C for 30 min. Total solid content of less than 525 526 0.01 wt% (0.008-0.009) was recorded. Other characteristics of the organic phase such as density, viscosity, ash content and calorific value reasonably conform to the ASTM specifications for 527

pyrolysis oil. Consequently, the organic phase bio-oil from NGS can be regarded as ASTM-Grade D bio-oil. Carbon and oxygen contents from the ultimate analysis revealed that the organic phase bio-oil has higher carbon (51.14 wt%) and lower oxygen (41.66 wt%) compared to the value reported by Lee et al. (2010) (6.04 wt%-carbon); Sousa et al. (2016) (41.85 wt%carbon). High carbon content in the organic phase connotes presence carbon-rich organic molecules which have resulted in the higher heating value (HHV) of 26.42 MJ/kg.

534 535

Table 7: Physicochemical properties of bio-oil produced at the optimized condition

- 537
- 538 3.6 Fourier-transform infra-red (FTIR)

FTIR spectra of chemical species in the bio-oil samples are shown in Figure 7 (a & b). The 539 common broad peak around 3439 cm⁻¹ implies that the samples contain chemical compounds 540 with hydroxyl group (O-H) such as water, alcohols and phenol (Guo et al., 2015; Bordoloi et 541 al., 2015). The peak at a frequency around 2970 cm⁻¹ is due C-H stretching vibration which is 542 present only in the Figure 6b indicating the presence of saturated hydrocarbon in the organic 543 phase while the peak at a frequency around 2100 cm⁻¹ common to both phases is ascribed to the 544 $C \equiv C$ functional group (Guo et al., 2015; Bordoloi et al., 2015). Vibration observed between 545 1625 cm⁻¹ and 1707 cm⁻¹ in both oil phases is attributed to C = O which signifies the presence 546 of aldehydes, ketones or carboxylic acids. The vibration around 1462 cm⁻¹ present in the organic 547 phase is ascribed C = H indicating the presence of alkenes/aromatic hydrocarbons while the 548 peak between 1388 and 1364 cm⁻¹ in both cases is due to C-H bond (Yorgun and Yildz, 2015). 549 The sharp band around 1221, 1269, 1016 and 1092 cm^{-1} are due to C – O vibration indicating the 550 presence of alcohol and esters. The fingerprint between 900 and 620 cm⁻¹ are ascribed to 551 aromatic C – H bending vibrations (Pan et al. 2012). 552

553

Figure 7: Averaged FTIR spectra (auto-smoothed and auto-baseline corrected) of bio-oil obtained at optimized condition (a) organic phase; (b) aqueous phase

556

557

558

559 *3.7 TGA of the organic phase bio-oil*

Thermogravimetric analysis of bio-oil provides data on weight loss by evaporation as sample is 560 heated over a certain temperature range. The resulting information similar to the distillation data, 561 562 which can be used to estimate amount of bio-oil that will distill into specific fuel products. In this study, TGA analysis of commercial fossil premium motor spirit (PMS), kerosene and diesel was 563 performed as standard. The organic bio-oil was subjected to the same thermal treatment. From 564 565 Figure 8 (a), final evaporation temperature of PMS, kerosene and diesel was found to be 126, 185 and 291 °C respectively. Using the final evaporation temperature, by extrapolation, the 566 organic phase bio-oil constitutes (Figure 8b) about 70wt% volatile fraction. The mass loss above 567 300 °C can be attributed to thermal decomposition of the residue. About 68 wt% of the oil has 568 boiling range similar to that of diesel. Similarly, approximately 53 wt% of the bio-oil is made up 569 of kerosene boiling fraction while 35 wt% has boiling characteristic comparable to that of PMS. 570 571 Also, information from the TGA simulated distillation can also be used as an indicator for selecting temperature condition for GC-MS analysis of the oil. Based on the injection 572 573 temperature selected (250 °C) for the bio-oil characterization, only about 62 % of the oil can be 574 analyzed by the GC-MS.

575 3.8 GC-MS analysis of the organic phase bio-oil

Identification of detail chemical compounds in the bio-oil samples was carried out by GC-MS. 576 Library search of twenty (20) most abundant compounds using MS NIST library 2011 showed 577 that the organic phase (Table 8a) consists predominantly benzene derivatives such as phenols, 578 579 methyl-phenol, ethyl-phenol, methoxy-phenols, methoxy-benzene, benzaldehyde and benzene 580 carboxylic acid, which constitute accounted for about 60 % of the total organic phase, which is in strong agreement with the bio-oil composition reported by Lim et al (2016). Other compounds 581 identified are hydrocarbons (dimethylcyclohexene and ethylbiphenyl). These together can be 582 583 processed to fuels and valuable chemicals. The aqueous phase (Table 8b) comprised mainly organic acids, esters, ketones, aldehydes and oxygenated aromatics. The composition of organic 584 585 phase bio-oil in this study is similar to the result of GC-MS analysis reported by Strezov et al. (2008). The oil analyzed by the author consisted large amount of benzene derivatives (24.46 586 wt%). The composition of bio-oil reported by Lee et al. (2010) was made up of 27.2 % organic 587 acids, 7.9 % phenols, which is comparable to the aqueous phase composition in this study. 588 589 Generally, the bio-oil aqueous phase is considered less important and often discarded as paralysis by-products due to lack of specific applications. However, recent studies have proposed 590 591 production of hydrogen via catalytic aqueous, and steam reforming processes from this stream (Resende et al., 2015; Chan et al., 2015). 592

- 593
- 594 595

Table 8: GC-MS analysis of bio-oil (a) organic phase, (b) aqueous phase obtained at the optimized condition

596

597 3.9 GC analysis of the non-condensable gas

Samples of non-condensable gas were collected at different temperature (450, 600 and 750 °C) 598 under the pyrolysis condition of 50 $^{\rm o}C/min$ and 5 L/min N_2 as shown in Figure 9. The gas 599 600 composition (nitrogen free basis) detected include hydrogen (H₂), carbon monoxide (CO), carbon dioxide (CO₂) and methane (CH₄). At 450 °C, the main components of the gas were CO₂ 601 (16.42 vol%), H₂ (15.44 vol%) and CO (9.79 vol%). The high amount of CO₂ and H₂ and the 602 release of CO are attributed to fragmentation and subsequent transformation of unstable carbonyl 603 and carboxyl groups from depolymerization of holocellulose (hemicellulose and cellulose) and 604 605 the resulting char transformation (Collard and Blin, 2014). The trace of CH_4 (0.84 vol%) detected at this temperature is mainly due to demethylation of char from hemicellulose and 606 fragmentation of methoxy group of lignin. As the temperature increased, the amount of CO₂ 607 declined considerably due to the fact that fragmentation of glyosidic bonds in the holocellulose 608 must have been completed. This observation is in strong agreement with the TG/DTG result. A 609 total of 14.04 and 12.56 vol% CO₂ was recorded at 600 and 750 °C respectively. The amount of 610 CO_2 recorded at 450°C (16.42 vol%) in this study is similar to the value (18.5% CO_2) reported by 611 Strezov et al. (2008) during the pyrolysis of NG at 500 °C. Decline in the amount of CO₂ with 612 613 increasing pyrolysis temperature has been reported by De Conto et al. (2016). The authors recorded CO₂ value of about 42.25 vol% at 500 °C, which declined substantially to around 12 vol 614 % at 600 °C and then to about 9.vol% at 700 °C. The continuous rise in the CO, H₂ and CH₄ 615 616 components of the non-condensable gas with pyrolysis temperature can be ascribed to carboncarbon scission within the lignin and demethylation of final residue from the sample. This can 617 618 also be attributed to the pyrolysis secondary reactions usually promoted by mineral elements in 619 the biomass, particularly the alkali and alkali element (K, Ca) (Mohammed et al., 2016b). At 600

^oC, the amount of H₂, CO and CH₄ recorded was 25.32, 13.60 and 3.36 vol% respectively. This 620 621 value increased to 35.45, 23.12 and 6.55 vol% correspondingly. The ratio of H_2/CO (vol%/vol%) increased from 1.58 to 1.86 at 450 to 600 °C and there after decreased to 1.53 at 750 °C. This 622 623 indicate that the non-condensable gas from NG pyrolysis is a suitable feedstock for Fischer-Tropsch (FT) synthesis, where the syngas can be further processed into liquid fuel (Pirola et al., 624 2014; De Conto et al., 2016). However, presence of hydrogen sulfide (H₂S) in bio-syngas is a 625 common characteristic of a non-condensable gas from biomass pyrolysis due to the presence of 626 sulfur in the feedstock (Mohammed et al., 2015b). Catalyst used in the FT synthesis are highly 627 628 sensitive to H₂S even at part per million level (Yamamoto et al., 2015). Gas cleaning is therefore 629 need prior to the FT process. H_2S can be selectively removed from gas stream by absorption with the aid of selexol solvent (Mohammed et al., 2014a). 630

631 3.10 Characteristics of produced bio-char

632 Physicochemical properties of NGS bio-char produced at different temperature and 50 °C/min and 5 L/min nitrogen flow are summarized in Table 9. Volatile matter (VM) and heating value of 633 the bio-char decreased with increasing pyrolysis temperature while the ash content (AC) and 634 fixed carbon (FC) increased. This means that with increasing temperature, more VM is released 635 and highly thermal stable and non-volatile components of the source biomass are left within the 636 remaining solid. Reduction in the heating value is attributed to the increased AC, which is non-637 combustible and generally has negative impact on the solid fuel (Mohammed et al., 2014b). The 638 ultimate analysis showed increased carbon content (C) with pyrolysis temperature while oxygen 639 640 (O), and hydrogen (H) decreased (Table 9). This can be attributed to structural transformation of 641 char via scission and cracking of weak bond within the bio-char structure. Similarly, nitrogen (N) and sulfur (S) content also declined with pyrolysis temperature. This shows that N and S 642

compounds are released during pyrolysis (De <u>Conto</u> et al., 2016), which is responsible for the
sulfur and nitrogenous compounds detected in the bio-oil (Table 8).

Thermal stability of bio-char can be evaluated from the proximate, ultimate and TGA analyses. 645 De Conto et al. (2016) reported that the ratio of FC to the sum of FC and VM [FC/ (VM+FC)] of 646 647 bio-char is an indicator of thermal stability. A ratio between 0.78 and 0.85 were recorded in this 648 study, which indicates high thermal stability of organic matter in the bio-char. Spokas (2010), reported that there is connection between bio-char stability and O/C molar ratio. The author 649 stated that O/C ratio below 0.2 signifies a minimum bio-char half-life of 1000 years. In this 650 study, O/C mole ratio of 0.22, 0.15 and 0.09 was observed for the bio-char produced at 450, 600 651 652 and 750°C respectively. Similarly, Harvey et al. (2012) proposed a recalcitrance index (R_{50}) for evaluating bio-char suitability for carbon sequestration. The authors defined R₅₀ as the ratio of 653 temperature of bio-char ($T_{50 \text{ bio-char}}$) to temperature of graphite ($T_{50 \text{ graphite}}$) at which both materials 654 655 decompose 50% initial weight under oxygen atmosphere. They grouped bio-char based on the R_{50} value into class A ($R_{50} \ge 0.70$), B ($0.50 \le R_{50} < 0.70$) or C ($R_{50} < 0.50$) where class A, B and 656 C represent carbon sequestration potential similar to graphite, intermediate sequestration 657 potential and carbon sequestration potential similar to the biomass plant. Thermal decomposition 658 659 of the produced bio-char in oxygen atmosphere (50 mL/min) was carried out from ambient to 900 °C at 10 °C/min (Figure 10). Thermal stability increased with pyrolysis temperature. 660 Decomposition of the bio-char produced at pyrolysis temperature of 450, 600 and 750 °C 661 commenced at 250, 309 and 320 °C respectively. The corresponding T_{50 bio-char} was 380, 385 and 662 392 °C. Using 886 °C for T_{50 graphite} (Harvey et al., 2012), the equivalent R₅₀ value was 0.43, 0.43 663 and 0.44. Hence, the bio-chars from NGS in this study are class C and have potential carbon 664 sequestration similar to the uncharred biomass plant. Also, the TGA result indicate that the char 665

can be used in processes under moderate temperature up to 300 °C. Residue (non-combustible)
from the TGA profile (Figure 10) increased with temperature pyrolysis temperature. This
observation is in good agreement with the proximate analysis results particular the increased ash
content observed with pyrolysis temperature.

670 Scanning electron micrograph and elemental mapping of the bio-char (Figure 11) showed that 671 the materials have porous structures, which increased with pyrolysis temperature. This result was further confirmed using physisorption analysis. The specific surface area recorded for the bio-672 char produced at 450, 600 and 750 $^{\circ}$ C was 0.014, 0.126 and 0.293 m²/g and the corresponding 673 pore volume of 0.08, 010 and $0.13 \text{ cm}^3/\text{g}$ was observed. From elemental mapping of the bio-char 674 (Figure 11), potassium (K), calcium (Ca), magnesium (Mg), phosphorus (P) and chlorine (Cl) 675 676 constitute the major mineral elements in the bio-char, which originated from the source biomass. K, Ca, Mg and P are macronutrients needed for most plant growth. Bio-char can therefore be 677 678 applied as a source of macronutrient for agricultural production.

679 **4.** Conclusions

Pyrolysis of Napier grass was carried out in a vertical fixed bed tubular reactor. Optimization of 680 681 process variables such as temperature, heating rate and nitrogen flow rate was performed by means of central composite design approach. The individual model equation for bio-oil, bio-char 682 and non-condensable gas yield was developed using set of experimental data and analysis of 683 variance. Temperature, nitrogen flow rate and heating rate had significant impact on the bio-oil 684 and non-condensable gas yield while the bio-char yield was mainly affected by the pyrolysis 685 temperature.. The oil yield was maximized by optimizing the process variables and optimum bio-686 oil yield of 51.94 wt% was predicted at 600 °C, 50 °C/min and 5 L/min nitrogen flow. This result 687 was further validated through multiple experiments and the average value of the experimental 688

689 results was in good agreement with the predicted value. Therefore, it can be inferred that the 690 developed quadric model for bio-oil yield is adequate. The optimum bio-oil yield recorded in this study is higher than the pyrolysis oil yield derived from Napier grass reported in the literature. 691 692 The bio-oil obtained throughout was two-phase liquid, the organic phase (high molecular weight 693 component) and aqueous phase (low molecular weight component). Both phases collected at 694 optimized condition were characterized using standard analytical techniques. The results revealed that the organic phase consists mainly various benzene derivatives and hydrocarbons 695 which can be further processed into fuels and valuable chemicals while the aqueous phase was 696 697 predominantly water, acids, ketones, aldehydes and some phenolics and other water-soluble organics. The non-condensable gas was made-up of methane, hydrogen, carbon monoxide and 698 carbon dioxide with high hydrogen/carbon monoxide ratio suitable for Fischer-Tropsch (FT) 699 700 synthesis. Biochar collected was a porous carbonaceous material, rich in mineral elements, which may be used as adsorbent, solid fuel or source of macronutrient for agricultural 701 production. The bio-char also possessed potential for carbon sequestration. This study 702 703 demonstrated that Napier grass biomass is a potential feedstock for production of high-value bioenergy precursors. 704

705 Acknowledgement

This project was supported by the Crops for the Future (CFF) and the University of Nottinghamunder the grant BioP1-005.

708

709

710

711

- 712
 713
 714
 715
- 716

717 **References**

Anca-Couce, A., 2016. Reaction mechanisms and multi-scale modelling of lignocellulosic
 biomass pyrolysis. *Progress in Energy and Combustion Science* 53, 41-79

Bordoloi, N., Narzari, R., Chutia, R.S., Bhaskar, T., Kataki, R., 2015. Pyrolysis of Mesua ferrea
 and Pongamia glabra seed cover: Characterization of bio-oil and its sub-fractions.
 Bioresource Technology 178, 83–89

- Braga, R.M., Melo, D.M.A., Aquino, F.M., Freitas, J.C.O., Melo, M.A.F., Barros, J.M.F., 2014.
 Characterization and comparative study of pyrolysis kinetics of the rice husk and the
 elephant grass. *Journal of Thermal Analysis and Calorimetry*. 115(2), 1915–1920
- Bridgwater, A.V., 2012. Review of Fast Pyrolysis of biomass and product upgrading. *Biomass Bioenergy* 38, 68–94.
- Brosse, N., Dufour, A., Meng, X., Sun, Q., Ragauskas, A., 2012. *Miscanthus*: a fast-growing
 crop for biofuels and chemicals production. *Biofuels, Bioproducts and Biorefining* 6, 580–
 598
- Cadoux, S., Riche, A.B., Yates, N. E., Machet, J-M., 2012. Nutrient requirements of Miscanthus
 x giganteus: Conclusions from a review of published studies. *Biomass and Bioenergy* 38, 14-22
- Caputo, A., 2014. Trends in European bioenergy law: problems, perspectives and risks. *Journal of Culture, Politics and Innovation* 2014 (1-2), 1-24.
- Carraretto, C., Macor, A., Mirandola, A., Stoppato, A., Tonon, S., 2004. Biodiesel as alternative
 fuel: experimental analysis and energetic evaluations. *Energy* 29, 2195–2211.
- Chaiwat, W., Gunawan, R., Gholizadeh, M., Li, X., Lievens, C., Hu, X., Wang, Y., Mourant, D.,
 Rossiter, A., Bromly, J., Li, C-Z., 2013. Upgrading of bio-oil into advanced biofuels and
 chemicals. Part II. Importance of holdup of heavy species during the hydrotreatment of biooil in a continuous packed-bed catalytic reactor. *Fuel* 112, 302–310
- Chen, A., Pan, C., Cao, D., Lou, H., 2015. Aqueous-phase reforming of the low-boiling fraction
 of bio-oil for hydrogen production: The size effect of Pt/Al₂O₃. *International Journal of Hydrogen Energy* 40(43), 14798–14805
- Cherubini, F., 2010. The biorefinery concept: using biomass instead of oil for producing energy
 and chemicals. *Energy Conversion and Management* 51, 1412–21.

- Collard, F.X., Blin, J., 2014. A review on pyrolysis of biomass constituents: Mechanisms and
 composition of the products obtained from the conversion of cellulose, hemicelluloses and
 lignin. *Renewable and Sustainable Energy Reviews* 38, 594–608
- De Conto, D., Silvestre, W.P., Baldasso, C., Godinho, M., 2016. Performance of Rotary Kiln
 Reactor for the Elephant Grass Pyrolysis. *Bioresource Technology* 218, 153-160
- Eom, I.Y., Kim, J.Y., Lee, S.M., Cho, T.S., Yeo, H., Choi, J.W., 2012. Comparison of pyrolytic
 products produced from inorganic-rich and demineralized rice straw (Oryzasativa L.) by
 fluidized bed pyrolyzer for future biorefinery approach. *Bioresource. Technology.* 128,
 664–672
- Flores, R. A., Urquiaga, S., Alves, B.J.R., Collier, L.S., Boddey, R. M., 2012. Yield and quality
 of elephant grass biomass produced in the cerrados region for bioenergy. *Engenharia Agrícola 32*(5), 831-839.
- Gómez, N., Rosas, J.G., Cara, J., Martínez, O., Alburquerque, J.A., Sánchez, M.E., 2016. Slow
 pyrolysis of relevant biomasses in the Mediterranean basin. Part 1. Effect of temperature on
 process performance on a pilot scale. *Journal of Cleaner Production* 120, 181-190.
- Guo, Y., Song, W., Lu, J., Ma, Q., Xu, D., Wang, S., 2015. Hydrothermal liquefaction of
 Cyanophyta: Evaluation of potential bio-crude oil production and component analysis. *Algal Research* 11, 242–247

Harvey, O.R., Kuo, L.J., Zimmerman, A.R., Louchouarn, P., Amonette, J.E., Herbert, B.E.,
2012. An index-based approach to assessing recalcitrance and soil carbon sequestration
potential of engineered black carbons (Biochars). *Environmental Science and Technology*46 (3), 1415–1421

- Heidenreich, S., Foscolo, P. U., 2015. New concepts in biomass gasification. *Progress in Energy and Combustion Science* 46, 72-95
- Imam,T., Capareda, S., 2012. Characterization of bio-oil, syn-gas and bio-char from switchgrass
 pyrolysis at various temperatures. *Journal of Analytical and Applied Pyrolysis* 93, 170-177
- Kan, T., Strezov, V., Evans, T.J., 2016. Lignocellulosic biomass pyrolysis: A review of product
 properties and effects of pyrolysis parameters. *Renewable and Sustainable Energy Reviews* 57, 1126–1140
- Lee, M.-K., Tsai, W.-T., Tsaic, Y.-L., Lin, S.-H., 2010. Pyrolysis of Napier grass in an
 induction-heating reactor. *Journal of Analytical and Applied Pyrolysis* 88, 110–116.
- Lim, C.H., Mohammed, I.Y., Abakr, Y.A., Kazi, F.K., Yusup, S. and Lam, H.L., 2016. Novel
 input-output prediction approach for biomass pyrolysis. *Journal of Cleaner Production*, *30*,
 1-11.

Lim, C.H., Mohammed, I.Y., Abakr, Y.A., Kazi, F.K., Yusup, S. and Lam, H.L., 2015. Element
 characteristic tolerance for semi-batch fixed bed biomass pyrolysis. *Chemical Engineering Transactions 45*, 1285-1290

784

.

- Mohammed, I. Y., Samah, M., Mohamed, A., & Sabina, G. 2014a. Comparison of SelexolTM
 and Rectisol® Technologies in an Integrated Gasification Combined Cycle (IGCC) Plant
 for Clean Energy Production. *International Journal of Engineering Research* 3(12), 742744
- Mohammed, I.Y., Abakr, Y.A., Kazi, F.K. and Yusuf, S., 2016a.. Effects of pretreatments of
 Napier grass with deionized water, sulfuric acid and sodium hydroxide on pyrolysis oil
 characteristics. *Waste and Biomass Valorization*, 1-19.
- Mohammed, I.Y., Abakr, Y.A., Kazi, F.K., Yusup, S., Alshareef, I., Chin, S.A., 2015b. Pyrolysis
 of Napier Grass in a Fixed Bed Reactor: Effect of Operating Conditions on Product Yields
 and Characteristics. *BioResources*10 (4), 6457–6478.
- Mohammed, I.Y., Abakr, Y.A., Kazi, F.K., Yusup, S., Alshareef, I., Chin, S.A., 2015a.
 Comprehensive Characterization of Napier Grass as a Feedstock for Thermochemical Conversion. *Energies* 8(5), 3403-3417.
- Mohammed, I.Y., Kazi, F.K., Yusuf, S.B., Alshareef, I. and Chi, S.A., 2014b. Higher Heating
 Values (HHV) Prediction Model from Biomass Proximate Analysis Data. In *International conference & exhibition on clean energy, October 20–22, Quebec, Canada.*
- Mohammed, I. Y., Abakr, Y. A., Musa, M., Yusup, S., Singh, A., & Kazi, F. K. (2016c).
 Valorization of Bambara groundnut shell via intermediate pyrolysis: Products distribution
 and characterization. *Journal of Cleaner Production*, 717-728.
- Mohammed, I.Y., Kazi, F.K., Yusup, S., Alaba, P.A., Sani, Y.M., Abakr, Y.A., 2016d. Catalytic
 Intermediate Pyrolysis of Napier Grass in a Fixed Bed Reactor with ZSM-5, HZSM-5 and
 Zinc-Exchanged Zeolite-A as the Catalyst" *Energies* 9 (4), 246.
- Mohammed, I.Y., Lim, C.H., Kazi, F.K., Yusup, S., Lam, H.L. and Abakr, Y.A., 2016b. Copyrolysis of Rice Husk with Underutilized Biomass Species: A Sustainable Route for
 Production of Precursors for Fuels and Valuable Chemicals. *Waste and Biomass Valorization*, 1-11.
- Ning, S-K., Hung, M-C., Chang, Y-H., Wan, H-P., Lee, H-T., Shih, R-F., 2013. Benefit
 assessment of cost, energy, and environment for biomass pyrolysis oil. *Journal of Cleaner Production* 59, 141-149

- Pan, C., Chen, A., Liu, Z., Chen, P., Lou, H., Zheng, X., 2012. Aqueous-phase reforming of the
 low-boiling fraction of rice husk pyrolyzed bio-oil in the presence of platinum catalyst for
 hydrogen production. *Bioresource Technology* 125, 335–339
- Patwardhan, P.R., Brown, R.C., Shanks, B.H., 2011. Product distribution from the fast pyrolysis
 of hemicellulose. *ChemSusChem* 4(5), 636–643.
- Pirola, C., Fronzo, A., Galli, F., Bianchi, C.L., Comazzi, A., Manenti, F., 2014. Biosyngas
 Conversion by Fischer Tropsch Synthesis: Experimental Results and Multi-scale
 Simulation of a PBR with High Fe Loaded Supported Catalysts. *Chemical. Engineering. Transaction* 37, 595-600.
- Resende, K.A., Ávila-Neto, C.N., Rabelo-Neto, R.C., Noronha, F.B., Hori, C.E., 2015.
 Thermodynamic analysis and reaction routes of steam reforming of bio-oil aqueous
 fraction. *Renewable. Energy* 80,166–176.
- Richter, G.M., Riche, A.B., Dailey, A.G., Gezan, S.A., Powlson, D.S., 2008. Is UK biofuel
 supply from Miscanthus water-limited? *Soil Use Manage* 24, 235–245.
- Samson, R., Mani, S., Boddey, R., 2005. The potential of C4 perennial grasses for developing a
 global bioheat industry. *Critical Reviews in Plant Sciences* 24(5-6), 461–495.
- Soetardji, J.P., Widjaja, C., Djojorahardjo, Y., Soetaredjo, F.E., Ismadjia, S., 2014. Bio-oil from
 jackfruit peel waste. *Procedia Chemistry* 9,158 164.
- Sousa, J.F., Bezerra, M.B.D., Almeida, M.B.B., Moure, G.T., Mesa-Perez, J.M., Caramao, E.B.,
 2016. Characteristics of Bio-oil from the Fast Pyrolysis of Elephant Grass (*Pennisetum purpureum Schumach*) in a Fluidized Bed Reactor. *American Chemical Science Journal*14(2), 1-10
- Spokas, K.A., 2010. Review of the stability of biochar in soils: predictability of O:C molar
 ratios. *Carbon Management* 1(2), 289–303
- Strezov, V., Evans, T.J., Hayman, C., 2008. Thermal conversion of elephant grass (Pennisetum
 Purpureum Schum) to bio-gas, bio-oil and charcoal. *Bioresource Technology* 99, 8394–
 8399.
- Tripathi, M., Sahu, J.N., Ganesan, P., 2016. Effect of process parameters on production of
 biochar from biomass waste through pyrolysis: A review. *Renewable and Sustainable Energy Reviews* 55, 467–481.
- Tsai, W.T. and Tsai, Y.L., 2016. Thermochemical characterization of Napier grass as an energy source and its environmental and economic benefit analysis. *Energy Sources, Part B: Economics, Planning, and Policy* 11(2), 130-136.

- Van de Velden, M., Baeyens, J., Brems, A., Janssens, B., Dewil, R., 2010. Fundamentals,
 kinetics and endothermicity of the biomass pyrolysis reaction. *Renewable Energy* 35, 232–
 242.
- Yamamoto, T., Tayakout-Fayolle, M., Geantet, C., 2015. Gas-phase removal of hydrogen
 sulfide using iron oxyhydroxide at low temperature: Measurement of breakthrough
 curve and modeling of sulfidation mechanism. *Chemical Engineering Journal* 262, 702–
 709.
- Yorgun, S., Yıldız, D., 2015. Slow pyrolysis of paulownia wood: Effects of pyrolysis parameters
 on product yields and bio-oil characterization. *Journal of Analytical and Applied Pyrolysis*114, 68–78
- Zhou, Y., Jiang, J., Peng, Z., Wang, Q., Xiong, D., 2012. Ecosystem Management in the Natural
 Rubber Industry. *Journal of Resources and Ecology* 3(3), 230-235
- 860 Standards used
- BS EN 14774-1. Solid Biofuels. Determination of Moisture Content. Oven Dry Method. Total
 Moisture. *British Standards Institution* 2009, London, UK
- BS EN 15148. Solid biofuels. Determination of the content of volatile matter. *British Standards Institution* 2009, London, UK
- BS EN 14775. Solid biofuels. Determination of ash content. *British Standards Institution* 2009,
 London, UK
- BS EN 15290. Solid biofuels. Determination of major elements-Al, Ca, Fe, Mg, P, K, Si, Na and
 Ti. *British Standards Institution* 2011, London, UK
- BS EN 14918. Solid biofuels. Determination of calorific value. *British Standards Institution*2009, London, UK
- Sluiter A, Hames B, Ruiz R, Scarlata C, Sluiter J, Templeton D, Crocker, D. Determination of
 Structural Carbohydrates and Lignin in Biomass Laboratory Analytical Procedure.
 National Renewable Laboratory 2012; NREL/TP-510-42618
- 20. ASTM D240. Standard test method for heat of combustion of liquid hydrocarbon fuels by
 bomb calorimeter. *ASTM International* 2009, West Conshohocken, PA.
- ASTM D7544 12 Standard Specification for Pyrolysis Liquid Biofuel. ASTM International
 2012, West Conshohocken, PA.
- 878

879